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ECOLOGICAL PATTERN OF MICROALGAL COMMUNITIES AND ASSOCIATED RISKS IN COASTAL ECOSYSTEMS

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Ecological Pattern of Microalgal Communities and Associated Risks in Coastal Ecosystems

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A thesis submitted in partial fulfilment of the requirements for the degree of Doctor of Philosophy

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Certificate of Originality

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Abstract

The occurrence of Harmful algal blooms (HABs) is a prominent environmental issue that pose significant threats to marine ecosystems, seafood safety, and human health. Over recent decades, the increasing frequency, intensity, and global distribution of HABs have underscored the urgency of understanding their ecological dynamics and associated risks. Beyond natural seawater, human-induced substrates such as plastics also provide a novel habitat, "the plastisphere" (a microecosystem with plastics as the matrix) for microalgal communities in the coastal ecosystem. However, the ecological dynamics and associated risks of microalgal communities in both seawater and plastisphere, as well as the associations between these two habitats, remain largely unexplored. Moreover, the absence of a specific and comprehensive sequence database for harmful and toxic microalgae has hindered the accurate identification and monitoring of these species. To address these gaps, this study developed the harmful and toxic microalgae database (HTMaDB) to reveal the ecological dynamics and microalgae-associated risks in the coastal ecosystem.

We constructed the HTMaDB, a comprehensive resource that consolidates 1,346 18S full-length reference sequences from public repositories. A total of 79 harmful and toxic species were identified in the typical coastal city Hong Kong surface seawaters with the application of curated database. The results revealed that dinoflagellates and diatoms constitute most of the toxic and harmful taxa, respectively. Seasonal variations in algal

community structures were observed. Specifically, temperature was found to be the primary factor shaping seasonal patterns of toxic algal communities, with increased toxic species abundance during the dry season. The dominance of causative taxa in this period was consistent with the higher prevalence of algal toxins, supporting the reliability of the database. These findings highlight the potential for escalating risks associated with toxic microalgae under a warming climate, as differential responses of specific species to temperature fluctuations contribute to the complexity of HAB occurrences. Species such as *Alexandrium* spp., *Dinophysis* spp., *Prorocentrum* spp., and *Karenia* spp. were identified as the primary toxin producers, underscoring the critical need for targeted management strategies to mitigate their ecological and health impacts.

In addition to exploring the ecological patterns of algal communities in the surface seawater, this study further investigated the ecological dynamics and associated risks of microalgal communities within the plastisphere. Plastic pollution has become a global environmental challenge, with millions of tonnes of plastic waste entering the oceans annually. The plastisphere provides a durable and hydrophobic surface that facilitates the colonization and transport of microorganisms, including harmful and toxic algae. By analyzing 92 paired samples of plastic debris and ambient water collected from coastal ecosystems, this research revealed that the plastisphere selectively harbors distinct algal communities compared to the surrounding waters. The relative abundance of Chlorophyta increased significantly, while that of Dinoflagellata

decreased within the plastisphere. Furthermore, the plastisphere exhibited significantly higher alpha diversity, indicating its potential to shelter unique or alien microalgal taxa transported from upstream systems. These results suggest that the plastisphere serves as a microhabitat that enables the selective assembly, enrichment, and transport of harmful and toxic algae across ecosystems.

The study also found that harmful and toxic diatoms, such as *Pseudo-nitzschia cuspidata*, *Chaetoceros socialis*, and *Skeletonema marinoi*, thrived within the plastisphere. Some harmful algal taxa were exclusively detected on plastic debris but absent from the surrounding water, highlighting the potential of plastics to act as vectors for the transport of harmful algae. Environmental factors such as temperature and salinity were positively associated with the abundance of risk-associated microalgae, whereas pH exhibited a negative correlation. These findings demonstrate that plastic pollution not only alters the composition and diversity of microalgal communities but also amplifies the ecological risks associated with HABs by facilitating the spread and enrichment of harmful algae and their toxins. The potential for plastics to carry harmful algal species across marine ecosystems poses significant challenges for managing HAB risks in the context of increasing plastic pollution.

This integrated study bridges critical knowledge gaps by systematically investigating the pelagic microalgal communities in the coastal ecosystem. By developing HTMaDB, this research provides a valuable resource for improving the accuracy and efficiency of

HAB monitoring and identification. The findings highlight the role of environmental factors, particularly temperature, in driving the seasonal dynamics of toxic algal communities and emphasize the impact of climate change on HAB risks. Additionally, the investigation into the plastisphere offers novel insights into how plastic debris assembled a distinct microalgal community compared to the seawater community and increased the risks associated with harmful and toxic algae. These findings contribute to a comprehensive ecological view of pelagic microalgal communities in the coastal ecosystem.

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As I stand at this juncture where this protracted academic journey approaches its terminus, retrospection unveils the perplexities encountered in literature reviews, the uncertainties faced with datasets, now interwoven with this moment of clarity and gratitude, coalesce into heartfelt appreciation to the numerous individuals who illuminated my path.

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Abbreviations

| Abbreviations | Full names |
|---------------|--|
| ASP | Amnesic shellfish poisoning |
| AST | Amnesic shellfish toxins |
| ASVs | Amplicon sequence variants |
| AZA-1 | Azaspiracid-1 |
| AZA-2 | Azaspiracid-1 |
| AZP | Azaspiracid poisoning |
| CE | Collisions energies |
| CEP | Collision cell entrance potentials |
| СР | Ciguatera poisoning |
| DA | Domoic acid |
| DDBJ | DNA Data Bank of Japan |
| DO | Dissolved oxygen |
| DOC | Dissolved organic carbon |
| DP | Declustering potentials |
| DSP | Diarrhetic shellfish poisoning |
| DST | Diarrhetic shellfish toxins |
| DTX-1 | Dinophysistoxin-1 |
| DTX-2 | Dinophysistoxin-2 |
| EMBL | European molecular biology laboratory |
| EP | Entrance potentials |
| EPS | Extracellular polymeric substances |
| GYM | Gymnodium |
| HABs | Harmful algal blooms |
| HAEDAT | Harmful algal event database |
| HTMaDB | Harmful and toxic microalgae database |
| HTS | High-throughput sequencing |
| INSDC | International nucleotide sequence database collaboration |
| LAT(s) | Lipophilic algal toxin(s) |
| LOD | Limits of detection |
| LOQ | Limit of quantification |

| LSU | Large subunit |
|-----------|---|
| MST | Modified stochasticity ratio |
| NCBI | National center for biotechnology information |
| NSP | Neurotoxic shellfish poisoning |
| NST | Normalized stochasticity ratio |
| OA | Okadaic acid |
| PCoA | Principal coordinate analysis |
| PERMANOVA | Permutational Multivariate Analysis of Variance |
| POPs | Persistent organic pollutants |
| PP | Polypropylene |
| PR2 | Protist ribosomal reference database |
| PSP | Paralytic shellfish poisoning |
| PSTs | Paralytic shellfish toxins |
| PTX-2 | Pectenotoxin-2 |
| PTXs | Pectenotoxins |
| RDA | Redundancy analysis |
| rRNA | Ribosomal RNA |
| SD | Standard deviation |
| SPM | Suspended particulate matters |
| SSU | Small subunit |
| VPA | Variation partitioning analysis |
| YTX | Yessotoxin |

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Chapter 1 Introduction

1.1 Background

The global range, regional intensity, and frequency of harmful algal blooms (HABs) have increased over the last few decades in coastal areas (Dai et al. 2023a). Some of HABs-forming species can release toxins leading to fish death, shellfish poisoning, and even human illnesses after consuming contaminated seafood (Berdalet et al. 2016). To date, over 300 phycotoxins and Furthermore, a number of seafood poisoning cases caused by algal toxins have been reported worldwide (Holmes & Teo 2002; Hossen et al. 2011; Jiang et al. 2017; Kim et al. 2010; Mackenzie 2019; Mafra et al. 2019; Visciano et al. 2016). These seafood toxins accounted for nearly half of total HAB events, and over the past three decades, an increase in the global frequency of causative organisms, such as Dinophysis spp., Alexandrium spp., and Pseudo-nitzschia spp. was observed according to the global HAB status report 2021. Some nontoxic microalgae form high biomass blooms and cause mortality of aquatic organisms by clogging gills or triggering hypoxia/anoxia conditions in the ambient water (Anderson et al. 2021a). This global ecological phenomenon has resulted in remarkable socioeconomic losses due to massive fish kills and fishery closure (Boyce et al. 2017).

Microalgae are not only found in natural water bodies, growing attention has been paid to the colonization in artificial or human-disturbed niches, such as the "plastisphere" (microbial habitat with plastic as the matrix) (Amaral-Zettler et al. 2020; Zettler et al.

2013). These non-traditional niches also participate in the dynamics of pelagic communities and have an important impact on their structure and function. The buoyancy, durability, and hydrophobic properties of plastics enable them to act as vectors for the transport and enrichment of harmful species like disease-causing microorganisms and antibiotic resistant bacteria (Bergmann et al. 2022; Bhagwat et al. 2021; Dey et al. 2022). These studies on plastisphere microbiology and associated risks mainly focus on bacteria, with comparatively less attention given to eukaryotes. Other life forms, such as microalgae, a key component of aquatic ecosystems, are poorly understood in the plastisphere. Uncovering the full picture of this novel, unique habitat generated in the Anthropocene and the associated potential ecological threats will help to achieve a more comprehensive understanding of plastic pollution risks and the influence of human activities on the Earth's microbiome (Li et al. 2024b; Rillig et al. 2024a). A previous investigation shows that one square centimeter of the plastic surface can harbor more than 80,000 diatom cells, pinpointing that marine plastispheres act as a suitable habitat for microalgae containing substantial microalgal biomass (Casabianca et al. 2019). However, the compositional dynamics and associated ecological risks of microalgal communities within distinct habitats, namely, the seawater and the plastisphere, remain poorly understood. Furthermore, the potential interactions, convergences, or divergences between these two ecological niches have not been thoroughly investigated, leaving a critical gap in our understanding of how microalgal populations function and evolve in anthropogenically influenced marine environments.

Metabarcoding monitoring efforts are highly beneficial for early HAB species detection and risk assessment due to robust sensitivity and specificity when compared to traditional time-consuming, labor-intensive morphology-based approach (Feist & Lance 2021). Many microalgal species with small sizes are difficult to distinguish via their morphology (Huang et al. 2021; Pearson et al. 2021). The accuracy of DNA metabarcoding will fluctuate depending on the reference database, however, there is no specific harmful and toxic microalgae sequence database till now.

Collectively, this PhD thesis provides significant contributions to the field of marine ecology by establishing a comprehensive database of harmful and toxic microalgae, exploring the ecological dynamics and risks of pelagic microalgae communities including surface seawater and plastisphere, and the interactions between these two habitats in the coastal ecosystem. Moreover, the exploration of the plastisphere reveals new insights into the role of plastic debris in altering microalgal community composition, thereby increasing the risks associated with harmful and toxic algae. The findings advance our understanding of how environmental and anthropogenic factors influence algal blooms and their ecological consequences. This work not only enhances our ability to monitor and manage HABs effectively but also underscores the emerging environmental challenge posed by plastic debris in marine ecosystems. Continued research and adaptive management strategies are essential to mitigate these ecological threats and protect marine biodiversity.

1.2 Research objectives

This thesis aims at conducting an in-depth examination of the ecological pattern of microalgal communities and the associated risks in coastal ecosystems. The specific objectives of this study are as follows:

- a) To develop and apply a comprehensive database (HTMaDB) containing taxonomic information on harmful and toxic microalgae, thereby enabling accurate identification and effective monitoring of HAB species;
- b) To analyze the spatiotemporal distribution patterns of harmful and toxic algae associated with key algal toxins in coastal surface seawater, in order to validate the reliability and applicability of the newly constructed database;
- c) To investigate the characteristics of microalgal communities within the plastisphere compared to the seawater microalgal community, with a particular focus on the enrichment of HAB species and their implications for marine ecosystem; and
- d) To determine what are the key environmental factors governing HAB community dynamics and the potential implications of climate change on the HAB community.

1.3 Organization

This thesis is structured into seven chapters, including "Introduction," "Literature Review," "Methodology," two results and discussion chapters, and a concluding chapter with "Conclusions and Recommendations." Chapter 1 introduces the background of the study, focusing on HABs and their ecological and economic impacts, as well as the

emerging challenges of plastic pollution and the plastisphere. Furthermore, the thesis structure is outlined as guidelines for the reader, and the research objectives are set out. Chapter 2 summarizes existing knowledge on phytoplankton communities in coastal ecosystems regarding their ecological impacts and environmental drivers. Several monitoring and detection techniques, traditional and DNA-based, molecular and HABspecific database development, are discussed. The chapter then looks at the plastisphere as a new microhabitat, how it forms, the ecological risks associated with living in it, and the emerging issue of the proliferation of HAB species within this microhabitat. Chapter 3 details the methodologies employed in the thesis. It includes the construction of the database (HTMaDB), covering the collection, taxonomic, and diversity of HAB species. It also describes marine sampling campaigns, including sampling strategies, environmental parameter measurements, DNA extraction, and sequencing techniques. Statistical analyses to quantify algal community patterns and chemical analyses for LATs are also described. In chapter 4 HTMaDB is applied to analyze distribution dynamics of microalgal communities in coastal seawater to validate the reliability of the dataset. Specifically, it maps harmful and toxic microalgae and evaluates environmental factors, such as temperature, shape toxic algal communities. It further explores the linkage among environmental parameters, non-toxic species interactions, and toxic algal community variations, identifying distribution pattern of potential causative taxa linked with lipophilic algal toxins (LATs). Chapter 5 examines the microalgal communities and associated risks in the plastisphere in coastal ecosystems, focusing on ecological shifts induced by plastic debris. It reveals how plastics

selectively enrich HAB species, their interactions with the environmental factors, as well as the potential risks posing to marine biodiversity and ecosystem dynamics. Finally, Chapter 6 summarizes the major findings and limitations of the current research, as well as provides a future perspective on advancing the study of microalgal communities and their associated risks in coastal ecosystems.

Chapter 2 Literature Review

Given the growing ecological threats posed by HABs and plastic pollution in coastal ecosystems, this chapter provides a comprehensive review of the interactions between HABs, plastic debris, and microalgal communities under a climate change world. The review focuses on understanding the ecological impacts and potential risks associated with these environmental challenges. From an ecological and environmental health perspective, this review emphasizes the role of toxigenic microalgae and how plastic pollution may influence the spread and dynamics of harmful algal species. It further explores the mechanisms driving the formation of HABs and plastisphere communities, their spatiotemporal variations, sources, and risk assessments, highlighting critical research gaps and guiding future studies toward effective monitoring and management strategies.

2.1 The phytoplankton in the coastal ecosystems

2.1.1 HABs and ecological impact

HABs, consisting of over-proliferation of algae in both marine and freshwater systems, represent a critical and escalating global environmental challenge. The frequency, regional intensity, and geographic distribution of HABs have increased over the last few decades in coastal areas, posing threats to ecosystems and human health (Anderson et al. 2021b; Dai et al. 2023b; McKenzie et al. 2021; Yu et al. 2023b). Of the nearly 5,000 documented marine phytoplankton species inhabiting global waters, researchers have identified more than 100 distinct variants that endanger human populations by

generating harmful toxins. These bioactive compounds compromise both food safety through contaminated marine harvests and public health via direct exposure risks (Hallegraeff et al. 2021a,b). The poisonous compounds directly impact aquatic ecosystems by triggering mass mortality events in both natural and farmed fish populations. Their presence disrupts coastal recreational activities and inflicts substantial financial damage on marine industries. Even non-toxic algal species contribute to harmful blooms when excessive growth leads to visible water discoloration, oxygen depletion, or sticky mucus production - all capable of severely disrupting ecological balance and coastal livelihoods. The Harmful Algal Event Database (HAEDAT) serves as the sole global repository documenting toxic bloom incidents across marine and freshwater systems, with significant disparities in regional reporting accuracy. Its records categorize algae-related phenomena that trigger regulatory interventions, incur financial losses, or alter ecosystem dynamics. These incidents encompass four operational definitions: water surface anomalies like colored patches or foam linked to socioeconomic disruptions from algal presence (toxic or benign); seafood contamination exceeding human safety thresholds from accumulated biotoxins; precautionary closures of fisheries or recreational areas due to algal overgrowth risks; and any algae-driven impacts on organisms, regardless of direct human health implications. Seafood biotoxin incidents (48%.) accounted for nearly half of total HAEDAT events across the globe. Among all events associated with seafood poisoning syndromes, Paralytic Shellfish Toxins (PST) contributed 35%, Diarrhetic Shellfish Toxins (DST) 30%, Ciguatera Poisoning (CP) and toxins from marine and

brackish water cyanobacteria each accounted for 9%, Amnesic Shellfish Toxins (AST) represented 7%, and other toxins made up the remaining 10%. (Fig.1) (Hallegraeff et al. 2021b).

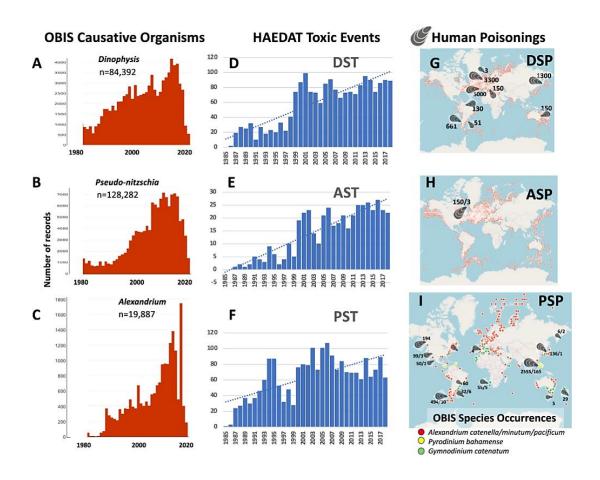


Figure 2-1 An increases global trend of HAB associated organisms, poisoning events and distributions of the toxin syndromes between 1985 and 2018. (A-C document the total number of global observations of causative microalgal organisms of *Dinophysis spp., Pseudo-nitzschia spp. and Alexandrium spp.*; D-F provid the number of records of HAEDAT database regarding toxic events of DST, AST and PST.; G–I show the distribution maps as of 2018 of *Dinophysis spp.* (DSP), *Pseudo-nitzschia spp.* (ASP) and *Alexandrium, Pyrodinium, Gymnodinium catenatum* (PSP) with red dots across the

globe. Toxic incidents that have led to human poisoning are represented on the map with shellfish icons, whose sizes correspond to the severity of the events. The first number shows the number of poisonings, the second number indicates fatalities. Specifically, for ASP, the notation 150/3 denotes 150 clinical cases with three fatalities) (Hallegraeff et al. 2021b).

The ecological impacts of HABs are profound and multifaceted. Toxins produced by harmful algae can bioaccumulate in marine organisms, such as fish, shellfish, and other invertebrates, leading to mass mortalities in marine life (Karlson et al. 2021). These toxins can also be transferred through the food web, impacting top predators, including marine mammals, seabirds, and even humans (Turner et al. 2021). Ingesting contaminated seafood can lead to a variety of shellfish poisoning illnesses. These conditions can lead to severe health complications and even death in some cases. Furthermore, some HABs are associated with aerosolized toxins that can be inhaled by people, causing respiratory issues, particularly in coastal communities (Patterson et al. 2016). For example, Karenia brevis, the species responsible for red tides in the Gulf of Mexico, can release toxic aerosols that cause respiratory irritation and exacerbate conditions like asthma in people living near affected areas (Jang et al. 2025). The socioeconomic consequences of HABs are staggering. In regions heavily dependent on fisheries and aquaculture, HABs can result in massive losses. The red tide event along U.S. West Coast in 2015 led to an estimated \$48.3 million losses (Anderson et al. 2021c). Similarly, in Asia, coastal communities rely on shellfish harvesting and aquaculture, which can be severely impacted by HABs. The presence of toxins in seafood can lead to closures of fisheries, tourist beach restrictions, and the contamination of commercial seafood, all of which result in significant economic losses.

2.1.2 Environmental drivers of HABs

Microalgal communities in marine and coastal ecosystems are highly sensitive to environmental changes, with temperature, salinity, and nutrient availability being the most influential factors shaping their structure, diversity, and functionality. These factors not only regulate the growth and composition of microalgal populations but also play a pivotal role in the formation and intensity of HABs. Understanding how these environmental variables interact with microalgal communities is critical for predicting ecosystem responses to natural and anthropogenic disturbances, including climate change and nutrient pollution.

Temperature is one of the most significant environmental drivers of microalgal dynamics. In general, temperature significantly influences the physiology and metabolism of phytoplankton (Falkowski et al. 2004). Various species of microalgae thrive at specific temperatures. For instance, widely cultivated algal species like *Chlorella, Scenedesmus, Chlorella vulgaris, Navicula, Chlamydomonas, Haematococcus*, and *Monoraphidium* typically flourish in temperatures between 15 and

35°C (Singh & Singh 2015). Microalgal communities react differently to specific temperatures based on their strain's optimal temperature, exhibiting negative, neutral, or positive responses (Kholssi et al. 2023). Rising water temperatures are expected to enhance both the abundance and diversity of *Gambierdiscus* and *Fukuyoa* species in the Gulf of Mexico and along the southeastern U.S. Atlantic coast. In contrast, species adapted to lower temperatures in the Caribbean Sea would likely decline in prevalence. However, their distribution is projected to extend further north into the Gulf of Mexico and potentially reach the western Atlantic (Kibler et al. 2015). Elevated temperatures can also disrupt predator-prey relationships, reduce grazing pressure on microalgae, and shift competitive advantages among species, thereby altering community dynamics (Anderson et al., 2002). Additionally, warmer temperatures often lead to stronger water column stratification, reducing vertical mixing and nutrient exchange, which can further stimulate harmful algal species that are adapted to low-nutrient conditions.

Temperature and salinity are two environmental factors which are closely related to climate change. With the increase of seawater temperature, the salinity changes accordingly (Frame & Stone 2013). Salinity significantly influences the distribution and composition of microalgal communities, especially in estuarine and coastal regions where freshwater and seawater interact. Changes in salinity can create osmotic stress, limiting the growth of some species while promoting the dominance of halotolerant or halophilic species. For example, *Prorocentrum minimum* and *Karlodinium veneficum*

are known to tolerate wide salinity ranges, enabling them to bloom in estuarine environments (Ajani et al. 2018; Zhang et al. 2008). Fluctuations in salinity caused by heavy rainfall, river discharge, or drought conditions can disrupt existing microalgal assemblages and facilitate the introduction and establishment of invasive or harmful species (Lazrak et al. 2024; Röthig et al. 2023).

Nutrient enrichment is another critical factor regulating microalgal community structure. Key nutrients, particularly nitrogen (N) and phosphorus (P), serve as essential resources for microalgal growth. Coastal eutrophication, often driven by agricultural runoff, wastewater discharge, and urbanization, can lead to excessive nutrient loading in aquatic ecosystems (Lie et al. 2011). The form and ratio of nutrients (e.g., N:P ratio) also influence species composition, with certain species favoring either nitrate- or ammonium-rich conditions (Lagus et al. 2004; Nhu Y et al. 2019). Moreover, nutrient limitation or imbalance can trigger toxin production in certain species such as *Prorocentrum hoffmannianum*, further elevating ecological and health risks (Accoroni et al. 2018).

The combined effects of temperature, salinity, and nutrients often interact in complex ways, amplifying their influence on microalgal community dynamics. Warming temperatures can intensify nutrient-driven eutrophication effects by accelerating nutrient cycling and stimulating harmful algal species adapted to high nutrient and

warm conditions. Similarly, altered precipitation patterns due to climate change can modify salinity regimes and nutrient inputs, creating conditions conducive to HAB outbreaks. Therefore, it is essential to consider the synergistic effects of these environmental factors when assessing microalgal community responses.

2.1.3 Assembly process in shaping the algal community

Understanding the relative contribution of different assembly processes in shaping microalgal communities over spatiotemporal scales is crucial for understanding how these communities respond to local, regional, and global changes (Skouroliakou et al. 2022). Deterministic and stochastic processes explain the assembly of microbial communities (Nemergut et al. 2013; Zhou & Ning 2017). Deterministic processes generally refer to factors that affect community structure through environmental selection and species interactions (Chesson 2000; Fargione et al. 2003). Environmental filtering plays a key role in deterministic processes by emphasizing how environmental factors like temperature, salinity, light, and nutrient availability influence the distribution of species. Water temperature directly affects the growth and reproduction of algal species (Matula et al. 2022). While some algae species are thermally sensitive, thermotolerant algae (cyanobacteria, green algae, certain species of dinoflagellates, etc.) can prevail in communities, especially in warmer waters (Rossi et al. 2023). Being more tolerant to high temperatures than other algae, they tend to increase reproduction at higher temperatures, which modulates the composition of the community. The

distribution of different species also depends on salinity variations; some species of algae perform better under low salinity while others are adapted to high salinity (Mo et al. 2021). Nutrient concentration is another critical environmental factor directly affecting algal growth rates and community composition, especially the concentrations of nitrogen and phosphorus, which often determine the prosperity of certain algal populations. In addition to environmental factors, species interactions are also a significant component of deterministic processes (Zhou & Ning 2017). The community structure is often affected by competitive, predatory, and symbiotic interactions among species (Solomon et al. 2022). Certain algal species may dominate specific ecological niches through competitive advantages, inhibiting the growth of other species. Some algae enhance their survivability through symbiotic relationships with microbial communities, obtaining nutrients. Additionally, predatory pressures can alter algal community structures, especially when planktonic predators consume algae (Behrenfeld et al. 2021; Thompson et al. 2024).

Stochastic ecological processes including dispersal, speciation, and ecological drift, which typically drive species composition and diversity (Chase & Myers 2011; Stegen et al. 2012). Dispersal refers to the movement of species across different areas, while speciation describes the generation of new genetic variation. Ecological drift, on the other hand, is the random fluctuations in the abundance of species over time, driven by stochastic processes (Fodelianakis et al. 2021). Fluctuations in the abundance of some

species within such communities may take place without any reference to environmental selection, and, hence, be purely due to random factors, when communities are in dynamic equilibrium. For instance, some algal species might rapidly occupy ecological niches due to initial accidental advantages (such as fewer predators or higher initial biomass), becoming dominant in the community, a phenomenon known as the priority effect (Debray et al. 2022; Reijenga et al. 2021). Gene flow is also part of stochastic processes, especially when there is genetic exchange between multiple communities (Arnold et al. 2022). Genetic diversity among species can be promoted by gene flow, and thus community structure and function (Irwin et al. 2022). Additionally, dispersal events are an important part of stochastic processes, particularly in open waters or large water bodies, where species dispersal is often unpredictable (Albright & Martiny 2018). Some algae may enter new habitats due to random dispersal events, profoundly impacting community structure. Deterministic and stochastic processes are not independent of each other but are intertwined and complementary (Skouroliakou et al. 2022). Environmental selection and species interactions usually collaborate with stochastic processes in shaping community structures in most ecosystems (Li et al. 2022b; Menéndez-Serra et al. 2023). For example, in ecosystems with stable environmental conditions, such as temperature and salinity, deterministic processes might be the dominant force, but in dynamic, fastchanging ecosystems, stochastic processes might play a larger role. In community succession, During community succession, deterministic and stochastic processes might gradually transition over time (Hanusch et al. 2022; Måren et al. 2017). During

the early stages of a community, stochastic processes may be the principal dynamic processes due to high species diversity and unstable environmental conditions. However, as the community reaches a stable phase, deterministic processes may become the dominating force as the impact of environmental factors on species selection becomes more pronounced. At this stage, the competitive and cooperative relationships among species will determine the final form of the community structure.

2.2 Advances in monitoring and detection techniques for HABs species

2.2.1 Traditional morphological identification

The identification and monitoring of microalgal species, particularly in the context of HABs, have historically relied on traditional morphological methods. Traditional morphological identification is based on the microscopic examination of physical and structural characteristics of algal cells (Manoylov 2014). Current taxonomy and ecology of algae have been based on this approach, for the most important information of species diversity and ecological functions. However, the efficacy of this method is limited by several factors. Firstly, the morphological features of many microalgae can be highly variable, depending on environmental conditions, which can lead to misidentification (Fawley & Fawley 2020). Moreover, some algal species exhibit minimal morphological variations, which complicates the identification of closely related species or different life stages within the same species (Krivina & Temraleeva 2020). Furthermore, traditional microscopy is labor intensive, requiring taxonomic

expertise, and can be time-consuming, which is impractical during urgent assessments of HABs where rapid response is critical.

2.2.2 DNA-based molecular techniques

Advances in molecular techniques have significantly transformed the landscape of algal research, offering new insights and overcoming many limitations inherent in traditional methods. The application of DNA barcoding as a tool for species identification has become routine use across the biological fields including phycology (Kowalska et al. 2019; Steinhagen et al. 2023). DNA barcoding refers to sequencing a short, standard length of DNA from a sample and matching it against reference sequences in a database (Antil et al. 2023). This technique allows for precise species identification, regardless of the life stage or condition of the organism, thereby addressing some of the key challenges of morphological methods. In microalgal research, commonly targeted genes for barcoding include the 16S rRNA for prokaryotic cyanobacteria and the 18S rRNA gene for eukaryotic algae (Karlusich et al. 2022).

Metabarcoding extends the principles of DNA barcoding to community-level assessments, enabling the identification of multiple species within environmental samples simultaneously (Jiménez-García et al. 2023). This technique combines high-throughput sequencing with specific molecular markers to identify and quantify the diversity of microbial communities. High-throughput sequencing (HTS) techniques,

including metabarcoding, offer significant advantages for the early detection and monitoring of HABs (Zaytsev et al. 2024). HTS can process numerous samples quickly and provide data on species composition and abundance at a much finer scale than traditional methods (Lopes Dos Santos et al. 2022). This rapid processing capability is crucial for the timely management of HABs, allowing for early warning and mitigation measures to be implemented before blooms reach harmful levels. Moreover, HTS can detect low-abundance species that might be overlooked by microscopy, providing an early indication of a potential bloom (Wang et al. 2022). The development of comprehensive, curated databases containing reference sequences is critical for the success of metabarcoding. Such databases allow researchers to assign taxonomy to the vast amounts of sequence data generated, facilitating rapid assessments of biodiversity and community structure. Databases such as the SILVA (ribosomal RNA sequences) and the Protist Ribosomal Reference database (PR2) have been instrumental in storing and providing access to reference sequences that enhance the accuracy and efficiency of species identification (Quast et al. 2013; Guillou et al. 2013).

In addition to 18S amplicon sequencing, metagenomic sequencing and metatranscriptome sequencing have also become essential molecular tools for analyzing the structural and functional characteristics of microalgal communities. 18S amplicon sequencing achieves high-throughput analysis of microalgae diversity and community composition in environmental samples by specifically amplifying the conserved region of the small subunit ribosomal RNA gene. This technology has

advantages of low cost, simple operation process and the ability to process large-scale samples, and is suitable for ecological monitoring, geographical distribution pattern analysis and time series research. However, limited by the species resolution of the 18S rRNA gene, some microalgae groups with similar systematic classification positions or highly conserved gene sequences are difficult to accurately distinguish by this method. In addition, the bias in the PCR amplification process may lead to bias in population abundance estimation, and this technology only provides "existence" information and cannot reveal the functional characteristics of microalgae. In contrast, metagenomic sequencing directly extracts the total DNA of all organisms in environmental samples without the need for specific primer amplification, thereby reducing amplification bias and being able to comprehensively depict the population composition of microalgae and their associated microorganisms, while analyzing their potential functional genomic characteristics (Aplakidou et al. 2024). Metagenomic data allows researchers to reconstruct partial or complete microalgae genomes (i.e., MAGs), and then conduct in-depth studies on metabolic pathways, environmental adaptation mechanisms, and microbial interaction networks. However, since environmental samples usually have extremely high biological complexity, especially in microalgae communities accompanied by a large number of bacteria, fungi, and viruses, the assembly and directional binning process of metagenomic data is extremely complex, and the coverage of low-abundance microalgae populations is often insufficient, resulting in the risk of information loss in subsequent analysis (Navgire et al. 2022). At the same time, the high cost of metagenomic sequencing also limits its application in large-scale sample screening. In contrast, metatranscriptome sequencing starts with total RNA, and high-throughput sequencing is performed after rRNA removal or mRNA enrichment, which can capture the active expression profiles of microalgae and related microorganisms in the environment under specific spatiotemporal conditions (Wahl et al. 2022). Metatranscriptome data not only reveals the level of "who is active", but also provides dynamic information at the functional level, which helps to analyze the transcriptional response of microalgae to environmental changes (such as light, temperature, salinity, nutrient changes, etc.), as well as the symbiotic, parasitic, or competitive relationship between microalgae and other microorganisms (Berman et al. 2020). However, since mRNA is easily degraded, metatranscriptomics have extremely high requirements for sample collection, preservation, and RNA extraction and purification. At the same time, there are still a large number of rRNA residues in environmental transcriptome data, resulting in a limited proportion of effective information. In addition, changes in gene expression levels may be due to environmental induction or fluctuations in community structure, which adds additional complexity to data interpretation. Therefore, in actual research design, the three technologies are often used in a reasonable combination according to research objectives and resource conditions: if the main focus is on the species diversity and community structure changes of microalgae, 18S amplicon sequencing is still the first choice with high cost performance; if the research focus is on the genomic characteristics and metabolic potential of microalgae and their associated microorganisms, metagenomic sequencing is more suitable; if it is necessary to reveal

the functional dynamics of microalgae under environmental stress or ecological interaction, metatranscriptome sequencing is indispensable. It is worth noting that with the development of sequencing technology and the continuous improvement of multi-omics joint analysis methods, future microalgae research will tend to integrate and analyze amplicon, metagenomic and metatranscriptome data to build a panoramic understanding from community structure to functional dynamics.

2.2.3 The database construction of HAB species

The downstream interpretation of molecular surveys relies heavily on the accuracy and completeness of the reference databases used for classifying taxonomy (Leray et al. 2022). The construction of comprehensive reference databases for harmful and toxigenic microalgae has emerged as a critical component in the effective monitoring and management of HABs. These databases not only facilitate accurate species identification but also enhance our understanding of the ecological impacts associated with algae. A comprehensive and accurate database should contain well-curated, vetted sequences that cover a wide range of known microalgal species, including those that are less common but potentially harmful. The quality of the database entries, including the correctness of the sequence data and the taxonomic resolution, directly impacts the success of molecular identification techniques such as DNA barcoding and metabarcoding. Mistakes in sequence annotation or the inclusion of poorly resolved sequences can lead to erroneous species identification, which in turn affects ecological

studies and management decisions.

Databases enable researchers to quickly identify the presence of harmful or toxic microalgae in water samples using molecular techniques. This capability is crucial for the early detection of HABs, allowing for timely interventions and mitigation strategies that can prevent ecological and economic damage. By providing detailed taxonomic and genetic information, these databases facilitate ecological studies that explore the distribution, diversity, and dynamics of algal populations. Understanding the impact of environmental changes on algal biodiversity requires robust baseline data, which wellmaintained databases can provide. Consequently, the development and application of databases for harmful and toxic microalgae are of great significance in addressing the global challenge posed by HABs. While there are significant challenges in achieving accurate species identification, the development of reliable and comprehensive databases is a critical step forward. These databases not only support effective monitoring and management but also enhance our scientific understanding of algal biology and ecology. As molecular techniques continue to advance, the importance of these databases will only grow, underscoring the need for ongoing investment in their development and maintenance.

2.3 Plastisphere: a novel microhabitat in coastal ecosystems

The pervasive presence of plastics poses a threat to various ecosystems around the

world. Their widespread presence is attributed to their ability to travel long distances by ocean currents or wind, reaching even the most remote areas of the planet. Plastic debris colonized by biofilm-forming microbial communities are known as the "plastisphere". The revelation that this unique substrate can aid microbial dispersal has piqued interest in the ground of microbial ecology.

2.3.1 Global plastic pollution and the formation of the plastisphere

The relentless accumulation of plastic waste in the world's oceans represents one of the most pressing environmental challenges of the 21st century (Geyer et al. 2017; Law 2017). Plastics have become the most common form of marine litter, with increasing amounts of post-consumer plastic waste inevitably entering the oceans from recycling and waste streams (MacLeod et al. 2021; Stubbins et al. 2021; Weiss et al. 2021; Li et al. 2021b). Plastic pollution is pervasive across all marine environments, from the poles to the equator and from the sea surface to the deep sea sediments (Kanhai et al. 2019). Plastic debris ranges in size from large, visible items to microplastics less than 5 millimeters in diameter. These plastics originate from both land-based and marine sources, including consumer and industrial waste that escapes waste management and treatment processes. Ocean currents carry these plastics across vast distances, leading to their accumulation in remote areas far from their original source and often converging in subtropical gyres where they form large floating "garbage patches" (Lebreton 2022). Currently, over 400 million tonnes of plastic waste are produced

globally each year. As estimated, approximately 5-13 million tonnes of plastic waste enter the ocean from coastal land annually (Jambeck et al. 2015). This causes more than 170 trillion pieces weighted approximately 2 million tonnes floating in the world's ocean (Eriksen et al. 2023). If no effective action, the plastic stock in the ocean could triple by 2040 (Lau et al. 2020).

The buoyancy, hydrophobic, organic, and durable surface of plastics in the environment provide a unique niche for microorganisms, resulting microecosystems with plastic as the matrix, termed "the plastisphere" (Amaral-Zettler et al. 2020; Zettler et al. 2013)... The huge stock of plastic waste in the environment, combined with the ongoing emission, indicates that the plastisphere is a habitat with a vast and expanding total area (Li et al. 2024a,c). It consists of a variety of microorganisms, including bacteria, algae, fungi, and small invertebrates (Wang et al. 2021a). These organisms colonize plastics in a biofilm, a robust layer of cells and extracellular matrix, adhering to the plastic surface. The formation of the plastisphere begins almost immediately after the plastic enters the water, with pioneer species attaching to the surface, followed by a succession of other organisms that form a complex, multi-species community. The community structure of the plastisphere is distinct from surrounding water column biota, often harboring species not commonly found in other marine environments (Pang et al. 2023). Growing evidence shows that the plastisphere may pose critical threats to the ecosystem function and biological health by enriching biogeochemical process-involved microorganisms as well as harmful species like disease-causing microorganisms and antibiotic resistant bacteria (Bergmann et al. 2022; Bhagwat et al. 2021; Dey et al. 2022). The unique conditions of the plastisphere, including the availability of new niche spaces and the presence of plastic additives, might encourage the evolution of novel biochemical pathways (Yu et al. 2023a). The durability of synthetic polymers, the building blocks of plastics, ensures that they persist in the marine environment for decades or even centuries. Unlike natural materials, plastics do not biodegrade but instead photodegrade under sunlight, breaking down into smaller and smaller pieces while never completely disappearing. This slow degradation process leads to the continuous accumulation of microplastics, which are ingested by marine organisms and can enter the food chain, leading to broader ecological consequences.

2.3.2 Potential ecological risks of the plastisphere

The ecological implications of the plastisphere are profound and multifaceted. Plastics, due to their durable and persistent nature, provide a unique substrate for microbial colonization in marine environments. These communities differ significantly from those found on natural substrates, partly due to the distinct chemical composition of plastics and their ability to persist and accumulate in marine environments. As the prevalence of plastic debris continues to increase globally, understanding the ecological implications of the plastisphere becomes critically important.

One of the primary ecological risks of the plastisphere is acting as a vector for invasive species. Plastics can travel long distances across oceans, carrying with them a wide range of colonizing organisms, which may not naturally occur in the habitats they reach. Habitat alterations can occur when non-native introduced species either compete with or displace indigenes marine life leading to the reduction of biodiversity and interference of local community structures. Moreover, animal, plant, and human pathogens are enriched in the plastisphere indicate that the plastisphere presents a significant biosecurity risk and a potential threat to human health (Li et al. 2024a). Pathogens harbored on plastic debris can infect animals, particularly aquatic species through filter feeding, adversely impacting their growth, behavior, and dietary habits (Jacob et al. 2020; Li et al. 2021a). There is evidence that even microplastics can be ingested and accumulate in terrestrial plants (Li et al. 2020). The exposure of plants to pathogens associated with the plastisphere could be a key factor behind reports of diminished crop yields(Zhang et al. 2020). Furthermore, reduced plant biomass has been linked to the presence of plastic contaminants (Zhang et al. 2022). The prevalence of pathogens in the plastisphere underscores an elevated risk of disease transmission. Plastics are known to absorb and concentrate persistent organic pollutants (POPs) from the marine environment, including PCBs, DDT, and PAHs (Koelmans et al. 2013; Wang et al. 2021b). Harmful compounds such as the phthalate plasticizers, bisphenol A, and metal additives such as zinc, copper, and nickel, can be released during plastic degrades process (Hahladakis et al. 2018). These chemicals can become part of the plastisphere and enter the food web through microorganisms that colonize plastics. This

incorporation leads to bioaccumulation and potentially biomagnification as toxins move up the food chain, reaching higher trophic levels and posing significant risks to marine wildlife. These pathogens and POPs might not only affect marine organisms but could also impact coastal communities and industries, such as fisheries and aquaculture, leading to economic losses and public health concerns. The complex interaction highlights the extensive impact of plastic pollution on ecosystem and human health (Li et al. 2024a).

Microbial communities within the plastisphere can significantly impact biogeochemical processes, particularly those related to carbon, nitrogen, and phosphorus cycling. The metabolic activities of these microorganisms can alter the rates of biogeochemical transformations in marine environments, potentially impacting nutrient availability, primary productivity, and overall ecosystem function. Regions with high plastic contamination tend to exhibit greater plastic biodegradation potential. Additionally, the accumulation of plastic residues has been shown to negatively impact soil organic matter in agricultural lands (Zhang et al. 2020). Another noteworthy discovery is that the higher degradation rates within the plastisphere may contribute to the release of greenhouse gases such as CO₂ and CH₄, which are by-products of plastic decomposition. The plastisphere also demonstrates enhanced potential for nitrogen-related processes, including denitrification, nitrogen respiration, and nitrate reduction, particularly within freshwater ecosystems (Li et al. 2024a). These processes increase the likelihood of

producing a potent greenhouse gas with a global warming potential, like N₂O (Kuypers et al. 2018; Repo et al. 2009). The metabolites released from the plastisphere into the surrounding environment disrupt the natural nutrient cycles. The potential for plastic to alter the metabolism of organic compounds, along with its significant influence on the nitrogen cycle, indicates that the plastisphere could disrupt biogeochemical processes and potentially exacerbate climate change (Bhagwat et al. 2021; Su et al. 2022). Beyond its ecological impacts, the accumulation of plastic waste and the ensuing formation of the plastisphere can also degrade the natural beauty of marine environments, potentially impacting tourism and recreational activities. Beaches and coastal areas littered with plastic waste are less appealing to tourists, which can result in significant economic losses for communities that rely on tourism revenue.

2.3.3 The HAB species in the plastisphere: emerging concerns

Previous studies on plastisphere microbiology and associated risks mainly focus on bacteria, with comparatively less attention given to fungi. Other life forms, such as microalgae, a key component of aquatic ecosystems, are poorly understood in the plastisphere. The convergence of two significant environmental issues, HABs and plastic pollution, presents new challenges for marine ecosystems globally. The plastisphere, a complex microbial community that colonizes floating plastic debris, provides a unique habitat that can facilitate the spread of harmful algal species and their toxins. In 2003, marine microplastics were recognized for the first time as potential

migration vectors for potential harmful dinoflagellates such as Ostreopsis sp. and Coolia sp. under macroscopic observation (Mercedes et al. 2003). One square centimeter of the plastic surface can harbor more than 80,000 diatom cells, pinpointing that marine plastispheres act as a suitable habitat for microalgae containing substantial microalgal biomass (Casabianca et al. 2019). The stable and durable nature of plastics offers a persistent platform for these algae, potentially allowing them to thrive and proliferate outside their usual geographic and temporal boundaries. The long-distance mobility of plastics across oceanic regions enables them to serve as carriers, transporting HAB species to previously unaffected areas, where these species may establish new outbreaks (Do Prado Leite et al. 2022). This capability is particularly concerning for regions previously unaffected by specific types of HABs, as it could lead to the introduction of new ecological threats to those ecosystems. Another critical concern is the accumulation of algal toxins on plastics. Studies have shown that plastics can adsorb and concentrate not only chemical pollutants but also biotoxins produced by toxic algae (Pires et al. 2025). This adsorption process can lead to higher concentrations of toxins on plastic surfaces than in surrounding waters, posing significant risks to marine life that ingest plastic debris. The ingestion of these toxin-loaded plastics by marine organisms can lead to bioaccumulation and biomagnification of toxins through the food web, impacting species up to and including humans. The risk is compounded by the potential for these toxins to be released into environments where they were previously absent, carried over vast distances by floating plastics. This scenario not only spreads the geographic impact of HABs but also complicates efforts to manage and mitigate their effects. Moreover, climate change adds a new dimension to this complex interaction. Global climate change exacerbates the uncertainty of marine environments, potentially altering basic environmental parameters such as ocean temperature, salinity, and circulation patterns, which in turn indirectly or directly affect the distribution and impact of plastic pollution and harmful algal blooms.

Chapter 3 Methodology

This section provides general information of database construction, sample collection, and detailed descriptions of the biological, statistical analyses and chemical applied in this study.

3.1 Construction of HTMaDB

3.1.1 Collection of HAB species

At the Fourth Session of the IOC Intergovernmental Panel on Harmful Algal Blooms, Vigo, Spain, 30 June-2 July 1997, it was decided to establish a Task Team on Algal Taxonomy with the aim of providing an agreed reference list of harmful algal species, including correct citation of the author(s), date of valid publication, and a list of synonyms. IOC-UNESCO Taxonomic Reference List of Harmful Micro Algae provides a reference for the use of names and information on each species of toxic microalgae. The list contains at the moment species producing toxins or toxic effects, and a few species suspected to form toxins. Note that a toxic species may not produce toxins everywhere it occurs. However, the reference list does not include species that cause harm due to high biomass. To include the harmful bloom-forming but not toxic species, we conducted literature review by searching publications with keywords "harmful microalgae" OR "toxic microalgae" OR "HAB species" OR "harmful and toxic algae" in Web of Science citation database Core Collection. After manually checking the collected 197 publications (Appendix 1), we further added 101 non-toxic but bloom causing species to the database. The detailed collected HAB species was provided in

3.1.2 18S sequence database construction for HAB species

3.1.2.1 Introduction of public databases

The International Nucleotide Sequence Database Collaboration (INSDC) is a global collaboration of independent governmental or non-profit organisations that manage nucleotide sequence databases capturing and preserving nucleotide sequence information and annotations to create a comprehensive collection that preserves the scientific record and enables broad sharing of such data (International Nucleotide Sequence Database Collaboration). INSDC Participating Databases including European Molecular Biology Laboratory (EMBL)-Bank, National Center for Biotechnology Information (NCBI)-GenBank, and DNA Data Bank of Japan (DDBJ). These three organizations exchange data on a daily basis. Although the three databases provide a wide range of sequence data, specialized databases such as SILVA and PR2 offer more in-depth and specialized resources in specific fields.

SILVA (from Latin silva, forest) database provides comprehensive, quality checked and regularly updated datasets of aligned small (16S/18S, SSU) and large subunit (23S/28S, LSU) ribosomal RNA (rRNA) sequences, which are widely used in microbial classification and ecological research (Quast et al. 2013). The SILVA databases are made available as releases, rather than being updated continuously, to enhance the

comparability of the studies employing these databases. Each release is numbered according to the EMBL-Bank release from which the sequence data were extracted and is permanently available for download via the SILVA website. The dataset encompasses a broad range of biological groups, including bacteria, archaea, and eukaryotes. The SILVA database includes taxonomic information, sequence data, and relevant functional annotations, facilitating users in conducting diversity analyses and phylogenetic studies.

The Protist Ribosomal Reference database (PR2) reference sequence database was initiated in 2010 in the frame of the BioMarks project from work that had developed in the previous ten years in the Plankton Group of the Station Biologique of Roscoff. The PR2 provides a unique access to eukaryotic small sub-unit (SSU) ribosomal RNA and DNA sequences, with curated taxonomy (Guillou et al. 2013). The database mainly consists of nuclear-encoded protistan sequences. However, metazoans, land plants, macrosporic fungi and eukaryotic organelles (mitochondrion, plastid and others) are also included because they are useful for the analysis of high-throughput sequencing data sets. Introns and putative chimeric sequences have been also carefully checked. Taxonomic assignation of sequences consists of nine unique taxonomic fields (from domain to species).

3.1.2.2 Collection of 18S full length from public databases

Advances in HTS have led to unprecedented growth in the amount of available sequencing data. Owing to existing reference databases such as SILVA and PR2, while widely used, exhibit notable limitations when it comes to the accurate identification and comprehensive coverage of harmful and toxic microalgae. Several key species known to be associated with HABs are entirely absent from these repositories. As a result, sequence alignment and taxonomic assignment using these databases may lead to incomplete characterization of harmful/toxic microalgal communities. To facilitate HAB research and related studies, we have developed the HTMaDB, comprehensive yet specific coverage of harmful and toxic algae taxa extracting from SILVA and PR2, which provides access to users for searching, downloading, aligning eukaryotic algal sequencing data. It offers enhanced taxonomic resolution and improved specificity for HAB-related organisms, thereby minimizing the risk of misidentification. The HTMaDB represents the first 18S full length sequence database for HAB species. Based on the scientific accepted species name, we extracted the full-length 18S sequences of the collected harmful and toxic taxa from eukaryotic microorganism databases (Silva 138.1 SSU and PR2 5.0.0 SSU Ref database). A total of 1346 sequences were obtained. We reorganized the taxonomic classification of the collected taxa by referring to AlgaeBase since inconsistent classification methods in different databases (M.D. Guiry & Guiry, G.M. 2024). The collected 18S full length sequences consist of 238 eukaryotic HAB species. The construction procedure was provided in Figure 3-1.

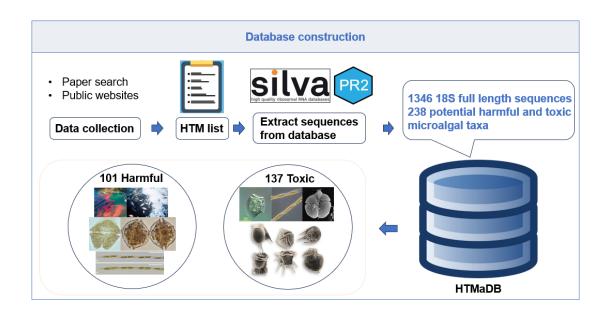


Figure 3-1 Flowchart of 18S sequence database construction of HTMaDB.

Table 3-1 The HAB species in the HTMaDB.

| уре | Species | | | | |
|-----|--|----------------------------|-------------------------------|---------------------------------|--------------------------------|
| | Akashiwo sanguinea | Coolia malayensis | Gambierdiscus toxicus | Prorocentrum concavum | Pseudo-nitzschia caciantha |
| | Alexandrium affine | Dinophysis acuminata | Gonyaulax bohaiensis | Prorocentrum cordatum | Pseudo-nitzschia calliantha |
| | Alexandrium andersonii | Dinophysis acuta | Gonyaulax spinifera | Prorocentrum emarginatum | Pseudo-nitzschia cuspidata |
| | Alexandrium australiense | Dinophysis caudata | Gymnodinium catenatum | Prorocentrum hoffmannianum | Pseudo-nitzschia delicatissima |
| | Alexandrium catenella | Dinophysis fortii | Heterocapsa bohaiensis | Prorocentrum leve | Pseudo-nitzschia fraudulenta |
| | Alexandrium fragae | Dinophysis infundibulum | Heterocapsa borneoensis | Prorocentrum lima | Pseudo-nitzschia fukuyoi |
| | Alexandrium hiranoi | Dinophysis miles | Heterocapsa circularisquama | Prorocentrum mexicanum | Pseudo-nitzschia galaxiae |
| | Alexandrium leei | Dinophysis norvegica | Heterocapsa horiguchii | Prorocentrum panamense | Pseudo-nitzschia granii |
| | Alexandrium limii | Dinophysis tripos | Heterocapsa niei | Prorocentrum rhathymum | Pseudo-nitzschia kodamae |
| | Alexandrium minutum | Fukuyoa paulensis | Heterocapsa pygmaea | Prorocentrum texanum | Pseudo-nitzschia lundholmiae |
| | Alexandrium monilatum | Fukuyoa ruetzleri | Karenia bicuneiformis | Protoceratium reticulatum | Pseudo-nitzschia multiseries |
| xic | Alexandrium ogatae | Fukuyoa yasumotoi | Karenia brevis | | Pseudo-nitzschia |
| | | | | Protoperidinium crassipes | pseudodelicatissima |
| | Alexandrium ostenfeldii | Gambierdiscus australes | Karenia mikimotoi | Pyrodinium bahamense | Pseudo-nitzschia pungens |
| | Alexandrium pacificum Gambierdiscus balechii Karenia | Karenia papilionacea | Pyrodinium bahamense var. | Pseudo-nitzschia seriata | |
| | | Кагена раршонасеа | compressum | | |
| | Alexandrium | Gambierdiscus belizeanus | Karenia selliformis | Takayama acrotrocha | Pseudo-nitzschia simulans |
| | pseudogonyaulax | Gambieraiscus belizeanus | | | |
| | Alexandrium tamarense | Gambierdiscus caribaeus | Karlodinium veneficum | Vicicitus globosus | Pseudo-nitzschia subcurvata |
| | Alexandrium tamiyavanichii | Gambierdiscus carolinianus | Lingulodinium polyedra | Chattonella marina | Pseudo-nitzschia turgidula |
| | Alexandrium taylorii | Gambierdiscus carpenteri | Margalefidinium fulvescens | Chattonella marina var. antiqua | Chrysochromulina leadbeater |
| | Amphidinium carterae | Gambierdiscus cheloniae | Margalefidinium polykrikoides | Chattonella marina var. ovata | Phaeocystis globosa |
| | Amphidinium gibbosum | Gambierdiscus excentricus | Ostreopsis ovata | Chattonella subsalsa | Phaeocystis pouchetii |

| | Amphidinium klebsii | Gambierdiscus holmesii | Ostreopsis rhodesiae | Fibrocapsa japonica | Prymnesium calathiferum |
|---------|------------------------------|-----------------------------|----------------------------|-----------------------------------|-----------------------------|
| | Amphidinium operculatum | Gambierdiscus honu | Ostreopsis siamensis | Heterosigma akashiwo | Prymnesium faveolatum |
| | Amphidoma languida | Gambierdiscus lapillus | Pfiesteria piscicida | Nitzschia bizertensis | Prymnesium parvum |
| | Azadinium dexteroporum | Gambierdiscus lewisii | Pfiesteria shumwayae | Pseudochattonella farcimen | Prymnesium polylepis |
| | Azadinium poporum | Gambierdiscus pacificus | Phalacroma mitra | Pseudochattonella verruculosa | Prymnesium zebrinum |
| | Azadinium spinosum | Gambierdiscus polynesiensis | Phalacroma rotundatum | Pseudo-nitzschia australis | |
| | Centrodinium punctatum | Gambierdiscus scabrosus | Polykrikos hartmannii | Pseudo-nitzschia batesiana | |
| | Coolia canariensis | Gambierdiscus silvae | Prorocentrum cassubicum | Pseudo-nitzschia brasiliana | |
| Harmful | Alexandrium cohorticula | Chaetoceros tenuissimus | Gymnodinium aureolum | Proboscia alata | Thalassiosira curviseriata |
| | Alexandrium margalefii | Chaetoceros throndsenii | Gymnodinium impudicum | Prorocentrum donghaiense | Thalassiosira rotula |
| | Alexandrium insuetum | Chaetoceros tortissimus | Protodinium simplex | Prorocentrum micans | Thalassiosira lundiana |
| | Asterionellopsis glacialis | Chaetoceros lorenzianus | Guinardia delicatula | Prorocentrum triestinum | Thalassiosira mala |
| | Aureococcus anophagefferens | Chaetoceros wighamii | Guinardia flaccida | Prorocentrum dentatum | Thalassiosira minima |
| | Aureoumbra lagunensis | Coscinodiscus radiatus | Halamphora coffeiformis | Prorocentrum redfieldii | Thalassiosira minuscula |
| | Blixaea quinquecornis | Coolia monotis | Heterocapsa triquetra | Prorocentrum gracile | Thalassiosira punctigera |
| | Chaetoceros affinis | Cylindrotheca closterium | Heterocapsa rotundata | Pseudocochlodinium profundisulcus | Thalassiosira weissflogii |
| | Chaetoceros atlanticus | Cyclotella meneghiniana | Heterocapsa pygmaea | Pyrophacus steinii | Thalassiosira pseudonana |
| | Chaetoceros constrictus | Dictyocha fibula | Lauderia annulata | Scrippsiella acuminata | Thalassiosira gravida |
| | Chaetoceros convolutus | Dictyocha speculum | Lepidodinium chlorophorum | Scrippsiella trochoidea | Thalassionema nitzschioides |
| | Chaetoceros curvisetus | Dinophysis hastata | Leptocylindrus danicus | Skeletonema tropicum | Tripos dens |
| | Chaetoceros danicus | Phalacroma rapa | Leptocylindrus minimus | Skeletonema costatum | Tripos furca |
| | Chaetoceros debilis | Diplopsalis lenticula | Levanderina fissa | Skeletonema marinoi | Tripos fusus |
| | Chaetoceros decipiens | Dunaliella salina | Lithodesmioides polymorpha | Skeletonema menzelii | Tripos lineatus |
| | Chaetoceros diadema | Eucampia zodiacus | Noctiluca scintillans | Skeletonema pseudocostatum | Tripos muelleri |
| | Chaetoceros laevisporus | Eutreptiella gymnastica | Nitzschia longissima | Takayama pulchella | Tryblionella compressa |
| | Chaetoceros pseudocurvisetus | Euglena viridis | Ostreopsis lenticularis | Takayama xiamenensis | |
| | | | | | |

| Chaetoceros rostratus | Gonyaulax polygramma | Paralia sulcata | Teleaulax acuta |
|-----------------------|----------------------|-----------------------------|-------------------------|
| Chaetoceros socialis | Gonyaulax verior | Plagioselmis prolonga | Tenuicylindrus belgicus |
| Chaetoceros simplex | Gonyaulax digitale | Plagioselmis nannoplanctica | Thalassiosira allenii |

3.1.3 Taxonomic composition and diversity of the HTMaDB

Among 238 eukaryotic HAB species. 137 toxic microalgae encompass Dinoflagellates (99), Ochrophyta (30), and Haptophyta (8) with, whereas 101 harmful microalgae consist of Dinoflagellates (37), Ochrophyta (56), Cryptophyta (3), Chlorophyta (1), and Euglenozoa (2) at phylum level (Fig.3-2a). This indicates that dinoflagellates are the dominant contributor of toxins, whereas harmful algal blooms are primarily caused by Ochrophyta.

Among the toxic and harmful genera in the database, Alexandrium is the most abundant, with an obvious higher sequence count than others (Fig.3-2b). This well-known toxic genus is associated with PSTs, posing severe risks to marine organisms such as shellfish and fish, as well as to human health. Gambierdiscus, ranking second in abundance, is linked to ciguatoxins, which are major contributors to food chain contamination, particularly in tropical and subtropical marine ecosystems. Skeletonema and Prorocentrum, ranked third and fourth, respectively, have distinct ecological impacts. While Skeletonema, primarily a diatom, can contribute to eutrophication in high concentrations, Prorocentrum is frequently associated with the production of algal toxins. Thalassiosira and Heterosigma also exhibit strong presence in the database. Thalassiosira, a diatom, generally does not produce toxins but can exacerbate eutrophication when overabundant, whereas Heterosigma is a known toxic genus capable of producing substances lethal to fish. Genera with lower abundance, such as Margalefidinium, Pfiesteria, and Cyclotella, are less frequent but still noteworthy. For instance, Pfiesteria is associated with neurotoxin production, posing a significant threat

to aquatic life, especially fish. This comprehensive assessment highlights the varying ecological and toxicological roles of these genera, emphasizing the importance of monitoring and managing their presence in marine ecosystems to mitigate potential risks.

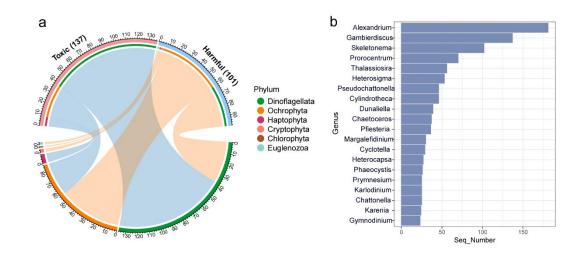


Figure 3-2 Taxonomic composition at phylum level (a) and top 20 genera based on sequence number (b) in constructed HTMaDB.

3.2 Marine sampling campaign

3.2.1 General information of the region studied

Hong Kong, a vibrant and densely populated coastal city in southern China, is known for its diverse marine ecosystems and thriving aquaculture industry. Its geographical location is at the mouth of the Pearl River in southern China, bordered by the South China Sea to the east and the Pearl River to the west, creating a diverse coastline that spans over 1,600 kilometers and includes various types such as bays, estuaries, beaches, and rocky shores. This supports a rich biological diversity, housing numerous species of fish, crustaceans, shellfish, and seaweed, making it an essential marine ecosystem.

Hong Kong located in the subtropical zone, significantly influenced by monsoon climate, with seasonal changes impacting the marine ecology and hydrological features.

3.2.2 Samling strategy

In this study, sample collection was first conducted in multiple sites under different areas during wet and dry seasons, to investigate the spatiotemporal distribution patterns and underlying mechanisms of the toxic algal community along Hong Kong's coastal waters using metabarcoding techniques. We collected seawater from 36 sampling sites of Hong Kong coastal ecosystems in wet and dry seasons (Figure 3-3 and Table 3-2). Surface seawater samples were collected using stainless-steel buckets and were transferred to 1-L polypropylene (PP) bottles. The bottles were covered with foil and kept in a portable icy incubator below 4°C during transportation to the laboratory. At each sampling location, a total of 5 L seawater samples were collected for the analysis of LATs (1 L), nutrients (1 L), phytoplankton density (1 L), and phytoplankton communities (2 L). Three bottles of Milli-Q water were also transported from the laboratory to the field site and then returned to the laboratory. These bottles served as field blanks to ensure negligible contamination. The samples were stored in a 4°C fridge and all samples were analyzed within 1 month. Detailed sampling information is provided in Appendix 2.

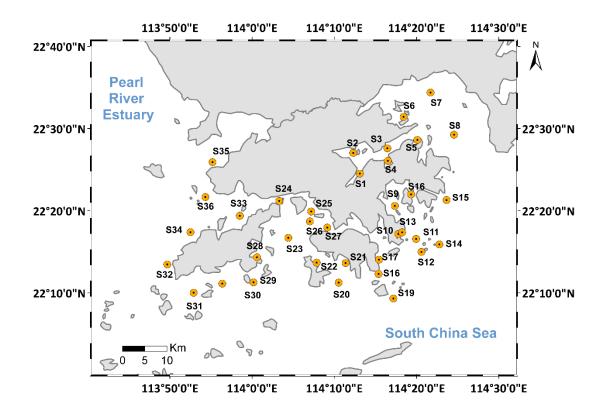


Figure 3-3 Sampling locations of surface seawater along the coastline of Hong Kong.

Table 3-2 Sampling period and number of samples in this study.

| Location | Location Sampling period | | Sample size |
|-----------------|---------------------------------|---------|-------------|
| Eastern Waters | Waters Wet season (August 2022) | | 30 |
| | Dry season (February 2023) | | |
| Southern Waters | Wet season (August 2022) | S16-S31 | 32 |
| | Dry season (February 2023) | | |
| Western Waters | Wet season (August 2022) | S32-S36 | 10 |
| | Dry season (February 2023) | | |

To reveal the ecological impacts of microalgae in coastal plastisphere, plastic debris and its ambient seawater samples were paired collected from coastal areas of Qingdao and Hong Kong, China, in August 2021 and February 2022. A Manta trawl (mesh size = 333 µm) was employed to collect plastic debris. A plastic sample was retrieved every

30 minutes of trawling. The collected plastic debris from the trawl was transferred into a 50-mL centrifuge tube. Simultaneously, 2 L of surface seawater was collected in a sterile glass bottle. All the samples were immediately stored in a 4 °C ice box. A total of 92 plastic samples and 92 ambient seawater samples were obtained during the field sampling. Samling information is listed in Table 3-3. Detailed information can be found in Appendix 2.

Table 3-3 Sampling period and number of samples in this study.

| Location | Sampling period | Sample type | Sample size |
|-----------|----------------------------|-------------|-------------|
| Hong Kong | Wet season (August 2021) | Plastic | 38 |
| | Dry season (February 2022) | | |
| | Wet season (August 2021) | Ambient | 38 |
| | Dry season (February 2022) | | |
| Qingdao | Wet season (August 2021) | Plastic | 54 |
| | Dry season (February 2022) | | |
| | Wet season (August 2021) | Ambient | 54 |
| | Dry season (February 2022) | | |

We visually confirmed that all collected debris pieces were plastic. To verify this objectively, we randomly selected a subset of debris for polymer identification via Midinfrared photothermal (MIP) microscopy, also called optical photothermal infrared (O-PTIR) microscopy, an emerging tool for bond-selective chemical imaging of living biological and material samples. O-PTIR spectroscopy is utilized to identify plastic polymer types through spectral analysis. Potential plastic samples are collected and mounted onto a glass slide without cleaning or peroxide treatment. A laser, set to 48% IR power and 4.4% probe power, induces photothermal effects to generate an infrared spectrum based on absorption characteristics. Spectra are recorded across 941–3007

cm⁻¹, including diagnostic bands such as C-H stretching (2900 cm⁻¹) and aromatic C-H bending (700 cm⁻¹). Polymer identification is performed by comparing spectral peaks to the KnowItAll library (Wiley). This method provides a non-destructive approach for identifying plastic polymers, enabling rapid characterization of samples. All tested pieces were confirmed as plastic (Appendix 3). According to our previous study, we propose a novel, conceptual framework of the "microplastome", encompassing the entirety of plastic particles and their associated entities, such as chemicals and microbes, within a given sample, along with their collective environmental and toxicological (Li et al. 2024b). In this study, we extracted DNA from all collected plastic samples, followed by sequencing, in order to provide an integrated perspective to capture the real-world ecology of plastics.

3.2.3 Determination of environmental parameters

Salinity, temperature, pH, and dissolved oxygen (DO)of seawater in each sampling site were measured *in situ* using the YSI Professional Plus Quatro water quality meter (YSI Incorporated, Yellow Springs, OH, USA). Dissolved organic carbon (DOC) was measured using a total organic carbon analyser (Elementar Acquray TOC cube, Frankfurt, Germany). Dissolved organic carbon (DOC) was measured using a total organic carbon analyser (Elementar Acquray TOC cube, Frankfurt, Germany). The concentrations of nutrients, including NH₄⁺, NO₂⁻, NO₃⁻ SiO₃²-, and PO₄³⁻ were determined in the laboratory by a continuous flow autoanalyser (Scalars San++, Skylar Analytical B.V., Breda, The Netherlands).

3.3 DNA extraction, sequencing and bioinformatic analysis

This section presents a comprehensive description of the molecular biological analyses employed in this study, containing the pretreatment of surface seawater samples, DNA extraction, library generation, 18S amplicon sequencing and (Figure 3-3).

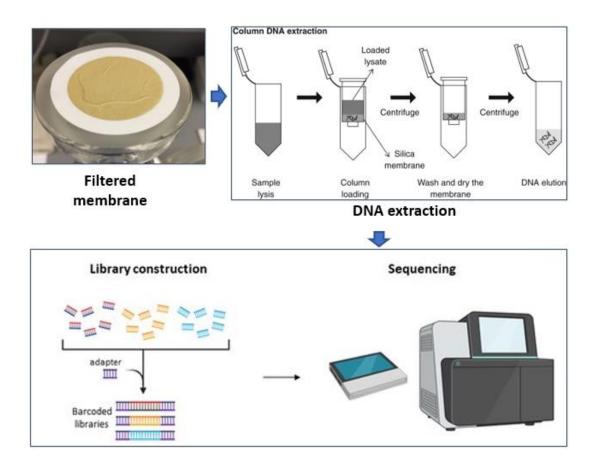


Figure 3-4 Flow chart of molecular biological analyses.

3.3.1 DNA extraction

For the extraction of total algal DNA, the seawater was filtered through a 0.45 µm MCE membrane. All filters were then stored at -80 °C until DNA extraction. The total DNA was extracted from the filters using FastSpinKit for soil (MP Biomedicals, Santa Ana, CA) following the manufacturer's instructions. Briefly, the membranes were cut with

sterile scissors and transferred to a lysing matrix E tube. Homogenize with the FastPrep instrument for 40 seconds at a speed setting of 6.0 m s⁻¹ with 978 µL sodium phosphate buffer and 122 µL MT buffer in it to mechanical disruption of algal cell walls and releasing nucleic acids into the protective buffer. The supernatant was transferred to a clean 2 mL microcentrifuge tube after centrifuge to pellet debris. Two hundred 50 µL of protein precipitation solution was added to separate the solubilized nucleic acids from the cellular debris and lysing matrix. The supernatant was transferred to a clean 15 mL tube after centrifuge to pellet precipitate for removal of flocculated proteins. Binding matrix solution (1 mL) was added for binding nucleic acids. The DNA solution was transferred to a spin filter tube and empty catch tube. The prepared SEWS-M solution was added to wash the spin filter for removing impurities by centrifuging through the filter while the purified DNA is still bound to the silica. The binding matrix in spin filter was resuspended in 100 µL elution buffer after the spin filter had been air dried for 5 minutes at room temperature for removal of residual ethanol. Purified DNA solution was ready after centrifuge. The concentration of DNA was measured using a NanoDrop One Microvolume UV-Vis Spectrophotometer (Thermo Fisher Scientific, Waltman, MA, USA).

3.3.2 18S amplicon sequencing and bioinformatic analysis

The *18S rDNA V9* region was amplified using the universal primers: 1380F (5'-CCCTGCCHTTTGTACACAC-3') and 1510R (5'-CCTTCYGCAGGTTCACCTAC-3') (Cheung et al. 2010; Zimmermann et al. 2011). The amplicons were paired-end sequenced using an DNBSEQ-G400 platform. Paired-end sequences were merged using USEARCH v10.0.240 (Edgar & Bateman 2010). Primer-cut from merged reads

and quality control was performed using "fastx_filter" command by VSEARCH v2.15 (Rognes et al. 2016). The quality-filtered reads of each sample were dereplicated using the "derep_fulllength" command, and less than 8 reads were removed from all samples as the potential sequence errors. The amplicon sequence variants (ASVs) were generated by denoising with unoise3 in USEARCH (Labouyrie et al. 2023; Li et al. 2022a). Taxonomic information of ASVs was annotated based on the SILVA database (Quast et al. 2013). Phytoplankton taxonomic assignment of ASVs was classified against the Silva Database (Release 123) (Liu et al. 2020). The harmful and toxic taxa in our samples were identified by aligning our sequences to the HTMaDB with a threshold of 0.8 using VSEARCH. Rarefaction curves were generated to assess the degree of sample saturation using picante and vegan function packages of R4.2.1 (Figure 3-4 and Figure 3-5).

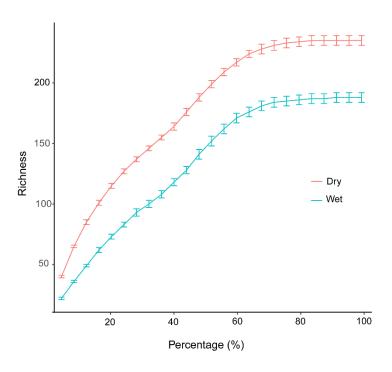


Figure 3-5 Rarefaction curves of the richness in the wet and dry seasons reach the

saturation stage with means and standard errors under 1% - 100%, indicating that the number of samples in our study is sufficient to capture most microorganisms in each season.

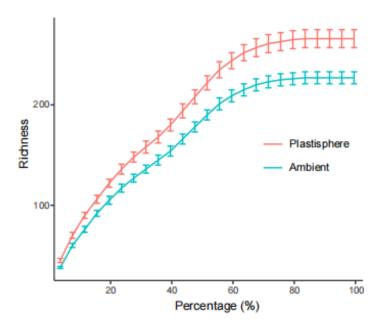


Figure 3-6 Rarefaction curves of detected phytoplankton richness from the plastisphere and the ambient environment in coastal ecosystem reach saturation stage with increasing sequencing depth.

3.4 Statistical analyses

3.4.1 The difference of the phytoplankton community in Hong Kong coastal seawater and plastisphere

Principal coordinate analysis (PCoA) based on Bray-Curtis dissimilarity was performed to evaluate the spatiotemporal pattern of phytoplankton community structure in Hong Kong coastal seawater. Permutational multivariate analysis of variance (PERMANOVA) was used to investigate differences in phytoplankton communities between sampling

seasons. Similarly, PCoA was utilized to illustrate the compositional difference between the plastisphere microalgal community and the ambient community, with the PERMANOVA used to test the statistical significance. The richness and Shannon indexes were calculated using the "vegan" package to reflect the alpha diversity of microalgal communities. Furthermore, Bray-Curtis distances between each two samples in each group were calculated to reflect the beta diversity of the communities with the "vegan" package. Differences in the abundances of ASVs between the plastisphere and the ambient microalgal community were analyzed using the "edgeR" package and were depicted in a Manhattan plot. For the identified harmful and toxic algal taxa, the differences of their abundances in the plastisphere community and the ambient community were assessed using the Wilcoxon rank-sum test.

3.4.2 Quantifying ecological stochasticity and niche breadth in the microalgal community

Understanding the community assembly mechanisms controlling biodiversity patterns is a central issue in ecology. An index, normalized stochasticity ratio (NST) was developed to quantify ecological stochasticity under different situations by considering abiotic filtering, competition, environmental noise, and spatial scales (Ning et al. 2019). Accuracy and precision of stochasticity estimation decreased when application to a global scale. Modified Stochasticity Ratio (MST) is a special form of NST, making it more general (Liang et al. 2020). Niche breadth is an important indicator of species adaptability to the environment within a community (von Meijenfeldt et al. 2023). Ecological niche breadth was calculated the Levins' niche breadth index (Levins 1968).

$$B_i = \frac{1}{\sum_{j=1}^{r} (P_{ij})^2}$$

Equation 3-1

In the above equation, *Bi* refers to the niche breadth of species i. *Pij* means the proportion of species i at a given location j. r is the number of locations. If a species has a higher B value, it indicates that the species has a greater niche breadth, with a wider, more even, and larger distribution. Conversely, a lower B value for a species suggests that its distribution is less frequent and more uneven (Pandit et al. 2009).

The MST and habitat niche breadth were estimated to reveal the underlying community assembly mechanisms of microalgal community using "NST" package. The MST model is typically applied to quantify the relative importance of stochasticity and determinism in the community assembly process. The values of MST range from 0 to 1, with 0.5 as the boundary defining deterministic (MST < 0.5) or stochastic (MST > 0.5) dominated assembly processes (Ning et al. 2019). To uncover the patterns of assembly process and their influence on microalgal communities across different habitats, we further calculated the niche breadth index of each species in the community. The habitat niche breadth was estimated based on the average niche width across all species within a community (Jiao et al. 2020; Richard 1968; Wu et al. 2018a). The analysis was conducted using the "niche.width" function in "spaa" package in R.

3.4.3 Identification of specialist algal taxa

Specificity refers to the average abundance of species (S) in the samples of a group (H);

and occupancy represents the relative frequency of occurrence of S in the samples of H (See Equation 3-2). Nindividual_{S,H} represents the average number number of individual species S across all samples in group H, while Nindividual_S is the total mean number of individual S across all groups; Nsites_{S,H} refers to the count of samples in H where S is present, whereas Nsites_H refers to the total number of samples in H (Dufrêne & Legendre 1997; Gweon et al. 2021).

$$Specificity = \frac{Nindividuals_{S,H}}{Nindividuals_{S}}$$

$$Occupancy = \frac{Nsites_{S,H}}{Nsites_H}$$

Equation 3-2

In our study, we calculated the specificity and occupancy of each toxic species across seasons to identify specialist taxa. Specificity is defined as the mean abundance of a species in the samples of a season, while occupancy is characterized as the relative frequency with which species occur within the same season. Species with specificity and occupancy of 0.7 or higher were classified as specialist species, indicating their preference for a particular season (Kang et al. 2024).

3.4.4 Associations between environmental factors and microalgal community

The correlation between toxic taxa and environmental factors in Hong Kong coastal seawater was demonstrated based on the spearman test. Redundancy analysis (RDA) was employed to identify potential environmental driving factors of the plastisphere

microalgal community and its ambient community. The Mantel test was carried out to investigate the potential driving environmental factors of plastisphere microalgal communities. Linear regression models were employed to determine relationships between the plastisphere microalgal risk and environmental factors. Procrustes analysis was performed to analyze the associations between the plastisphere community and the ambient community.

3.4.5 Quantifying the relative contribution of biotic and abiotic factors to microalgal community

Variation Partitioning Analysis (VPA) is a statistical technique used in ecology to assess the contribution of various environmental factors to the variability in biological data (Figure 3-10). Variation partitioning analysis results are typically shown using a Venn diagram, displaying the variance percentage explained by each data set. In a case where we are partitioning the variation among two explanatory matrices, the result could be represented as Figure 3-11. Fraction [a + b + c] represents the explained variance by both X1 and X2, calculated through an RDA of Y with both X1 + X2. Fraction [d] indicates the unexplained variance by X1 and X2, derived from the same RDA. Fraction [a] represents the explained variance by X1 alone, calculated through a partial RDA of Y by X1|X2 (controlling for X2). Fraction [c] represents the explained variance by X2 alone, calculated using a partial RDA of Y by X2|X1 (controlling for X1). Fraction [b] is determined by subtraction: b = [a + b] + [b + c] - [a + b + c]. Because [b] is not the result of an RDA, it cannot be tested for significance. It can also be negative, which indicates that the response matrix is better explained by the combination of X1 and X2 than by either matrix on its own (Peres-Neto et al. 2006; Tedersoo et al. 2016; Lai et al. 2022). Here we employed VPA to explain the contribution of the community variations

to co-occurring species interaction and environmental factors using the 'vegan' package in R (Labouyrie et al. 2023; Tedersoo et al. 2016).

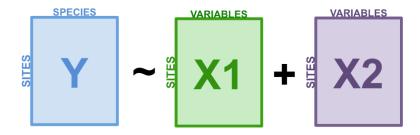


Figure 3-7 The basic structure of variation partitioning.

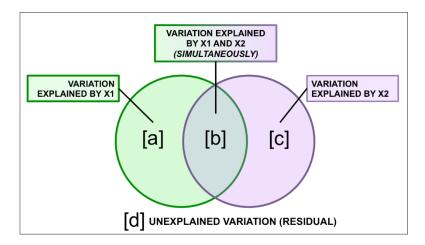


Figure 3-8 Representing variation partitioning results.

3.5 Chemical analysis

3.5.1 Methods for extracting LATs

For LATs extraction, seawater samples were filtered through 0.45 µm glass microfiber filters (GF/A, Whatman Schleicher & Schuell, Maidstone, England) to remove visible particulate matter and algal cells. SPE was performed using Oasis HLB cartridges (200 mg, 6 mL, Waters, Medford, MA, USA), following the procedure outlined in a previous study (Li et al. 2014). The cartridges were pre-conditioned with 4 mL of ammonium hydroxide/methanol (3:1000, v:v), followed by 4 mL of methanol and 4 mL of

deionized water. Then, 250 mL of seawater samples were passed through each cartridge at a flow rate of 1 mL·min-1. Afterward, the cartridges were rinsed with 4 mL of methanol/water (15:85, v:v) and dried by centrifugation at 3500 rpm for 2 minutes. The extracts were obtained by eluting the cartridge three times with 3 mL of ammonium hydroxide/methanol (3:1000, v:v). The extracts were then evaporated until dry under a gentle nitrogen stream at 40°C, reconstituted with 100 μL of methanol, centrifuged at 3500 rpm for 3 minutes, and transferred to an analysis vial. For extraction of SPM, each filter loaded with suspended particulate matter was cut off in pieces and placed in 50 mL centrifuges. Samples were extracted with 8 mL 26 mM Ammonia-MeOH by 30 mins ultrasonication. The liquid from the centrifuge tube was transferred to a 5 mL syringe, filtered through a PTFE filter membrane (0.22μm) into a centrifuge tube (15 mL). Conduct other 2 extractions with 5 mL 26 mM Ammonia-MeOH separately. The extract was dried under a gentle nitrogen stream, reconstituted with 100 μL of methanol, and filtered through a 0.22 μm PTFE filter before being transferred to an analysis vial. The samples were then stored at -20°C until further analysis.

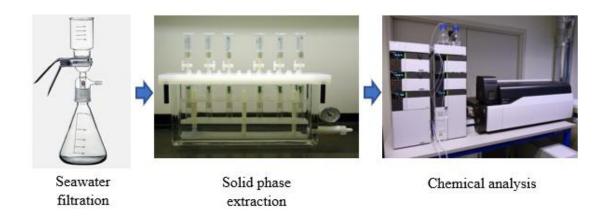


Figure 3-9 Surface seawater treatment for algal toxins.

The standards of gymnodium (GYM), okadaic acid (OA), pectenotoxin-2 (PTX-2),

dinophysistoxin-1 and -2 (DTX-1 and -2), and azaspiracid-1 and -2 (AZA-1 and -2) were purchased from the National Research Council, Institute for Marine Biosciences (Halifax, Nova Scotia, Canada). The properties of algal toxins were shown in Table 3-4. The standards were stored at -20°C. Acetonitrile and methanol of gradient grade were purchased from Merck (Darmstadt, Germany). Ammonium hydroxide (≥ 25%) of MS grade was purchased from SUPELCO (Bellefonte, PA, USA). Ultrapure water was obtained from a Milli-Q water-purification system (Millipore, Billerica, MA, USA).

Table 3-4 Molecular weight, molecular formula, and octanol-water partition coefficient of the LATs.

| Compound | CAS number | Molecular Weight | Molecular Formula | Log Kow |
|----------|-------------|------------------------|---|-------------------------|
| | | (g·mol ⁻¹) | | (Takahashi et al. 2007) |
| OA | 78111-17-8 | 805.0 | $C_{44}H_{68}O_{13}$ | 5.05 |
| DTX-1 | 81720-10-7 | 819.0 | $C_{45}H_{70}O_{13}$ | 6.88 |
| DTX-2 | 139933-46-3 | 805.0 | $C_{44}H_{68}O_{13}$ | 5.61 |
| PTX-2 | 97564-91-5 | 857.0 | $C_{47}H_{70}O_{14}$ | 6.47 |
| GYM | 173792-58-0 | 507.7 | C ₃₂ H ₄₅ NO ₄ | 6.64 |
| AZA-1 | 214899-21-5 | 842.1 | $C_{47}H_{71}NO_{12}$ | 7.54 |
| AZA-2 | 265996-92-7 | 856.1 | $C_{48}H_{73}NO_{12}$ | 8.18 |

3.5.2 Instrumental analysis of LATs

Separation of the seven LATs in seawater samples was conducted on an ExionLC UHPLC system (Sciex, Foster City, CA, USA) with an ACQUITY UPLC BEH C18 Column (2.1 mm \times 50 mm, 1.7 μ m particles, Waters, Medford, MA, USA), maintained at 40°C. The gradient condition began with 5% solvent B, increasing to 45% at 3.5 minutes, followed by a linear rise to 55% at 6.5 minutes, and finally reaching 100% at 7.5 minutes. After 1 minute, the gradient returned to the initial conditions at 9.5 minutes and was held for 2 minutes. The total run time was 11.5 minutes, with 10 μ L injected for each sample. During analysis, the autosampler's sample illumination was turned off.

The MS method was developed and optimized for maximum sensitivity using the 6500+ QTRAP (García-Altares et al. 2013). Instrumental data were acquired and processed with Analyst 1.6.3 software (Sciex, Foster City, CA, USA). Chromatographic analysis employed dual fragment ions derived from parent ions for compound verification, with quantification based on the higher-intensity transition. Retention time alignment combined with parent-fragment ion transitions provided three-dimensional confirmation of chromatographic peaks. Seawater analyte concentrations were determined using external calibration curves matching established reference standards. Detailed parameters can be found in Table 3-5 and Table 3-6. The QA/QC measures included evaluating for method limits of detection (LODs), quantification (LOQs), calibration curves, procedure blanks. Each sample was extracted and analyzed in duplicate, and for every batch of real samples, procedural blank (n = 2), and blank-spiked recoveries (n = 2) were also included.

Table 3-5 Source parameters in negative mode and positive mode.

| Parameters | Negative mode | Positive mode |
|------------------|---------------|---------------|
| Curtain Gas | 25 | 25 |
| Collision Gas | Low | Low |
| Voltage (V) | -4500 | 5000 |
| Temperature (°C) | 550 | 550 |
| Gas 1 (psi) | 60 | 60 |
| Gas 2 (psi) | 60 | 60 |

Table 3-6 Optimized mass spectrometry parameters for LATs detection: Transitions monitored, dwell times, declustering potentials (DP), entrance potentials (EP), collision cell entrance potentials (CEP) and collisions energies (CE)

| Compound | Transitions (m/z) | Time (ms) | DP (V) | EP (V) | CE (V) | CXP (V) | Precursor ion |
|----------|-------------------|-----------|-----------|-----------|-----------|------------|-----------------------------------|
| OA | 803.1>255.1 | 40 | 140 | -10 | -60 | -20 | [M-H] ⁻ |
| OA | 803.1>563.1 | 40 | 140 | -10 | -58 | -25 | [141-11] |
| DTX-2 | 803.5>255.2 | 40 | 140 | -10 | -60 | -38 | LW F11- |
| D1X-2 | 803.5>563.2 | 40 | 140 | -10 | -58 | -17 | [M-H] ⁻ |
| DTX-1 | 817.6>255.2 | 40 | 140 | -10 | -60 | -15 | LW F11- |
| DIA-I | 817.6>563.2 | 40 | 140 | -10 | -58 | -30 | [M-H] ⁻ |
| YTX | 570.4>467.2 | 40 | 140 | -10 | -40 | -28 | [M-2H] ²⁻ |
| 117 | 570.4>501.8 | 40 | 140 | -10 | -31 | -30 | [101-211] |
| | 508.3>490.3 | 40 | 140 | 10 | 32 | 26 | |
| GYM | 508.3>392.3 | 40 | 140 | 10 | 46 | 30 | $[M+H]^+$ |
| | 508.3>202.4 | 40 | 140 | 10 | 50 | 10 | |
| PTX-2 | 876.6>823.3 | 40 | 140 | 10 | 34 | 20 | [M+NH ₄] ⁺ |
| 1 1 A-2 | 876.6>805.3 | 40 | 140 | 10 | 37 | 20 | [1V1 + 1N114] |

| AZA-1 | 842.5>824.6 | 40 | 140 | 10 | 43 | 27 | $[M+H]^+$ |
|-------|-------------|----|-----|----|----|----|--------------------|
| AZA-I | 842.5>806.2 | 40 | 140 | 10 | 46 | 35 | [101+11] |
| AZA-2 | 856.7>838.4 | 40 | 140 | 10 | 41 | 38 | [M+H] ⁺ |
| ALA-L | 856.7>820.5 | 40 | 140 | 10 | 49 | 40 | |
| | | | | | | | |

To determine the LODs and LOQs for LATs, we conducted the same extraction procedure with real environment samples on 11 blank samples. The standard deviation (SD) of the signals was calculated, LOD and LOQ can be estimated to 3 times and 10 times of the standard deviation respectively (Table 3-7). Calibration curves for the seven LATs had concentrations (seven data points) with correlation coefficients ranging from 0.9977 to 0.9999. Recoveries of the 7 LATs varied from 52.9% to 92.5% in seawater and from 89.3% to 116.3% in suspended particulate matters (SPM) samples (Figure 3-10). Importantly, none of the 7 target LATs were found in procedure blanks. The reported LATs concentrations were not surrogate recovery corrected.

Table 3-7 Limits of detection, limits of quantification, correlation coefficients of the external calibration (r^2) , and recoveries of the targeted LATs.

| LATs | LOD (pg L ⁻¹) | LOQ (pg L ⁻¹) | r ² | Recovery Mean (SW) | Recovery Mean (SPM) |
|-------|------------------------------|------------------------------|----------------|--------------------|---------------------|
| PTX-2 | 2.83 | 9.43 | 0.9977 | 81.2% | 116% |
| AZA-2 | 0.58 | 1.94 | 0.9988 | 52.9% | 91.2% |
| GYM | 1.15 | 3.83 | 0.9993 | 66.3% | 89.3% |
| AZA-1 | 0.66 | 2.20 | 0.9999 | 56.3% | 101% |
| OA | 9.60 | 32.0 | 0.9984 | 84.7% | 105% |
| DTX-1 | 3.12 | 10.4 | 0.9988 | 63.0% | 96.9% |
| DTX-2 | 2.54 | 8.46 | 0.9980 | 92.5% | 108% |

Recovery Test Suspended Particulate Matters Seawater Suspended Particulate Matters Seawater PTX-2 AZA-2 GYM AZA-1 OA DTX-1 DTX-2

Figure 3-10 Recovery of targeted LATs.

Chapter 4 Ecological pattern of microalgal communities in coastal ecosystems – An application of HTMaDB in Hong Kong

For better identification of harmful and toxic microalgae, we constructed a HTMaDB, which contained 1346 18S full-length reference sequences. With marine field sampling campaign, the spatiotemporal variations in microalgal communities were investigated along Hong Kong coastal zone during wet (August 2022) and dry seasons (February 2023). The phytoplankton community was far more affected by seasonal change compared with spatial distribution. Based on the application of HTMaDB, we revealed the driving factors of toxic algal community and potential causative organisms of targeted toxins.

4.1 Distinct seasonal pattern of phytoplankton community across Hong Kong coastal seawater

A total of 1,649 ASVs were identified as eukaryotic microalgae based on amplicon data. PCoA and PERMANOVA was conducted to discern disparities in spatiotemporal distribution of the structure of the eukaryotic phytoplankton community in Hong Kong coastal surface seawater (Figure 3-1a). Our results unveiled a notable differentiation in taxonomic compositions among two sampling seasons compared to different regions, suggesting a greater influence of seasonality over spatial factors on a small localize scale ($R^2_{\text{season}} = 0.3164 > R^2_{\text{region}} = 0.057$). Given this seasonal variation, we further characterized taxonomic richness across the two sampling seasons. An increase was observed in taxonomic richness in dry season compared to wet season (Figure 4-1b).

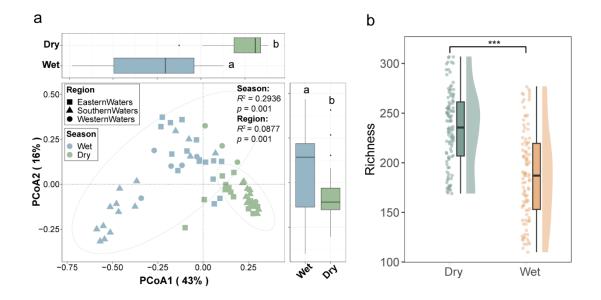


Figure 4-1 Significant seasonal pattern of community structure. **a**, Unconstrained principal coordinate analysis (PCoA) with permutational multivariate analysis of variance (PERMANOVA) showing a significant compositional difference between the wet and dry seasons. **b**, Differences in the Richness index between the wet and dry seasons (***p < 0.001; Wilcoxon rank-sum test).

Distinct patterns in phytoplankton composition during different seasons (Figure 4-2). The eukaryotic phytoplankton were categorized into 7 phyla (Dinoflagellata, Ochrophyta, Chlorophyta, Haptophyta, Cryptophyta, Rhodophyta, and Cercozoa). Dinoflagellates and Ochrophyta predominated in taxonomic composition, accounting for more than 70% in both seasons. The relative abundance of dinoflagellates was highest in the dry season, whereas Ochrophyta peaked in the wet season.

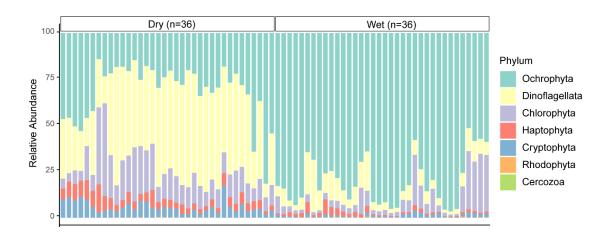


Figure 4-2 Taxonomic composition of phytoplankton community.

4.2 The deterministic process shaping phytoplankton community

Community assembly describes how different ecological processes shape microbial community composition and structure. Modified Stochasticity Ratio (MST) was calculated based on the null model to assess the relative importance of deterministic and stochastic processes. The importance of phytoplankton communities was shaped by different assembly processes during two sampling periods (Figure 4-3).

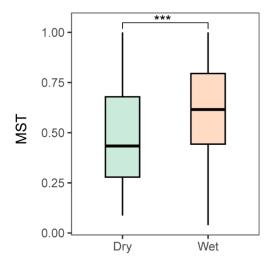


Figure 4-3 Assembly process based on modified stochasticity ratio (MST). The MST index has a boundary of 50% to distinguish between more deterministic (<50%) and more stochastic (>50%) assembly (***p < 0.001; Wilcoxon rank-sum test).

4.2 Mapping harmful and toxic microalgae using HTMaDB

Out of 1,649 ASVs, a total of 403 were identified as eukaryotic harmful and toxic microalgae in Hong Kong coastal seawater, belonging to the 9 class Dinophyceae, Diatomea, Coccolithophyceae, Raphidophyceae, Prymnesiophyceae, Chlorophyceae, Cryptophyceae, Pelagophyceae, and Dictyochophyceae (Figure 4-4a). Dinophyceae and Diatomea were the dominant class of toxic and harmful taxa, respectively. In the case of toxic ASVs, Coccolithophyceae represented a significant proportion as well. For harmful ASVs, both Chlorophyceae and Cryptophyceae also constituted considerable proportions. The relative abundance of 192 toxic ASVs and 211 harmful ASVs displayed different seasonal patterns. Toxic algae were prevalent in the dry season, whereas harmful taxa were relatively stable across seasons (Figure 4-4b).

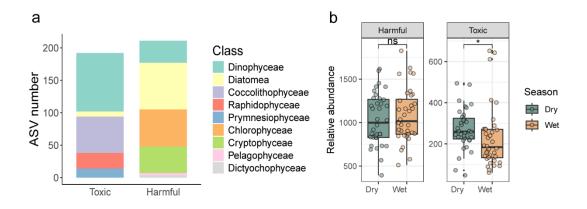


Figure 4-4 The composition and relative abundance of identified harmful and toxic ASVs using HTMaDB. The proportion of classified ASV numbers at class level (a), significantly higher relative abundance of all toxic taxa in dry season (b).

The abundance of 22 harmful and toxic algal genera in wet and dry seasons was showed in Figure 4-5. Among HAB algal taxa, the diatom in the genus of *Pseudo-nitzschia*

displayed the highest absolute abundance, indicating the potential of bloom forming and toxin contamination of DA especially in the wet season. A higher number of dinoflagellates were identified for the genera of *Alexandrium*, *Dinophysis*, *Gymnodinium*, and *Prorocentrum*. Additionally, the genera of *Chrysochromulina*, *Phaeocystis*, and *Prymnesium* in Haptophyta were also detected. These harmful and toxic algal taxa result in varied harmful effects on marine ecosystems. The occurrence of water discolorations and oxygen depletion owed to harmful bloom-forming species, for example, most diatom taxa and some of the dinoflagellates.

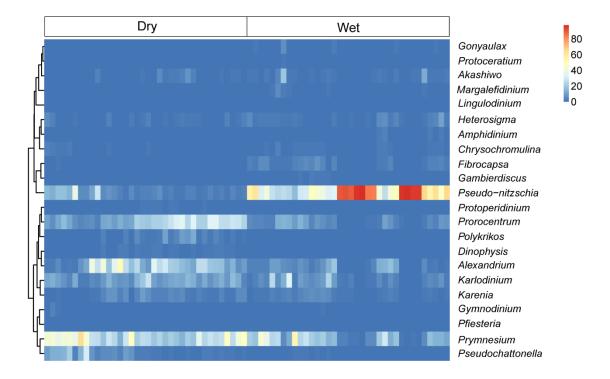


Figure 4-5 Temporal distribution of harmful and toxic taxa at genus level in two seasons.

The identified 22 HAB genera belonged to the 30 toxic and 49 harmful algal species (Table 4-1). Toxic algal taxa pose a threat to fish and shellfish and may cause mass mortality in aquaculture organisms, such as ichthyotoxic dinoflagellates

Margalefidinium polykrikoides and M. fulvescens (previously known as Cochlodinium polykrikoides and C. fulvescens) (Hofmann et al. 2021). Alexandrium spp. are the major producers of PSP toxin (Valbi et al. 2019). A total of twenty-six ASVs were attributed to five species of Alxandrium, all of them (including A. andersonii, A. hiranoi, A. leei, A. ostenfeldii, and A. tamarense) were proven to be PSP producers (Lassus et al. 2016). Pseudo-nitzschia spp., the only producers of domoic acid (DA), such as Pseudonitzschia australis and P. pungens, can lead to ASP (Wang et al. 2023). Dinophysis spp. and *Prorocentrum* spp. are source organisms of PTX-2, OA and its derivatives DTX-1, which are responsible for DSP (Marzidovšek et al. 2024). Azadinium. spp were associated with AZP. Karenia mikimotoi can produce the brevetoxins, which cause NSP (Li et al. 2024d). Diverse yessotoxin (YTX) producers, including Gonyaulax spinifera and Protoceratium reticulatum, were identified in coastal waters. The widespread species Prymnesium parvum, primarily in coastal waters, but also found in rivers and marine environments, have been reported with large-scale fish mortality events globally (Wagstaff et al. 2021). Furthermore, recent study indicates that prymnesins, rather than organic micropollutants, possess strong in vitro neurotoxic effects (Escher et al. 2024). A total of 49 harmful algal taxa were identified, most of them are diatoms.

 Table 4-1 The identified HAB species in coastal seawater.

| Phylum | Species | Impacts | Phylum | Species | Impacts |
|-----------|----------------------------------|---------|-------------|--------------------------------|---------|
| | Akashiwo sanguinea | Toxic | | Heterosigma akashiwo | Toxic |
| | Alexandrium andersonii | Toxic | | Fibrocapsa japonica | Toxic |
| | Alexandrium hiranoi | Toxic | | Pseudo-nitzschia australis | Toxic |
| | Alexandrium leei | Toxic | | Pseudo-nitzschia cuspidata | Toxic |
| | Amphidinium klebsii | Toxic | | Pseudo-nitzschia delicatissima | Toxic |
| | Alexandrium ostenfeldii | Toxic | | Pseudochattonella verruculosa | Toxic |
| | Alexandrium tamarense | Toxic | | Asterionellopsis glacialis | Harmful |
| | Amphidinium klebsii | Toxic | | Aureococcus anophagefferens | Harmful |
| Dinanhyta | Dinophysis acuminata | Toxic | Ochmonhysto | Chaetoceros affinis | Harmful |
| Dinophyta | Dinophysis miles | Toxic | Ochrophyta | Chaetoceros debilis | Harmful |
| | Gambierdiscus scabrosus | Toxic | | Chaetoceros diadema | Harmful |
| | Gonyaulax spinifera | Toxic | | Chaetoceros lorenzianus | Harmful |
| | Gymnodinium catenatum | Toxic | | Chaetoceros pseudocurvisetus | Harmful |
| | Karenia mikimotoi | Toxic | | Chaetoceros rostratus | Harmful |
| | Karlodinium veneficum | Toxic | | Chaetoceros simplex | Harmful |
| | Lingulodinium polyedra | Toxic | | Chaetoceros socialis | Harmful |
| | Margalefidinium fulvescens Toxic | | | Chaetoceros tenuissimus | Harmful |
| | Margalefidinium polykrikoides | Toxic | | Chaetoceros throndsenii | Harmful |

| | Pfiesteria piscicida | Toxic | Coscinodiscus radiatus | Harmful |
|-------------|-----------------------------------|---------|----------------------------|---------|
| | Polykrikos hartmannii | Toxic | Cyclotella meneghiniana | Harmful |
| | Prorocentrum cordatum | Toxic | Cylindrotheca closterium | Harmful |
| | Prorocentrum rhathymum | Toxic | Guinardia flaccida | Harmful |
| | Protoceratium reticulatum | Toxic | Lepidodinium chlorophorum | Harmful |
| | Protoperidinium crassipes | Toxic | Levanderina fissa | Harmful |
| | Gonyaulax polygramma | Harmful | Dictyocha fibula | Harmful |
| | Gymnodinium impudicum | Harmful | Eucampia zodiacus | Harmful |
| | Heterocapsa rotundata | Harmful | Lauderia annulata | Harmful |
| | Heterocapsa triquetra | Harmful | Leptocylindrus danicus | Harmful |
| | Noctiluca scintillans | Harmful | Nitzschia longissima | Harmful |
| | Pseudocochlodinium profundisulcus | Harmful | Paralia sulcata | Harmful |
| | Prorocentrum triestinum | Harmful | Proboscia alata | Harmful |
| | Protodinium simplex | Harmful | Skeletonema marinoi | Harmful |
| | Pyrophacus steinii | Harmful | Tenuicylindrus belgicus | Harmful |
| | Tripos furca | Harmful | Thalassiosira allenii | Harmful |
| | Prymnesium parvum | Toxic | Thalassiosira curviseriata | Harmful |
| Haptophyta | Chrysochromulina leadbeateri | Toxic | Thalassiosira mala | Harmful |
| | Prymnesium polylepis | Toxic | Thalassiosira pseudonana | Harmful |
| Cryptophyta | Plagioselmis prolonga | Harmful | Thalassiosira punctigera | Harmful |
| | | | | |

| | Teleaulax acuta | Harmful | Thalassiosira weissflogii | Harmful |
|-------------|-------------------|---------|---------------------------|---------|
| Chlorophyta | Dunaliella salina | Harmful | | |

4.3 Deterministic processes driven by temperature shaping toxic algal communities

To explore the complexities of toxic algal communities and the assembly processes shaping them, it is crucial to delve deeper into how these processes are modulated by environmental stress, particularly temperature fluctuations. Deterministic processes are pivotal in defining the community structure of phytoplankton by favoring species that can adapt to prevailing conditions, thus reducing community stochasticity and increasing predictability in species composition.

4.3.1 The deterministic process shaping toxic algal community

As mentioned above, toxic algae displayed a significant seasonal difference in sampling period. The proportion of deterministic process increased with increasing abundance of toxic microalgae from wet season to dry season (Figure 4-3). The difference in toxic microalgae abundance was significantly negatively correlated with stochasticity for the toxic algal community ($R^2 < 0.01$, p = 0.025), suggesting the importance of deterministic process increased with the toxic microalgae stress, which align with the theoretical framework of deterministic assembly increases as stress increases (Figure 4-6) (Ning et al. 2024). High stress usually imposes strong selective pressure. As stress increases, many species will be more suppressed, but those with higher tolerance or adaptation to the stressor(s) will thrive, leading to more deterministic community assembly. Determinism indicates that selection is the primary force that shapes microbiomes, including environmental filtering and species interactions (Kang et al. 2024).

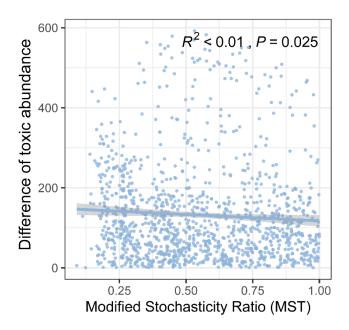


Figure 4-6 Relationships of the stochasticity and the difference of toxic algal abundance.

Environmental factors such as temperature and light directly influence the growth and metabolic activities of algae (Deutsch et al. 2008; Liao 2024). When environmental conditions shift to favor the growth of specific toxic algae, these organisms can rapidly proliferate, thereby increasing their abundance within the community. For instance, toxic algae, such as *Dinophysis* spp., flourish in warm water conditions (Fiorendino et al. 2020). Nutrient conditions also exert significant selective pressure. An increase in specific nutrients, such as phosphates and nitrates, can enhance the growth of certain toxic algae, which may be more sensitive to or capable of more effectively utilizing resources (Bonilla 2023). When referring to the species interaction, toxic algae reduce the risk of predation by producing toxins, a defensive mechanism that can decrease predator pressure and allow toxic algae to occupy a larger proportion within the community (Borrell 2010).

Moreover, some toxic algae engage in symbiotic relationships with other microorganisms, such as bacteria (Higashi et al. 2016). These relationships are mutually beneficial, as the associated microorganisms can assist the algae in more efficiently utilizing available nutrients or even provide other survival benefits (Fei et al. 2025; Gajardo et al. 2023). Such symbiotic interactions not only help toxic algae thrive in their existing environments but also enhance their adaptability to changing ecological conditions, thereby reinforcing their presence in the ecosystem.

4.3.2 The relationship between environmental factors and toxic algal community

Network correlation test was applied to further elucidate the major influencing environmental variables, including salinity, pH, temperature, DO, NO₂-, NO₃-, SiO₃²-, PO₄³-, and NH₄+. Significant correlations were identified linking toxic algal taxa with temperature, salinity, pH, and concentrations of PO₄³- and SiO₃²- (Figure 4-7a). Responses to environmental drivers varied among different algal taxa.

Temperature emerged as the most influential factor among all the environmental parameters evaluated, with 5 toxic algal taxa, including the genera *Dinophysis*, *Fibrocapsa*, *Herterosigma*, *Protoperidinium*, and *Pseudo-nitzschia*, primarily displaying a positive correlation with temperature except for the *Dinophysis*. The genus of dinoflagellates, *Dinophysis*, were more abundant at the cooler temperatures observed during the dry season (18–20°C) may owing to their optimal growth at temperatures ranging from 18 to 24°C (Fiorendino et al. 2020).

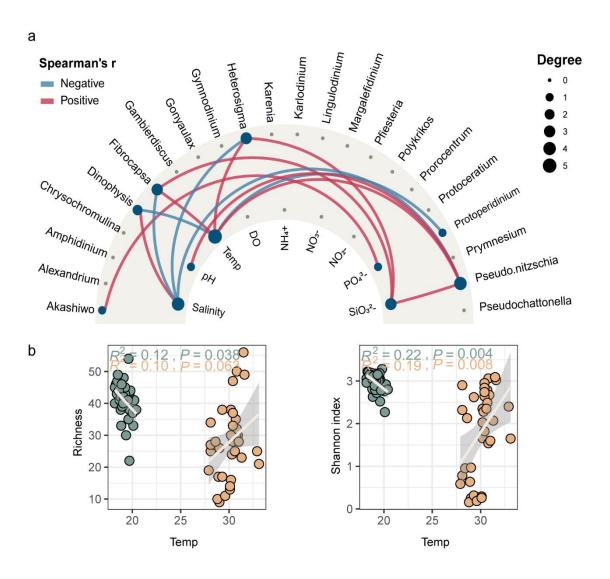


Figure 4-7 Major driving environmental factors shaping toxic community. (a) Relationships between identified communities and environmental factors. When p<0.05, a significant correlation is represented by a line. Different colors are used to distinguish between positive and negative correlations (red: positive; blue: negative). (b) The size of the circles represents magnitude of the correlation coefficient, with larger circles indicating a greater impact. Relationships between temperature and toxic algal community alpha diversity.

Growth responses towards temperature change were very variable and remained

inconsistent across the different algal groups, species and strains, reflecting evolutionary constraints and past adaptations to its environmental regime (Boyd et al. 2013; Brandenburg et al. 2019). Species that thrive at higher optimal growth temperatures typically exhibit faster growth under favorable conditions compared to those with lower optimal temperatures (Suzuki & Takahashi). When the temperature exceeds a species' optimal range, growth rates tend to decline sharply (García et al. 2018). However, evolution surveys have demonstrated that phytoplankton species can rapidly adapt to higher temperatures. Diatom species have increased their temperature optima by 1 °C, with adjust their metabolic pathways to optimize resource utilization at different temperatures (Hattich et al. 2024). This adaptation driven by evolutionary changes in baseline gene expression, a process known as transcriptional investment or divestment (Liang et al. 2019).

Linear regression analysis also confirmed that temperature significantly affected the alpha diversity of the toxic algal community (Figure 4-7b). Under elevated temperature stress, both species richness and Shannon diversity index of toxic algal community displayed a significant positive correlation. This suggests that increasing environmental stress fosters toxic algal assemblages, potentially due to the enhanced selection for thermotolerant and opportunistic species that thrive under changing conditions. The negative correlations between 3 algal taxa with salinity levels were observed, suggesting the possibility of presence of salinity stress. The growth of diatoms (*Pseudonitzschia*), Dinoflagellate (*Akashiwo*), and Heterokontophyta (*Herterosigma*) were promoted with increasing nutrients of PO₄³⁻ and SiO₃²⁻ on a limited scale. The risk of HABs and toxin contamination could be more severe by eutrophication of coastal zones.

Biotic filtering processes such as species competition, facilitation and predation, and abiotic filtering where functional differences between individual species play a central role (Yuan et al. 2021). The occupancy and specificity analysis revealed that the numbers of keystone toxic species in dry and wet season were 25 and 7, respectively (Figure 4-8 and Table 4-2). Specialist species in the dry season were mostly dinoflagellates, including *Alexandrium*, *Dinophysis*, *Karlodinium*, *Polykrikos*, and *Prorocentrum*, whereas *Pseudo-nitzschia*, belonging to diatoms accounted for nearly half of specialist species in the wet season. The increasing abundance of toxic algal community in the dry season could be attributed to these characteristic dinoflagellate species.

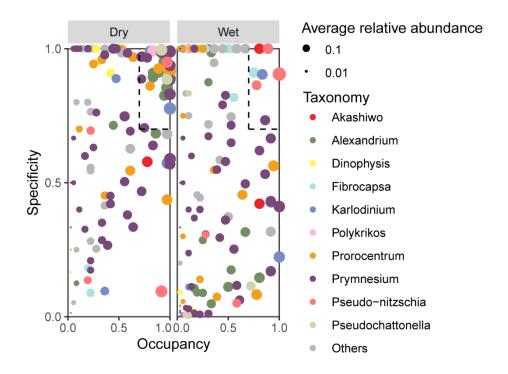


Figure 4-8 The specificity-occupancy plot shows the distribution and specificity of the abundant toxic species with a mean relative abundance higher than 0.01% in each season. Species with specificity and occupancy greater or equal to 0.7 are specialist

species.

Table 4-2 Specialist species in the dry and wet seasons.

| NO. | Season | OTU | Specificity | Occupancy | Species |
|-----|--------|----------|-------------|-------------|--------------------------------|
| 1 | Wet | ASV_342 | 1 | 0.80555556 | Akashiwo sanguinea |
| 2 | Wet | ASV_1576 | 0.910852713 | 0.75 | Fibrocapsa japonica |
| 3 | Wet | ASV_292 | 0.904225352 | 0.833333333 | Karlodinium veneficum |
| 4 | Wet | ASV_438 | 0.733333333 | 0.861111111 | Prymnesium parvum |
| 5 | Wet | ASV_337 | 1 | 0.88888889 | Pseudo-nitzschia cuspidata |
| 6 | Wet | ASV_641 | 0.863829787 | 0.77777778 | Pseudo-nitzschia cuspidata |
| 7 | Wet | ASV_13 | 0.905371556 | 1 | Pseudo-nitzschia delicatissima |
| 8 | Dry | ASV_209 | 0.951785714 | 0.94444444 | Alexandrium andersonii |
| 9 | Dry | ASV_634 | 0.91111111 | 0.88888889 | Alexandrium andersonii |
| 10 | Dry | ASV_115 | 0.854388635 | 0.972222222 | Alexandrium hiranoi |
| 11 | Dry | ASV_153 | 0.888530928 | 1 | Alexandrium tamarense |
| 12 | Dry | ASV_525 | 0.825174825 | 0.94444444 | Alexandrium tamarense |
| 13 | Dry | ASV_574 | 0.909482759 | 0.833333333 | Alexandrium tamarense |
| 14 | Dry | ASV_431 | 0.893548387 | 0.833333333 | Alexandrium tamarense |
| 15 | Dry | ASV_726 | 0.974358974 | 0.75 | Dinophysis miles |
| 16 | Dry | ASV_701 | 0.991836735 | 0.80555556 | Karenia mikimotoi |
| 17 | Dry | ASV_290 | 0.776990649 | 1 | Karlodinium veneficum |
| 18 | Dry | ASV_460 | 0.992025518 | 0.833333333 | Polykrikos hartmannii |
| 19 | Dry | ASV_503 | 0.925816024 | 0.88888889 | Prorocentrum cordatum |
| 20 | Dry | ASV_989 | 0.860606061 | 0.80555556 | Prorocentrum cordatum |
| 21 | Dry | ASV_70 | 0.917433538 | 1 | Prorocentrum rhathymum |
| 22 | Dry | ASV_117 | 0.830131827 | 1 | Prymnesium parvum |
| 23 | Dry | ASV_163 | 0.937427578 | 1 | Prymnesium parvum |
| 24 | Dry | ASV_111 | 0.988690476 | 1 | Prymnesium parvum |
| 25 | Dry | ASV_353 | 0.985887097 | 0.916666667 | Prymnesium parvum |
| 26 | Dry | ASV_173 | 0.990983607 | 1 | Prymnesium parvum |
| 27 | Dry | ASV_451 | 0.882926829 | 1 | Prymnesium parvum |
| 28 | Dry | ASV_763 | 0.966292135 | 0.77777778 | Prymnesium parvum |
| 29 | Dry | ASV_881 | 0.704761905 | 0.75 | Prymnesium parvum |
| 30 | Dry | ASV_1178 | 0.887096774 | 0.72222222 | Prymnesium parvum |
| 31 | Dry | ASV_308 | 0.970260223 | 0.72222222 | Prymnesium polylepis |

| 32 | Dry | ASV_225 | 0.949579832 | 0.972222222 | Pseudo-nitzschia australis |
|----|-----|---------|-------------|-------------|-------------------------------|
| 33 | Dry | ASV_105 | 0.995830438 | 0.916666667 | Pseudochattonella verruculosa |
| 34 | Dry | ASV_254 | 0.88558952 | 0.972222222 | Pseudochattonella verruculosa |

4.3.2 The contribution of environmental factors and non-toxic species interactions to toxic algal community variation.

VPA was used to further quantify the contribution of environmental factors and nontoxic species interactions to toxic algal community variation. A combined 47% of the variation in toxic algal community could be explained, with non-toxic algal taxa accounting for 12%, environmental factors for 2%, together for 33%; the remaining 53% of variation was unexplained (Figure 4-9). The unaccounted variability in toxic algal distribution, as indicated by the VPA, might also be influenced by complex interactions between **Symbiotic** relationships viruses. bacteria. and algae. between Planctomycetaceae bacteria and diatoms can play a significant role in the resurgence of algal blooms (Ma et al. 2022). These bacteria not only provide essential nutrients that aid algal growth but may also indirectly affect algae by modifying environmental conditions through changes in bacterial physiological traits (Seymour et al. 2017). Therefore, a more thorough investigation into the myriad interactions that impact toxic taxa is essential. To effectively protect marine ecosystems and ensure public health and safety, a comprehensive understanding of the composition, ecological dynamics, and toxin production mechanisms of harmful algae, as well as the pathways through which they spread, is crucial.

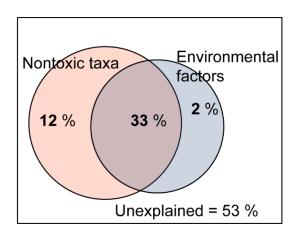


Figure 4-9 Variation partitioning showed the effects (%) of co-occurring nontoxic algal taxa and environmental factors (temperature, salinity, pH, DO, and nutrients) on the spatiotemporal distribution of toxic community.

4.4 Potential causative taxa of targeted LATs

Toxic algal taxa exhibited the same seasonal pattern as targeted LATs (Figure 4-10). LATs are of high priority due to their bioaccumulation effects. We further investigated the levels of typical LATs in surface seawaters of Hong Kong. Out of the seven targeted LATs, PTX-2, OA, GYM, and DTX-1 were present in Hong Kong waters. In contrast, DTX-2, AZA-1, and AZA-2 were undetected at all sampling sites (Appendix 3). The majority of LATs were detected in the dissolved seawater phase, with the only exception of PTX-2, which was more prevalent in SPM during the dry season. Within the dissolved phase, PTX-2 and OA emerged as the dominant toxins, collectively accounting for approximately 80% of the total concentrations of the detected LATs. These patterns mirror those found in the nearby waters of the Pearl River Estuary (Liu et al. 2020) and the northern South China Sea (Li et al. 2022a). Notably, PTX-2 and OA concentrations surged during the dry season compared to the wet season. DTX-1 was observed exclusively in the dry season. GYM concentrations remained no significant seasonal fluctuation. Notably, *Dinophysis* spp. are the sole known producers of PTX-2,

and they exhibit optimal growth at temperatures ranging from 18 to 24°C, which could explain the higher concentrations of PTX-2 observed during the dry season. In contrast, OA and DTX-1 maintain stability in seawater, resulting in less pronounced seasonal variations in their distribution when compared to the distribution pattern of PTX-2 (García-Altares et al. 2016).

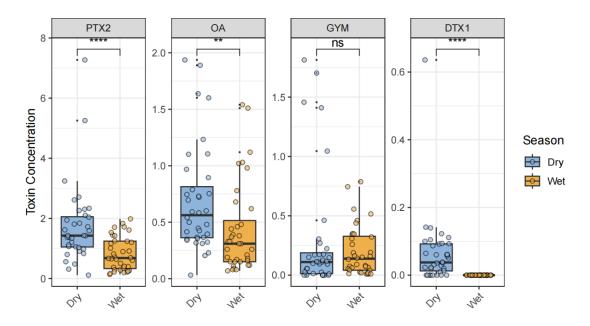


Figure 4-10 Seasonal patterns of targeted LATs.

PTXs are synthesized by toxic *Dinophysis* species, which are also capable of producing OA and DTXs (Reguera et al. 2012). Six ASVs related to *Dinophysis* spp. were identified as *D. acuminata* and *D. mils*. PTX-2, in particular, has been linked to *D. acuminata* and *D. mils*. These potential producers have been observed in Hong Kong waters, their toxin profiles are not yet fully understood (Lu & Hodgkiss 2004). OA and DTX-1 are primarily produced by *Dinophysis* spp., and epibenthic dinoflagellates *Prorocentrum* (Lee et al. 2020). ASVs linked to *Prorocentrum* species in this study, including *P. cordatum* and *P. rhathymum*, *P. rhathymum* have been reported to produce

OA. GYM toxins are associated with *Karenia mikimotoi* (formerly *Gymnodinium mikimotoi*) and *Alexandrium ostenfeldii* in our study (Molgó et al. 2017). Toxic algae related to detected toxins were summarized in Table 4-3. The relative abundance of toxic algal species was significantly higher in the dry season, sharing the same seasonal pattern with corresponding LATs (Figure 4-11). Accordingly, these can be assumed as potential causative organisms of LATs in Hong Kong coastal waters.

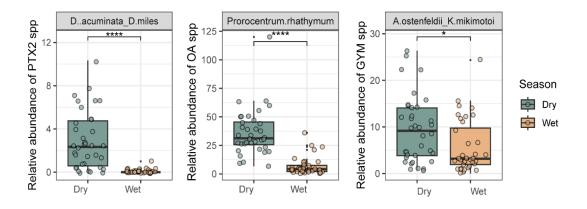


Figure 4-11 Seasonal patterns of potential producers.

Beyond the primary focus on specific LATs and their direct producers, it is imperative to also consider the broader influence of other harmful algal species present in Hong Kong's marine waters. For example, species within the *Prorocentrum* group have been linked to fish mortality events, causing hypoxia or anoxia that can lead to massive fish kills (Lu & Hodgkiss 2004). Equally concerning is the production of ciguatoxins by *Gambierdiscus scabrosus*, which are notable for their bioaccumulation through the marine food web-from herbivorous to carnivorous reef fish-and the consequent health risks they pose to humans upon consumption of contaminated fish (Pisapia et al. 2017). Further research is necessary to identify the toxin profiles of the prevailing algal species in the region.

 Table 4-3 Toxic algae related to detected toxins.

| Phylum | Class | Order | Family | Genus | Species | Toxin | References |
|----------------|-------------|----------------|-----------------|--------------|---------------|---------------|--------------------------|
| Dinoflagellate | Dinophyceae | Dinophysiales | Dinophysaceae | Dinophysis | D. acuta | OA, | (Fux et al. 2010) |
| | | | | | | DTX-2, PTX-2, | |
| | | | | | | PYX-2sa | |
| | | | | | | OA, | (Lee et al. 1989) |
| | | | | | | DTX-1 | |
| | | | | | D. acuminata | OA | (Lee et al. 1989) |
| | | | | | | DTX-1, | (Kamiyama & Suzuki 2009) |
| | | | | | | PTX-2 | |
| | | | | | D.caudate | OA, PTX-2, | (Fernández et al. 2006) |
| | | | | | | PTX-2sa | |
| | | | | | D. fortii | OA, | (Lee et al. 1989) |
| | | | | | | DTX-1, | |
| | | | | | | PTX-2 | |
| | | | | | D. miles | OA, DTX-1 | (Marasigan et al. 2001) |
| | | | | | D. sacculus | OA, DTX-1 | (Giacobbe et al. 2000) |
| | | | | | D. norvegica | OA, | (Lee et al. 1989) |
| | | | | | | DTX-1 | |
| | | | | | | PTX-2 | (Suzuki et al. 2009) |
| | | | | | D. mitra | DTX-1 | (Lee et al. 1989) |
| | | | | | D. rotundata | DTX-1 | (Lee et al. 1989) |
| | | | | | D. tripos | DTX-1 | (Lee et al. 1989) |
| | | Prorocentrales | Prorocentraceae | Prorocentrum | P. lima | OA | (Murakami et al. 1982) |
| | | | | | | OA, | (Lee et al. 1989) |
| | | | | | T. 0 | DTX-1 | (1.5. (1.000) |
| | | | | | P. faustiae | OA, | (Morton 1998) |
| | | | | | D | DTX-1 | (7 1 2020) |
| | | | | | P. concavum | OA | (Zou et al. 2020) |
| | | | | | P. | OA | (Accoroni et al. 2018) |
| | | | | | hoffmannianum | | |

| | | | P. maculosum | OA | (Zhou & Fritz 1993) |
|---------------|-----------------|-------------|-----------------|-----------|--------------------------------|
| | | | P. belizeanum | OA | (Cruz et al. 2006) |
| | | | P. porosum | OA | (Arteaga-Sogamoso et al. 2023) |
| | | | P. rhathymum | OA | (Luo et al. 2017) |
| Gymnodiniales | Kareniaceae | Karenia | K. selliformis | GYM | (Seki et al. 1995) |
| | | | (Gymnodinium | GYM, | (Miles et al. 2000) |
| | | | selliforme,) | GYM-B | |
| Gonyaulacales | Ostreopsidaceae | Alexandrium | A. ostenfeldii | 12-methyl | (Wagoner et al. 2011) |
| | | | (A. peruvianum) | GYM | |
| | | | | GYM, | (Waal et al. 2015) |
| | | | | 12-methyl | |
| | | | | GYM | |

4.5 Summary

This chapter investigated the distribution dynamics of microalgae communities, mainly focused on harmful and toxic algal community in coastal seawater using the self-constructed HTMaDB to verify reliability. The major findings are as follows:

A total of 1,649 amplicon sequence variants (ASVs) representing eukaryotic microalgae were identified. Principal coordinate analysis (PCoA) revealed that seasonal variation $(R^2 season = 0.3164)$ played a greater role than spatial variation $(R^2 region = 0.057)$ in shaping the phytoplankton community structure. Taxonomic richness was significantly higher during the dry season than the wet season, with dinoflagellates and Ochrophyta dominating both seasons. Dinoflagellates were more abundant in the dry season, while Ochrophyta peaked during the wet season. The modified stochasticity ratio (MST) suggested that deterministic processes, such as environmental filtering, became more pronounced with increasing toxic algal abundance. Out of the 1,649 ASVs, 403 were identified as harmful and toxic microalgae, belonging to nine classes, including Dinophyceae and Diatomea as dominant classes. Toxic algae, such as *Pseudo-nitzschia* spp., Alexandrium spp., and Dinophysis spp., were more prevalent in the dry season, whereas harmful, non-toxic taxa were relatively stable across seasons. Species like Pseudo-nitzschia australis, Alexandrium ostenfeldii, and Prorocentrum cordatum were identified as major toxin producers, contributing to PSP, ASP, and DSP. These harmful and toxic taxa have significant ecological impacts, including hypoxia, anoxia, and fish mortality.

The abundance of toxic algae increased during the dry season, correlating with temperature-driven deterministic processes. Key environmental factors influencing the toxic algal community included temperature, salinity, pH, and nutrients such as phosphate (PO4³¬) and silicate (SiO₃²¬). Temperature emerged as the most influential factor, with some toxic genera, such as *Dinophysis*, *Pseudo-nitzschia*, and *Fibrocapsa*, displaying positive correlations with temperature. Controlling inputs of nutrients like phosphates could effectively suppress the proliferation of harmful algae such as Akashiwo, reducing bloom risks. Monitoring salinity changes could aid in predicting the distribution of specific algal taxa. Regulating key environmental factors could mitigate threats posed by algal toxins to aquaculture and human health. Specialist species such as *Alexandrium* and *Prorocentrum* were more abundant in the dry season, driving the seasonal increase in toxic algal abundance. VPA showed that 47% of the variation in the toxic algal community could be attributed to environmental factors (2%) and non-toxic algal interactions (12%), with the remaining 53% unexplained, possibly due to complex interactions with bacteria and viruses.

The strong alignment between toxin profiles and the relative abundance of toxic taxa highlights the robustness and reliability of our curated dataset. Beyond the targeted toxins, other harmful species such as *Gambierdiscus scabrosus*, associated with ciguatoxins, and hypoxia-inducing diatoms and dinoflagellates were identified. These species can have cascading ecological impacts, including bioaccumulation through food webs and large-scale fish mortality events. Improved identification of harmful and toxic species (via HTMaDB) and understanding of their environmental triggers enable early warning systems, supporting the development of effective management and

mitigation strategies to minimize their impacts on ecosystems and human health.

Chapter 5 Microalgae and associated risks in coastal plastisphere

With the application of our newly constructed database, we explored the spatiotemporal distribution pattern and major driving factors of toxic microalgal communities in coastal ecosystem in chapter 4. The continued input of plastic waste into the ocean can alter marine microalgal community. Microbial risks stemming from the plastisphere (a microecosystem with plastics as the matrix) intensify the complexity and danger of marine plastic pollution. To further explore the ecological pattern and risk of microalgal community in the plastisphere, we collected 92 paired samples from Hong Kong and an additional coastal city, Qingdao, providing insights into the role of plastic debris as a novel habitat for microalgae in this chapter. Based on the amplicon sequencing data, the contrast in microalgal community structure among plastisphere and ambient are discussed. Additionally, the interactions between environmental factors such as temperature, salinity, and nutrients, and their impact on the abundance and diversity of microalgae within the plastisphere, are explored. Furthermore, the potential ecological risks posed by the proliferation of toxic microalgae within the plastisphere are discussed considering climate change and rising plastic pollution. By highlighting the growing concerns over algal blooms in the plastisphere, this chapter aims to provide a deeper understanding of how plastic pollution may exacerbate HABs in coastal ecosystems and contribute to the broader ecological challenges in marine environments.

5.1 Characteristics of microalgal communities in plastispheres

Overall, the composition of the plastisphere microalgal community and its ambient

community was significantly different (Figure 5-1a-b). Ochrophyta (64.3%, 61.3%), Dinoflagellata (10.3%, 21.0%), and Chlorophyta (18.8%, 11.6%) were the dominant phyla in both the plastisphere and the seawater. However, the abundances of Chlorophyta were significantly higher in the plastisphere, while the abundance of Dinoflagellata was notably lower. (Figure 5-1c-d). Among 1270 ASVs, 423 ASVs showed significant changes in abundance with 170 ASVs (40.2%) significantly enriched in the plastisphere after filtering for an average relative abundance below 0.01% (Figure 5-1e).

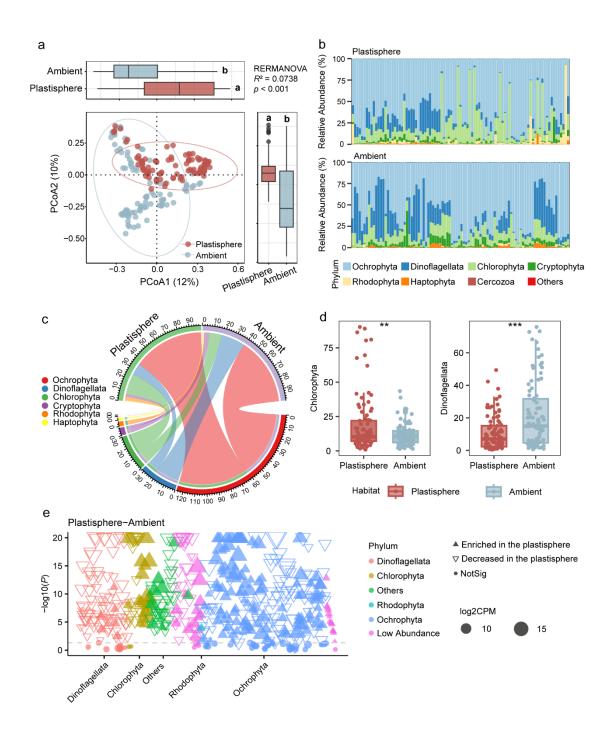


Figure 5-1 Characteristics of the plastisphere microalgal community. (a) Unconstrained principal coordinate analysis (PCoA) with permutational multivariate analysis of variance (PERMANOVA) showing a significant compositional difference between the plastisphere microalgal community and its ambient community. (b) The taxonomic composition at the phylum level of microalgal communities in each site in

the plastisphere and the ambient. (c) The taxonomic proportion of the microalgal community in the plastisphere and its ambient environment. (d) The difference of dominant phylum in the plastisphere and its ambient environment. (e) Manhattan plots showing differences in ASVs between the plastisphere and the ambient. Each circle or triangle represents a single ASV. An upward and filled triangle represents an ASV significantly enriched in the plastisphere, a downward and empty triangle represents an ASV significantly depleted in the plastisphere, while a circle represents an ASV with a nonsignificant difference between the two habitats (*p<0.05; Wilcoxon rank-sum test). CPM, counts per million.

Both richness and Shannon indexes indicated that the plastisphere microalgal community, rather than its ambient community, exhibited a significantly higher level of alpha-diversity (Fig. 1d). Additionally, more unique ASVs were found in the plastisphere and a considerable proportion of ASVs was only detected in the plastisphere (Fig. 1e). These results suggest that the plastisphere could harbor alien microalgae from upstream systems into coastal ecosystems, potentially posing microalgal invasion risks. The plastisphere microalgal community also exhibited a significantly lower beta-diversity, suggesting that the plastisphere provides a shelter with less variations in environmental conditions for microalgae (Fig. 1f). The heightened microbial diversity observed within plastisphere communities cannot be attributed solely to passive accumulation of ambient microorganisms. This enrichment implies that plastic debris actively facilitates the transport and survival of non-indigenous species from upstream environments to new habitats, intensifying concerns over ecological colonization by foreign taxa. Notably, the presence of non-native algal species within plastispheres introduces critical ecological uncertainties. The prolonged

retention and potential release of these algae into nutrient-rich coastal zones may induce localized HABs. Such proliferation events risk destabilizing native trophic networks through toxin-mediated species displacement and oxygen depletion, potentially triggering cascading disruptions across coastal ecosystems—from benthic community collapse to fisheries resource degradation.

Microalgae represent the primary productivity of marine ecosystems and regulate many biogeochemical cycling processes (Sommeria-Klein et al. 2021; De Vargas et al. 2015; Worden et al. 2015; Arrigo 2005). With the ever-increasing emission of plastic waste into the ocean, the density of plastisphere-associated microalgal communities with distinct compositions and higher taxonomic diversity will increase in the ocean. As a result, the overall structure and diversity of marine microalgal communities could be altered, and the associated ecological processes and biodiversity within marine ecosystems could be influenced.

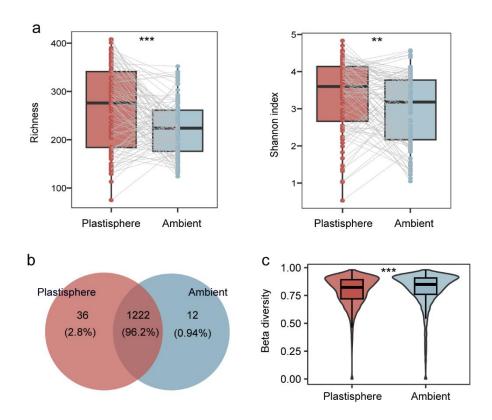


Figure 5-2 Taxonomic diversity in the plastisphere. (a) Differences in the alphadiversity indexed by Richness and Shannon between the plastisphere microalgal community and its ambient community (**p < 0.01, **p < 0.001; Wilcoxon rank-sum test). (b) Overlap ASVs between plastisphere and ambient algal community. (c) Compositional dissimilarity between the plastisphere microalgal community and its ambient community based on Bray-Curtis distance.

5.2 The plastisphere provide a shelter to microalgal community

We then explored assembly mechanisms of the plastisphere microalgal community. The MST model showed that the assembly of both the plastisphere microalgal community and its ambient community was dominated by deterministic processes (MST < 0.5), but stochastic assembly played a more important role in the plastisphere community assembly than in the ambient community assembly (Figure 5-3a). Additionally, the

habitat-level niche breadth was found to be significantly higher in the plastisphere (Figure 5-3b), suggesting that microalgae in the plastisphere might be less affected by environmental filtering (Li et al. 2021c; Wu et al. 2018b).

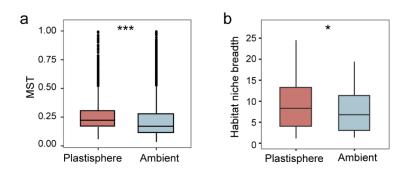


Figure 5-3 Community assembly mechanism of plastisphere microalgal community and its ambient community. (a) The modified stochasticity ratio (MST) model showing the relative importance of the stochastic process in the plastisphere and ambient microalgal community assembly (***p < 0.001; Wilcoxon rank-sum test). (b) Habitat niche breadth in the plastisphere and ambient microalgal community (*p < 0.05; Wilcoxon rank-sum test).

Procrustes analysis demonstrated that a significant correlation existed between the plastisphere microalgal community and its ambient community ($M^2 = 0.746$; p < 0.001; Figure 5-4a), indicating the plastisphere microalgal community was noteworthy shaped by surrounding community composition. The RDA result revealed that all detected physicochemical parameters were significantly associated with the microalgal community in seawater (Figure 5-4b). However, only pH, salinity, temperature, and phosphate concentration showed strong associations with the plastisphere microalgal community composition, while concentrations of nitrate and DOC did not (Figure 5-4c). The detected physicochemical factors explained 10.7% of variations in the

plastisphere microalgal community, much less than the 19.6% explained in the ambient community. VPA was employed to further quantify the contribution of environmental factors and ambient species interactions to plastisphere algal community variation. A combined 25% of the variation in algal community could be explained, with non-toxic algal taxa accounting for 6%, environmental factors for 2%, and their combined effects for 17%; the remaining 75% of variation was unexplained (Figure 5-4d). The unexplained variability in the plastisphere algal community might also be influenced by intricate interactions among viruses, bacteria, and algae. These microbes not only supply crucial nutrients that promote algal growth but may also indirectly influence algae by altering environmental conditions through their physiological changes (Ma et al. 2022; Seymour et al. 2017).

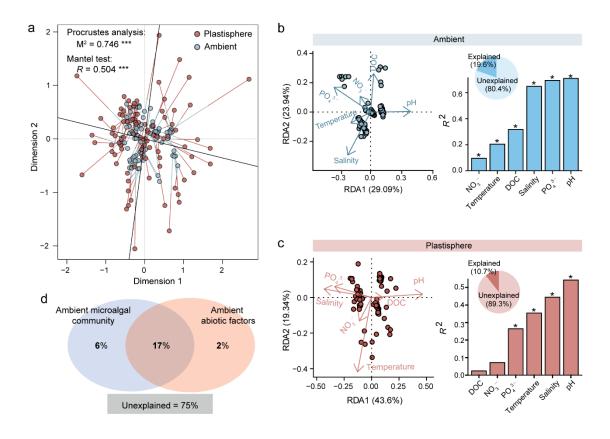


Figure 5-4 Potential driving factors shaping the plastisphere microalgal community and

its ambient community. (a) Significant correlation of microalgal community profile in the plastisphere with the ambient seawater based on Procrustes analysis (***p<0.001). (b) Redundancy analysis between environmental variables and ambient algal communities (*p<0.05). (c) Redundancy analysis between environmental variables and plastisphere algal communities (*p<0.05). (d) Variation partition analysis showing the variations in the plastisphere microalgal community explained by ambient microalgal community and environmental factors.

The phenomenon of reduced environmental filtering effects on microalgae within the plastisphere is potentially due to the formation of biofilms by microalgae in conjunction with other microorganisms on plastic surfaces, which offer protective benefits, enhancing their resilience against adverse environmental conditions (Yan et al. 2024). The relatively stable microenvironments provided in the plastisphere can mitigate fluctuations in factors such as light, temperature, and water flow, facilitating easier adaptation and reproduction of microalgae, thereby further reducing the intensity of environmental filtering (Schaum 2019). Additionally, the adsorption and accumulation of organic matter from surrounding environmental media on plastic surfaces make the plastisphere a nutrient-rich environment (Bowley et al. 2021). This microenvironmental condition diminishes the reliance of microalgae on traditional environmental constraints, enabling them to survive and reproduce in relatively nutrient-poor marine ecosystems. This could explain why the microalgal community in the plastisphere was not affected by nitrate and DOC. The enhanced stochasticity and the reduced environmental filtering within the plastisphere reflect the unique and resource-rich growth environment provided by plastic surfaces. This environment facilitates the

survival and reproduction of microalgae, particularly in nutrient-depleted marine ecosystems, thereby altering the dynamics and ecological functions of microbial communities. These findings provide crucial insights into understanding the impacts of plastic pollution on marine ecosystems and underscore the significance of the plastisphere as a new habitat influencing algal community structure.

5.3 Microalgal hazards harbored by plastispheres

Using our constructed database HTMaDB, a total of 373 ASVs among 1270 ASVs (29%) from our samples were identified as harmful/toxic microalgae. It is reasonable to expect that there are some taxa (2.4% ASVs) observed in the ambient community that were not detected in the plastisphere because some taxa may not have been exposed to or were not adapted to the plastisphere microenvironment. However, notably, 14 (3.8%) HAB ASVs were only detected in the plastisphere and were absent from the ambient community (Figure 5-5a). This result indicates that the plastisphere could harbor harmful/toxic microalgae from upstream environments entering coastal ecosystems, which may potentially trigger harmful algal blooms, thereby disturbing the balance and functioning of coastal ecosystems. The identified harmful/toxic ASVs attribute to 30 toxic and 44 harmful microalgal species (Figure 5-5b), primarily belonging to Diatomea (58.0%), Dinophyceae (28.3%), and Chlorophyceae (6.25%) (Figure 5-5c).

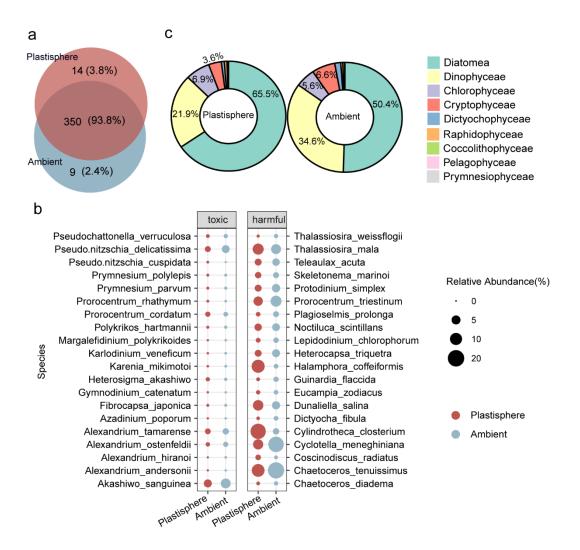


Figure 5-5 Harmful and toxic microalgal taxa. (a) The overlap of ASVs between the plastisphere and the bulk environment. (b) Top 20 relative abundance of harmful and toxic microalgae species in the two habitats. (c) The taxonomic composition of harmful and toxic algal community in the plastisphere and the ambient.

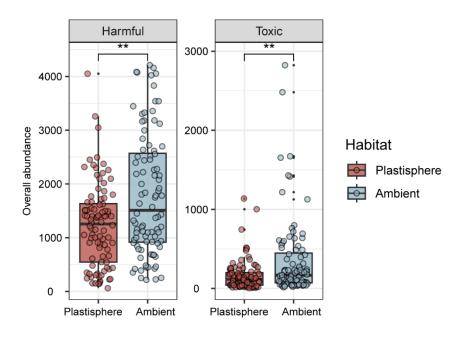


Figure 5-6 Comparison total abundance of HAB taxa in the two habitats.

Although the total abundance of these taxa was lower in the plastisphere (Figure 5-6), many harmful and toxic algal taxa were enriched in the plastisphere compared to the ambient community (Figure 5-7), The most abundant taxa were all diatoms in the plastisphere. Of the top 18 abundant harmful/toxic species, seven were enriched in the plastisphere, namely *Chaetoceros socialis*, *Cylindrotheca closterium*, *Halamphora coffeiformis*, *Nitzschia longissima*, *Skeletonema marinoi*, *Pseudo-nitzschia cuspidate* and *Thalassiosira allenii*. Notably, all these plastisphere-enriched harmful/toxic microalgae belong to diatoms. Most studies have reported that diatoms are common and primary residents of the plastisphere (Nava & Leoni 2021). A similar finding was observed for marine plastic debris collected from both pelagic and benthic environments along the Mediterranean coastal waters of Greece, Italy, and Spain, reporting that diatoms were present on nearly 100% of the plastic debris samples (Masó et al. 2016).In general, diatoms could be classified as centric (radial) or pennate

(bilateral) according to their valve symmetry (Williams & Kociolek 2011). Pennate diatoms including *C. Closterium*, *H. coffeiiformis*, *N. longissimi* and *P. cuspidate* in our study were found to be abundant in both benthic and planktonic environment and capable of adhesion and motility on natural or artificial substrata. In contrast, centric diatoms are predominantly planktonic and are rarely observed in sediments or attached to various surfaces. However, a few centric diatom taxa, such as *C. socialis*, *S. marinoi*, and *T. allenii*, were found to be enriched within the plastisphere, suggesting that they can adapt to specialized habitats, such as the microenvironment provided by the plastisphere.

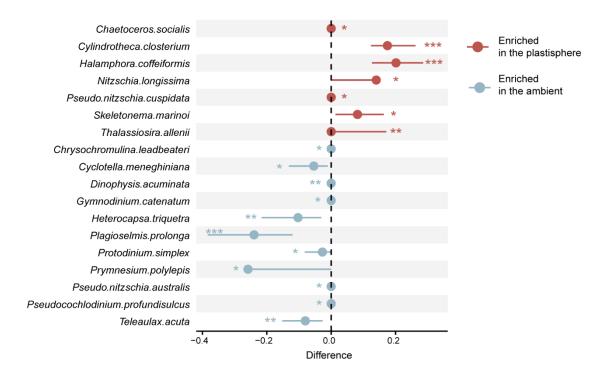


Figure 5-7 The enrichment of specific algal species in the plastisphere (*p < 0.05, **p < 0.01, ***p < 0.001; Wilcoxon rank-sum test).

The successful colonization of diatoms in the plastisphere is owing to their exceptional adhesion capabilities (Casabianca et al. 2019). Diatoms possess specialized structures, including a unique surface roughness from their nanoporous silica, which facilitates their attachment to various substrates, including plastic surfaces (Khan et al. 2020). The secretion of extracellular polymeric substances (EPS) by members of diatoms such as *Chaetoceros* and *Thalassiosira* during growth plays a crucial role in this process, creating a sticky matrix that allows diatoms to firmly anchor to the plastic (Vidal-Melgosa et al. 2021). This strong adhesion not only enables diatoms to resist detachment from the substrate but also promotes the formation of biofilms (Wright et al. 2020). Plastispheres are typically nutrient-rich environments, since they can sorb organic matter from the surrounding environment and leach organics from their interior (Rillig et al. 2024b; Sheridan et al. 2022), providing nutrient sources for diatoms (Christie-Oleza et al. 2017).

Diatoms typically exhibit rapid growth and reproduction rates, enabling them to occupy available ecological niches quickly (Inomura et al.; Taurozzi et al. 2024). This competitive advantage facilitates the establishment of dense populations on plastic surfaces. In contrast, dinoflagellates have slower growth rates and longer generation times, which may hinder their ability to compete effectively for resources in the dynamic environment of the plastisphere. The ability of diatoms to rapidly respond to environmental changes, such as fluctuations in light availability and nutrient concentration, enhances their competitiveness in these novel niches (Zhou et al. 2021).

Moreover, the interactions between diatoms and other microorganisms present in the plastisphere may further contribute their adaptation in the plastisphere. The ability of

diatoms to form complex biofilms with bacteria enhances their nutrient acquisition and provides protection against predation (Flemming & Wingender 2010; Sun et al. 2023). These biofilms create a microenvironment that can stabilize nutrient availability, thereby supporting the growth of diatoms.

Together, the enhanced ability of diatoms to colonize the plastisphere can be attributed to their superior adhesion capabilities, rapid growth rates, efficient nutrient utilization, broad ecological adaptability, and beneficial interactions with other microorganisms. As plastic debris continues to accumulate in marine environments, the proliferation of diatoms in the plastisphere may have far-reaching implications for nutrient cycling, food web dynamics, and ecosystem health.

Coastal plastispheres enriched with harmful/toxic microalgae pose critical threats to biological health. For example, the plastisphere-enriched species *Pseudo-nitzschia cuspidate* is a producer of domoic acid (DA), a kind of neurotoxin that can cause animal seizures, disorientation, and even death (Cook et al. 2015). The widespread and long-distance transport of plastic debris facilitates the proliferation of toxic algal blooms. A High concentration of DA exposure have been linked to increased stranding events of California sea lions, with substantial impacts on their population dynamics (Scholin et al. 2000).

Beyond the enrichment of diatom species in the plastisphere, ichthyotoxic dinoflagellates such as *Margalefidinium polykrikoides* and *Prymnesium parvum* were also detected (Fig. S4) (Hofmann et al. 2021; Wagstaff et al. 2021). The presence of *Alexandrium* spp., *Dinophysis* spp., *Prorocentrum* spp., *Azadinium* spp., and *Karenia* spp. pose a threat to various shellfish poisoning (Lassus et al. 2016; Li et al. 2024d;

Marzidovšek et al. 2024; Valbi et al. 2019). A previous study has reported that the cultured *A. pacificum* strains isolated from plastic debris could produce paralytic shellfish toxins (Casabianca et al. 2019). The incidence of shellfish poisoning due to the consumption of contaminated seafood warrants attention, as it poses direct health risks to humans and further complicates the ecological ramifications of plastic pollution in marine ecosystems (Trainer et al. 2012). Additionally, many species of dinoflagellates such as *Alexandrium* spp. are able to form resting cysts (Dai et al. 2020), which are particularly relevant to survival and transport of these species on plastic debris. These interconnected issues underscore the urgency of addressing plastic debris and its role in the proliferation of harmful algal blooms, as well as the consequent effects on marine life and human health.

Monitoring data reveal that the frequency and intensity of harmful algal bloom events are increasing with human activities in the Anthropocene (Dai et al. 2023b). According to our findings, it is reasonable to hypothesize that the ever-increasing emission of plastic waste and its long-distance, cross-ecosystem transportable property may be an overlooked contributor to the increasing harmful algal bloom outbreaks. Further monitoring and more modeling evidence are needed to test this hypothesis. If confirmed, controlling plastic pollution will become a crucial strategy for managing and preventing algal bloom outbreaks in coastal ecosystems.

5.4 Risks associated with toxic and harmful communities under the background of climate change

We performed the Mantel test to investigate the major driving environmental drivers of

the harmful and toxic algal community in the plastisphere. The result showed that harmful and toxic algal communities in both the plastisphere and its ambient seawater were significantly associated with temperature, pH, and salinity, with temperature showing the largest correlation coefficient (Figure 5-8a). Linear regression models were further applied to uncover how the total abundance of harmful and toxic microalgae responds to environmental factors (Figure 5-8b). The result indicated that the total abundance of harmful and toxic microalgae in both the plastisphere and its ambient was positively associated with temperature, while it was negatively associated with pH levels. In the future, the interwoven challenges of global changes, including climate warming leading to a warmer ocean, increasing CO₂ levels resulting in a more acidic ocean, and the ever-increasing trajectory of plastic emissions, may exacerbate the microalgal risk in coastal ecosystems (Simmer et al. 2023). Although current results provide valuable preliminary insights into the microalgal-associated risks with plastics, their application in risk assessment remains limited. Beyond species identification, future investigations should incorporate functional assessments, including toxin production, pathogenicity, and microbial viability analyses, to better understand the actual risk potential of these organisms. Ultimately, addressing the complex challenges posed by microplastic pollution and the associated biological risks will require interdisciplinary collaboration across microbiology, ecology, oceanography, toxicology, and public health sectors. Only through such concerted efforts can we better quantify risks and design targeted interventions to mitigate the impacts of plastic pollution on marine ecosystems.

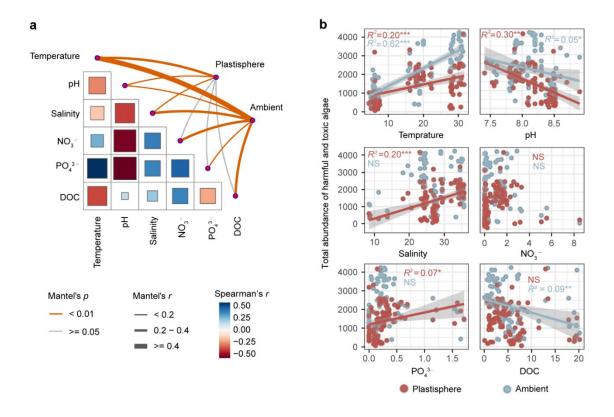


Figure 5-8 Associations between environmental factors and the microalgal risk. (a) Correlations of the plastisphere and ambient algal community with environmental factors using Mantel test. (b) Linear regression analysis between the driving environmental factors and total abundance of harmful and toxic microalgae in the plastisphere and ambient seawater. Environmental factors showing significant associations with harmful and toxic microalgal community structure in the Mantel test were all included in the linear regression models (*p < 0.05, **p < 0.01, ***p < 0.001, NS = not significant).

5.5 Summary

This chapter investigated the composition, diversity, and ecological risks associated with microalgae in the coastal plastisphere, providing new insights into the potential impacts of plastic pollution on marine ecosystems. The study focused on the

differentiation of microalgal communities between the plastisphere and ambient seawater, the environmental drivers influencing these communities, and the enrichment of harmful and toxic taxa in the plastisphere. Key findings are summarized below:

The microalgal community composition in the plastisphere was significantly different from that in ambient seawater. Ochrophyta, Dinoflagellata, and Chlorophyta were the dominant phyla in both environments; however, Chlorophyta were significantly enriched in the plastisphere, whereas Dinoflagellata were notably less abundant. Of the 1,270 ASVs identified, 423 showed significant abundance changes, with 170 ASVs enriched in the plastisphere. Both richness and Shannon indexes indicated higher alphadiversity in the plastisphere, which also harbored unique ASVs not present in ambient seawater. These findings suggest that the plastisphere serves as a stable and nutrientrich microhabitat capable of harboring alien microalgae, potentially introducing invasion risks into coastal ecosystems. The assembly of plastisphere microalgal communities was shaped by both deterministic and stochastic processes, with stochasticity playing a larger role compared to ambient seawater. The broader habitatlevel niche breadth observed in the plastisphere indicated reduced environmental filtering effects, likely due to the stable microenvironment and biofilm formation on plastic surfaces. Procrustes analysis revealed a strong correlation between plastisphere and ambient microalgal communities, while RDA showed that pH, salinity, temperature, and phosphate concentrations were key factors influencing plastisphere community composition. However, the overall contribution of environmental factors to plastisphere community variation was much lower than that for ambient seawater, highlighting the unique ecological dynamics of the plastisphere.

Using the HTMaDB database, 373 harmful and toxic ASVs were identified, including 14 taxa exclusively detected in the plastisphere. Most harmful and toxic taxa belonged to diatoms, such as Pseudo-nitzschia cuspidate, Chaetoceros socialis, and Skeletonema marinoi. These species are known producers of toxins like DA, posing significant risks to marine ecosystems and human health. The enhanced colonization of diatoms in the plastisphere was attributed to their strong adhesion capabilities, rapid growth rates, and ability to form biofilms. In addition to diatoms, harmful dinoflagellates were also detected, raising concerns about the plastisphere's role in exacerbating HABs. The plastisphere's ability to enrich harmful and toxic microalgae poses significant ecological and health risks. The proliferation of toxin-producing algae in the plastisphere can disrupt nutrient cycling, food web dynamics, and ecosystem functions, potentially exacerbating the frequency and intensity of HABs. The presence of harmful algae, such as Pseudo-nitzschia cuspidate, in the plastisphere highlights its role in facilitating long-distance transport and proliferation of toxic algal blooms, which have been linked to mass mortality events in marine life and human shellfish poisoning incidents.

The Mantel test identified temperature, pH, and salinity as major environmental drivers of harmful and toxic algal communities in the plastisphere, with temperature showing the strongest correlation. Linear regression models further revealed that the abundance of harmful and toxic algae was positively associated with temperature and negatively associated with pH. These findings suggest that ongoing global changes, including ocean warming, acidification, and increasing plastic pollution, could amplify the risks

posed by harmful algae in coastal ecosystems.

In summary, this chapter revealed significant compositional differences and higher diversity in plastisphere microalgal communities compared to ambient seawater. The plastisphere serves as a unique and nutrient-rich microhabitat that facilitates the survival, enrichment, and transport of harmful and toxic algae. The findings underscore the critical ecological risks posed by the plastisphere, particularly in exacerbating HABs and associated toxin production. The reduced environmental filtering effects and enhanced stochasticity within the plastisphere highlight its role as a novel ecological niche influencing marine microalgal dynamics. However, a large proportion of the observed variation in plastisphere algal communities remains unexplained, likely due to intricate interactions among algae, bacteria, and viruses. Future research should focus on these complex interactions and their implications for ecosystem health. Additionally, the role of the plastisphere in amplifying the impacts of climate change and plastic pollution on marine ecosystems warrants further investigation. Continued monitoring and management efforts are essential to mitigate the ecological and public health risks posed by the plastisphere and its associated harmful microalgae.

Chapter 6 Conclusions and Recommendations

6.1 Overall summary and major conclusions

This thesis undertook a comprehensive examination of the ecological patterns and driving mechanisms of microalgal community and associated risks in coastal ecosystems with our curated database. The research focused on understanding the dynamics and underlying mechanism of HAB species, as well as their colonization characteristic within the plastisphere-the unique ecological niche formed on the surfaces of plastic debris. This niche not only serves as a habitat for diverse microalgal communities but also significantly alters their composition and diversity. The main findings of the thesis are as follows:

- Construction of a comprehensive 18S full length database: The utilization of the
 HTMaDB has significantly enhanced the capacity for accurately identifying and
 monitoring harmful and toxic microalgae. This tool is instrumental in assessing the
 diversity and distribution of algal species in affected marine regions, providing a
 crucial resource for researchers and policymakers aiming to mitigate the impacts of
 HABs.
- 2. Ecological impact of the plastisphere: The studies highlighted the plastisphere's role as a distinct ecological niche that selectively enriches and alters the composition of microalgal communities. Microalgal communities in the plastisphere were less constrained by ambient physicochemical factors, enabling the shelter role of the plastisphere for microalgae. Compared to its ambient community, the microalgal community in the plastisphere harbored a higher diversity of algal species,

including prevelance of harmful and toxic diatoms (e.g., *Pseudo-nitzschia cuspidate*, *Chaetoceros socialis*, and *Skeletonema marinoi*). The plastisphere could facilitate the proliferation of these species by providing a stable habitat that is somewhat insulated from environmental stressors typically present in the surrounding marine environment.

- 3. Influence of environmental factors: Temperature and salinity are identified as critical environmental factors that influence algal community dynamics. The findings underscore that these abiotic factors, especially temperature, drive the seasonal patterns of algal communities and enhance the prevalence of HABs during certain times of the year. This relationship suggests that changes in these environmental variables due to climate change could potentially exacerbate the occurrence and severity of HABs.
- 4. Potential for increased HAB occurrences: By providing a refuge and a means of transport, plastic debris may increase the frequency and geographic spread of HABs. The ability of the plastisphere to act as a vector for harmful algae and their toxins introduces new dynamics into marine ecosystems, potentially facilitating the spread of these organisms beyond their natural geographic boundaries and leading to new outbreaks of HABs in previously unaffected areas.

The findings highlight the need for a paradigm shift in managing harmful algal blooms (HABs) – moving from reactive suppression strategies to proactive measures

addressing plastic pollution. One approach is to designate high-risk regions, such as the Pearl River estuaries, as "HAB-plastic control zones," where stringent regulations on plastic use and disposal would be enforced. Additionally, developing chitosan-based coatings to prevent diatom adhesion without contributing to microplastic pollution represents a promising solution. This work provides a scientific foundation for incorporating plastic pollution management as an integral part of HAB mitigation efforts, which is essential for advancing toward Sustainable Development Goal 14 (Life Below Water) in the face of escalating global environmental challenges.

6.2 Limitations of the current study and future perspectives

Although the research study provided significant contributions to the field of marine ecology by establishing a comprehensive database of harmful and toxic microalgae, which enhance our ability to monitor and manage HABs effectively. It also highlighted the emerging environmental challenge posed by plastic debris in marine ecosystems. However, there are limitations that must be acknowledged to refine future research and interventions. These limitations, coupled with suggested future perspectives, can help guide subsequent studies aiming to further unravel the complexities of HAB dynamics and the plastisphere.

The current research is primarily focused on surface seawater, neglecting the investigation of harmful and toxic algae in the benthic layers. This oversight can lead to an incomplete understanding of the spatial distribution and full ecological impacts of HABs, as different environmental conditions and algal communities may exist in deeper

waters. Sampling was conducted only during the dry and wet seasons, which does not allow for continuous, long-term monitoring of phytoplankton dynamics. This limitation restricts the ability to capture seasonal variations fully and understand long-term trends and fluctuations in phytoplankton communities, which are crucial for predicting and managing algal blooms effectively. In the case of VPA, only interactions among algal species were considered. This approach overlooks the potential influence of other biological factors such as bacteria and viruses, which can significantly impact algal growth and community dynamics through various interactions, including symbiosis, competition, and predation. The current work mainly relies on culture-independent approaches, such as high-throughput qPCR and 18SrRNA gene amplicon sequencing. Correlation analysis was conducted between the microbial community and environmental factors. It will be more important to set up lab experiments to validate if there is causality, rather than correlation. The study did not include measurements of algal toxins and quantification of microalgae-associated risk within the plastisphere. This is a significant oversight, as plastics can absorb and concentrate toxins produced by harmful algae. Without assessing the presence and concentration of toxins on plastic debris, it is difficult to fully understand the role that the plastisphere plays in the distribution and potential enhancement of algal toxin risks in marine environments. This limitation hinders the ability to evaluate the health risks posed by toxin-laden plastics to marine life and, by extension, to human health through bioaccumulation and the food chain.

To overcome these limitations, future studies should aim to include deeper water layers in their sampling regimen to provide a more holistic view of the algal communities across different marine strata. Additionally, extending the monitoring periods to encompass multiple seasons or even continuous year-round sampling could offer a more detailed picture of phytoplankton dynamics and their responses to environmental changes. Including a broader range of biological interactions in the analysis, particularly with non-algal microorganisms like bacteria and viruses, would also enrich the understanding of community dynamics and ecological balances. It might be relevant to isolate representative harmful and toxic taxa from seawater samples by employing culture-dependent methods to further consolidate the conclusions. Finally, quantifying the impacts of plastic pollution on marine microalgal communities and bloom events through simulation experiments and big-data modelling could yield more precise data on how different plastics affect marine life, thereby facilitating more targeted mitigation strategies against plastic pollution.

Appendix 1

This section provides detailed database information in this study.

1 Source of eukaryotic toxic and harmful microalgae species.

Public website IOC-UNESCO Taxonomic Reference List of Harmful Micro Algae, Lundholm, N.; Churro, C.; Escalera, L.; Fraga, S.; Hoppenrath, M.; Larsen, J.; Mertens, K.; Moestrup, Ø.; Murray, S.: Tillmann, U.; Zingone, A. (Eds) (2009 onwards). IOC-UNESCO Taxonomic Reference List of Harmful Micro Algae. Accessed at https://www.marinespecies.org/hab on 2024-10-26. doi:10.14284/362 Paper search Number **Author Full Names Article Title Source Title** Roselli, Leonilde; Caroppo, Carmela; Bevilacqua, Stanislao; Harmful algae and pressure-impact relationship: Noxious blooms MARINE ENVIRONMENTAL Ciciriello, Pierangelo Cosimo; Ungaro, Nicola; Vadrucci, and toxic microalgae occurrence from coastal waters of the Apulia RESEARCH Maria Rosaria region (Adriatic and Ionian Seas, Mediterranean) Hargraves, PE; Maranda, L Potentially toxic or harmful microalgae from the northeast coast NORTHEASTERN NATURALIST Hernandez-Becerril, David U.; Alonso-Rodriguez, Rosalba; Toxic and harmful marine phytoplankton and microalgae (habs) in

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| 148 | Antonopoulou, Maria; Vlastos, Dimitris; Dormousoglou, Margarita; Bouras, Spyridon; Varela-Athanasatou, Maria; Bekakou, Irene-Eleni | Genotoxic and Toxic Effects of The Flame Retardant Tris(Chloropropyl) Phosphate (TCPP) in Human Lymphocytes, Microalgae and Bacteria | TOXICS |
| 149 | Davarpanah, Elham; Guilhermino, Lucia | Are gold nanoparticles and microplastics mixtures more toxic to the marine microalgae Tetraselmis chuii than the substances individually? | ECOTOXICOLOGY AND ENVIRONMENTAL SAFETY |

| | Zingone, Adriana; Escalera, Laura; Aligizaki, Katerina; | Toxic marine microalgae and noxious blooms in the Mediterranean | HARMFUL ALGAE |
|------|--|---|----------------------------------|
| | Fernandez-Tejedor, Margarita; Ismael, Amany; Montresor, | Sea: A contribution to the Global HAB Status Report | |
| 150 | Marina; Mozetic, Patricija; Tas, Seyfettin; Totti, Cecilia | • | |
| | Muthuraman, R. M.; Murugappan, A.; Soundharajan, B. | Highly effective removal of presence of toxic metal concentrations | APPLIED NANOSCIENCE |
| 151 | | in the wastewater using microalgae and pre-treatment processing | |
| | Loukas, Christos-Moritz; McQuillan, Jonathan S.; | Detection and quantification of the toxic microalgae Karenia brevis | JOURNAL OF MICROBIOLOGICAL |
| 1.50 | Laouenan, Florian; Tsaloglou, Maria-Nefeli; Ruano-Lopez, | using lab on a chip mrna sequence-based amplification | METHODS |
| 152 | Jesus M.; Mowlem, Matthew C. | | CHENGERY A FUROREAN |
| | Hakamada, Mayu; Tokairin, Chihiro; Ishizuka, Hayate; Adachi, Kanna; Osawa, Toma; Aonuma, Shiori; Hirozumi, | Synthesis and Identification of decarbamoyloxysaxitoxins in Toxic Microalgae and their Reactions with the Oxygenase, sxtt, Reveal | CHEMISTRY-A EUROPEAN JOURNAL |
| | Ryosuke; Tsuchiya, Shigeki; Cho, Yuko; Kudo, Yuta; | Saxitoxin Biosynthesis | JOURNAL |
| | Konoki, Keiichi; Oshima, Yasukatsu; Nagasawa, Kazuo; | Saxitoxiii Biosylidiesis | |
| 153 | Yotsu-Yamashita, Mari | | |
| | Zhang, Weixia; Sun, Shuge; Du, Xueying; Han, Yu; Tang, | Toxic impacts of microplastics and tetrabromobisphenol A on the | GONDWANA RESEARCH |
| 154 | Yu; Zhou, Weishang; Shi, Wei; Liu, Guangxu | motility of marine microalgae and potential mechanisms of action | |
| | Sun, Ying-ying; Wang, Hui; Guo, Gan-lin; Pu, Yin-fang; | Isolation, purification, and identification of antialgal substances in | ENVIRONMENTAL SCIENCE AND |
| | Yan, Bin-lun; Wang, Chang-hai | green alga Ulva prolifera for antialgal activity against the common | POLLUTION RESEARCH |
| 155 | | harmful red tide microalgae | |
| 156 | Penna, Antonella; Galluzzi, Luca | The quantitative real-time PCR applications in the monitoring of | ENVIRONMENTAL SCIENCE AND |
| 156 | CI WILLIAMI, II. W W. D. I | marine harmful algal bloom (HAB) species (vol 20, pg 6851, 2013) | POLLUTION RESEARCH |
| | Cho, Kichul; Ueno, Mikinori; Liang, Yan; Kim, Daekyung; Oda, Tatsuya | Generation of Reactive Oxygen Species (ROS) by Harmful Algal Bloom (HAB)-Forming Phytoplankton and Their Potential Impact | ANTIOXIDANTS |
| 157 | Oda, Taisuya | on Surrounding Living Organisms | |
| 137 | Chen, Junhui; Wang, Yanlong; Pan, Lei; Shen, Huihui; Fu, | Separation and purification of two minor typical diarrhetic shellfish | JOURNAL OF SEPARATION |
| | Dan; Fu, Boqiang; Sun, Chengjun; Zheng, Li | poisoning toxins from harmful marine microalgae via combined | SCIENCE |
| 158 | Dan, I a, Doquaig, San, Changjan, Enang, Er | liquid chromatography with mass spectrometric detection | Selbiveb |
| | Herrera-Sepulveda, Angelica; Hernandez-Saavedra, Norma | Capillary electrophoresis finger print technique (CE-SSCP): an | ENVIRONMENTAL SCIENCE AND |
| | Y.; Medlin, Linda K.; West, Nyree | alternative tool for the monitoring activities of HAB species in Baja | POLLUTION RESEARCH |
| 159 | | California Sur Costal | |
| | Bazzoni, Anna Maria; Caddeo, Tiziana; Pulina, Silvia; | Spatial distribution and multiannual trends of potentially toxic | ENVIRONMENTAL MONITORING |
| 1.60 | Padedda, Bachisio M.; Satta, Cecilia T.; Sechi, Nicola; | microalgae in shellfish farms along the Sardinian coast (NW | AND ASSESSMENT |
| 160 | Luglie, Antonella | Mediterranean Sea) | GOVERNOR OF THE TOTAL |
| | Wang, Lei; Huang, Xulei; Lim, Dorothy Jingwen; Laserna, Anna Karen Carrasco; Li, Sam Fong Yau | Uptake and toxic effects of triphenyl phosphate on freshwater microalgae Chlorella vulgaris and Scenedesmus obliquus: Insights | SCIENCE OF THE TOTAL ENVIRONMENT |
| 161 | Aima Kaich Carrasco; Li, Sam Fong Tau | from untargeted metabolomics | ENVIRONWENT |
| 101 | Petsas, A. S.; Vagi, M. C.; Pavlaki, M. D.; Smaragdaki, N. | Toxic effects of azinphos methyl and azinphos ethyl on growth and | PROCEEDINGS OF THE 13TH |
| | M.; Kostopoulou, M. N.; Lekkas, T. D. | chlorophyll pigments production of marine unicellular microalgae | INTERNATIONAL CONFERENCE |
| | 1.1., 120000poulou, 11.111, Dollino, 11.D. | tetraselmis suecica | ON ENVIRONMENTAL SCIENCE |
| 162 | | | AND TECHNOLOGY |

| | Thiagarajan, Vignesh; Ramasubbu, Seenivasan | Extraction and characterization of sub-micron sized tio2 from | TOXICOLOGY AND |
|-----|--|---|--|
| 163 | Tinagarajan, vigitesii, Ramasuoou, Seenivasan | toothpaste: evaluation of their toxic effects in marine microalgae Chlorella sp. | ENVIRONMENTAL HEALTH SCIENCES |
| 164 | Chen, Daijie; Wang, Guobao; Chen, Chiyu; Feng, Zekai; Jiang, Yuanyuan; Yu, Hang; Li, Mengyao; Chao, Yuanqing; Tang, Yetao; Wang, Shizhong; Qiu, Rongliang | The interplay between microalgae and toxic metal(loid)s: mechanisms and implications in AMD phycoremediation coupled with Fe/Mn mineralization | JOURNAL OF HAZARDOUS MATERIALS |
| 165 | Wang, Jiayin; Tan, Liju; Ni, Ziqi; Zhang, Na; Li, Qi; Wang, Jiangtao | Is hydrodynamic diameter the decisive factor?-Comparison of the toxic mechanism of nsio2 and mps on marine microalgae Heterosigma akashiwo | AQUATIC TOXICOLOGY |
| 166 | Doblin, Martina A.; Coyne, Kathryn J.; Rinta-Kanto, Johanna M.; Wilhelm, Steven W.; Dobbs, Fred C. | Dynamics and short-term survival of toxic cyanobacteria species in ballast water from NOBOB vessels transiting the Great Lakes - implications for HAB invasions | HARMFUL ALGAE |
| 167 | Norén, F; Moestrup, O; Rehnstam-Holm, AS | Parvilucifera infectans Noren et Moestrup gen. Et sp nov (Perkinsozoa phylum nov.):: a parasitic flagellate capable of killing toxic microalgae | EUROPEAN JOURNAL OF PROTISTOLOGY |
| 168 | Wodeyar K, K. Abhilash; Akter, Sahina; Nama, Suman; Nayak, Binaya Bhusan; Gogoi, Pranab; Deshmukhe, Geetanjali; Jaiswar, Ashok Kumar; Landge, Asha Taterao; Layana, Porayil | Unveiling the spatio-temporal variation of harmful algal bloom (HAB) species assemblages and their relationship with environmental factors in a tidal creek environment | AQUATIC SCIENCES |
| 169 | Wang, Hao; Bouwman, Alexander F.; Wang, Junjie; Yu, Zhigang; Ran, Xiangbin | Competitive advantages of HAB species under changing environmental conditions in the coastal waters of the Bohai Sea, Yellow Sea and East China Sea | CONTINENTAL SHELF RESEARCH |
| 170 | Loukil-Baklouti, Amira; Feki-Sahnoun, Wafa; Hamza, Asma; Abdennadher, Moufida; Mahfoudhi, Mabrouka; Bouain, Abderrahmen; Jarboui, Othman | Controlling factors of harmful microalgae distribution in water column, biofilm and sediment in shellfish production area (South of Sfax, Gulf of Gabes) from southern Tunisia | CONTINENTAL SHELF RESEARCH |
| 171 | Ohkubo, Nobuyuki; Tomaru, Yuji; Yamaguchi, Haruo; Kitatsuji, Saho; Mochida, Kazuhiko | Development of a method to assess the ichthyotoxicity of the harmful marine microalgae Karenia spp. Using gill cell cultures from red sea bream (Pagrus major) | FISH PHYSIOLOGY AND BIOCHEMISTRY |
| 172 | Das, Soupam; Chandrasekaran, N.; Mukherjee, Amitava | Unmasking effects of masks: Microplastics released from disposable surgical face masks induce toxic effects in microalgae Scenedesmus obliquus and Chlorella sp. | COMPARATIVE BIOCHEMISTRY AND PHYSIOLOGY C- TOXICOLOGY & PHARMACOLOGY |
| 173 | Ebenezer, Vinitha; Ki, Jang-Seu | Quantification of toxic effects of the herbicide metolachlor on marine microalgae Ditylum brightwellii (Bacillariophyceae), Prorocentrum minimum (Dinophyceae), and Tetraselmis suecica (Chlorophyceae) | JOURNAL OF MICROBIOLOGY |
| 174 | Li, Ming; Wan, Chengyan; Pan, Xiaojie; Zou, Yi; Chang, Jianbo; Xie, Ping | Acute toxic effects of zinc, cadmium, and mercury on the growths of three unicellular green microalgae with relatively high initial densities | FRESENIUS ENVIRONMENTAL BULLETIN |

| 175 | Wang, Zikang; Yu, Simin; Nie, Yufan; Zhang, Ning; Zhu, Wentao; Zhou, Zhiqiang; Diao, Jinling | Interspecific insights into direct and indirect effects of acetochlor on interactions between daphnids and microalgae: Toxic, trophic, and grazer-infochemical responses | SCIENCE OF THE TOTAL ENVIRONMENT |
|-----|--|---|---|
| 176 | Sanchez-Parra, Elisabet; Boutarfa, Soumia; Aboal, Marina | Are Cyanotoxins the Only Toxic Compound Potentially Present in Microalgae Supplements? Results from a Study of Ecological and Non-Ecological Products | TOXINS |
| 177 | Zhang, Chun Yun; Chen, Guo Fu; Cai, Pan Pan; Wang, Yuan Yuan; Guo, Chang Lu | Development and evaluation of a reverse dot blot assay for the simultaneous detection of common toxic microalgae along the Chinese coast | HARMFUL ALGAE |
| 178 | Moreira-Gonzalez, Angel Ramon; Brustolin, Marco Colossi; Mafra Junior, Luiz Laureno | Composition and abundance of benthic microalgae from the Estuarine Complex of Paranagua Bay (southern Brazil) with special emphasis on toxic species | OCEAN AND COASTAL RESEARCH |
| 179 | Orlova, Tatiana Yu. | Resting stages of planktonic microalgae in recent marine sediments from the east coast of Russia with special emphasis on toxic species | PROCEEDINGS OF THE RUSSIA- CHINA BILATERAL SYMPOSIUM ON MARINE ECOSYSTEMS UNDER THE GLOBAL CHANGE IN THE NORTHWESTERN PACIFIC |
| 180 | Oloketuyi, Sandra; Mazzega, Elisa; Zavasnik, Janez; Pungjunun, Kingkan; Kalcher, Kurt; de Marco, Ario; Mehmeti, Eda | Electrochemical immunosensor functionalized with nanobodies for the detection of the toxic microalgae Alexandrium minutum using glassy carbon electrode modified with gold nanoparticles | BIOSENSORS & BIOELECTRONICS |
| 181 | Moreira-Gonzalez, Angel R.; Domit, Camila; Rosa, Kaianan M. S.; Mafra Jr, Luiz L. | Occurrence of potentially toxic microalgae and diarrhetic shellfish toxins in the digestive tracts of green sea turtles (Chelonia mydas) from southern Brazil | HARMFUL ALGAE |
| 182 | Battocchi, Cecilia; Totti, Cecilia; Vila, Magda; Maso, Mercedes; Capellacci, Samuela; Accoroni, Stefano; Rene, Albert; Scardi, Michele; Penna, Antonella | Monitoring toxic microalgae Ostreopsis (dinoflagellate) species in coastal waters of the Mediterranean Sea using molecular PCR-based assay combined with light microscopy | MARINE POLLUTION BULLETIN |
| 183 | Toldra, Anna; Jauset-Rubio, Miriam; Andree, Karl B.; Fernandez-Tejedor, Margarita; Diogene, Jorge; Katakis, Ioanis; O'Sullivan, Ciara K.; Campas, Monica | Detection and quantification of the toxic marine microalgae Karlodinium veneficum and Karlodinium armiger using recombinase polymerase amplification and enzyme-linked oligonucleotide assay | ANALYTICA CHIMICA ACTA |
| 184 | Hu, Zhangxi; Xu, Ning; Gu, Haifeng; Chai, Zhaoyang; Takahashi, Kazuya; Li, Zhun; Deng, Yunyan; Iwataki, Mitsunori; Matsuoka, Kazumi; Tang, Ying Zhong | Morpho-molecular description of a new HAB species, Pseudocochlodinium profundisulcus gen. Et sp. Nov., and its LSU rrna gene based genetic diversity and geographical distribution | HARMFUL ALGAE |
| 185 | Shimada, Hiroshi; Sakamoto, Setsuko; Yamaguchi, Mineo; Imai, Ichiro | First record of two warm-water HAB species Chattonella marina (Raphidophyceae) and Cochlodinium polykrikoides (Dinophyceae) on the west coast of Hokkaido, northern Japan in summer 2014 | REGIONAL STUDIES IN MARINE SCIENCE |
| 186 | Beyer, Valentin P.; Blockx, Jonas; Maciel, Ayessa Pires; Tyagi, Sakshi; Demir-Yilmaz, Irem; Formosa-Dague, Cecile; Muylaert, Koenraad; Thielemans, Wim | Use of PDMAEMA as a non-toxic synthetic flocculant for harvesting microalgae and evaluation of the influence of polymer size on flocculation behaviour | JOURNAL OF ENVIRONMENTAL CHEMICAL ENGINEERING |

| 107 | Bouquet, Aurelien; Perdrau, Marie Anais; Laabir, Mohamed; Foucault, Elodie; Chomerat, Nicolas; Rolland, Jean Luc; | Liza ramada Juveniles after Exposure to the Toxic Dinoflagellate Vulcanodinium rugosum: Effects on Fish Viability, Tissue | TOXINS |
|------------|--|---|--|
| 187 | Abadie, Eric Zhang, YH; Fu, FX; Whereat, E; Coyne, KJ; Hutchins, DA | Contamination and Microalgae Survival after Gut Passage Bottom-up controls on a mixed-species HAB assemblage:: A comparison of sympatric Chattonella subsalsa and Heterosigma akashiwo (Raphidophyceae) isolates from the Delaware Inland Bays, USA | HARMFUL ALGAE |
| 189 | Machado, Rodrigo R.; Lourenco, Sergio O. | Nutritional properties of microalgae used as food for bivalve molluscs: a review | APLICACOES DA FICOLOGIA: ANAIS DO XI CONGRESSO BRASILEIRO DE FICOLOGIA E SIMPOSIO LATINO-AMERICANO SOBRE ALGAS NOCIVAS |
| 190 | Draredja, Mohamed Anis; Frihi, Hocine; Boualleg, Chahinaise; Gofart, Anne; Abadie, Eric; Laabir, Mohamed | Seasonal variations of phytoplankton community in relation to environmental factors in a protected meso-oligotrophic southern Mediterranean marine ecosystem (Mellah lagoon, Algeria) with an emphasis of HAB species | ENVIRONMENTAL MONITORING AND ASSESSMENT |
| 191 | Garcia-Balboa, C.; Baselga-Cervera, B.; Garcia-Sanchez, A.; Igual, J. M.; Lopez-Rodas, V.; Costas, E. | Rapid adaptation of microalgae to bodies of water with extreme pollution from uranium mining: An explanation of how mesophilic organisms can rapidly colonise extremely toxic environments | AQUATIC TOXICOLOGY |
| 192 | Naik, Ravidas Krishna; Naik, Milind Mohan; D'Costa, Priya Mallika; Shaikh, Fauzia | Microplastics in ballast water as an emerging source and vector for harmful chemicals, antibiotics, metals, bacterial pathogens and HAB species: A potential risk to the marine environment and human health | MARINE POLLUTION BULLETIN |
| 193 | Hazeem, Layla J.; Yesilay, Gamze; Bououdina, Mohamed; Perna, Simone; Cetin, Demet; Suludere, Zekiye; Barras, Alexandre; Boukherroub, Rabah | Investigation of the toxic effects of different polystyrene micro-and nanoplastics on microalgae Chlorella vulgaris by analysis of cell viability, pigment content, oxidative stress and ultrastructural changes | MARINE POLLUTION BULLETIN |
| 194 | Gerardo, Solis-Gonzalez; Alondra Alelie, Cortes-Tellez; Luis, Chacon-Garcia; Martha-Estrella, Garcia-Perez; Hector, Martinez-Flores E.; Maria Carmen, Bartolome Camacho | Prediction of the toxic impact on the freshwater microalgae Scenedesmus intermedius produced by the interaction of copper sulfate and copper oxychloride in a binary mixture with glyphosate | LIMNETICA |
| 195 | Touzet, Nicolas; Keady, Evelyn; Raine, Robin; Maher, Majella | Evaluation of taxa-specific real-time PCR, whole-cell FISH and morphotaxonomy analyses for the detection and quantification of the toxic microalgae Alexandrium minutum (Dinophyceae), Global Clade ribotype | FEMS MICROBIOLOGY ECOLOGY |
| | BEHRINGER, DM; MEYER, KH; VEH, RW | Antibodies against neuroactive amino acids and neuropeptides. II. Simultaneous immunoenzymatic double staining with labeled primary antibodies of the same species and a combination of the ABC method and the hapten-anti-hapten bridge (HAB) technique. | JOURNAL OF HISTOCHEMISTRY & CYTOCHEMISTRY |
| 196 197 | Lourenco, Sergio O.; Barbarino, Elisabete; Bispo, Maria das | Effects of light intensity on growth, inorganic nitrogen storage, and | APLICACOES DA FICOLOGIA: |

| Ī | Gracas S.; Borges, Davilma A.; Coelho-Gomes, Carina; | gross chemical composition of four marine microalgae in batch | ANAIS | DO | XI | CONGRESSO |
|---|--|---|----------|------|-------|-------------|
| | Lavin, Paris L.; Santos, Fernanda | cultures | BRASILE | IRO | DE 1 | FICOLOGIA E |
| | | | SIMPOSIO |) LA | ATINC | O-AMERICANO |
| | | | SOBRE A | LGAS | NOC | IVAS |

2 The taxonomic composition and sequence number of toxic algal species in HTMaDB.

| Species_database | Kingdom | Phylum | Class | Order | Family | Genus | SeqNumber |
|-----------------------------|-----------|----------------|-------------|-------------------|-----------------|-------------|-----------|
| Akashiwo_sanguinea | Alveolata | Dinoflagellata | Dinophyceae | Gymnodiniphycidae | Gymnodiniaceae | Akashiwo | 15 |
| Alexandrium_affine | Alveolata | Dinoflagellata | Dinophyceae | Gonyaulacales | Pyrocystaceae | Alexandrium | 9 |
| Alexandrium_andersonii | Alveolata | Dinoflagellata | Dinophyceae | Gonyaulacales | Pyrocystaceae | Alexandrium | 5 |
| Alexandrium_australiense | Alveolata | Dinoflagellata | Dinophyceae | Gonyaulacales | Pyrocystaceae | Alexandrium | 1 |
| Alexandrium_catenella | Alveolata | Dinoflagellata | Dinophyceae | Gonyaulacales | Pyrocystaceae | Alexandrium | 23 |
| Alexandrium_fragae | Alveolata | Dinoflagellata | Dinophyceae | Gonyaulacales | Pyrocystaceae | Alexandrium | 1 |
| Alexandrium_hiranoi | Alveolata | Dinoflagellata | Dinophyceae | Gonyaulacales | Pyrocystaceae | Alexandrium | 2 |
| Alexandrium_leei | Alveolata | Dinoflagellata | Dinophyceae | Gonyaulacales | Pyrocystaceae | Alexandrium | 1 |
| Alexandrium_limii | Alveolata | Dinoflagellata | Dinophyceae | Gonyaulacales | Pyrocystaceae | Alexandrium | 2 |
| Alexandrium_minutum | Alveolata | Dinoflagellata | Dinophyceae | Gonyaulacales | Pyrocystaceae | Alexandrium | 11 |
| Alexandrium_monilatum | Alveolata | Dinoflagellata | Dinophyceae | Gonyaulacales | Pyrocystaceae | Alexandrium | 1 |
| Alexandrium_ogatae | Alveolata | Dinoflagellata | Dinophyceae | Gonyaulacales | Pyrocystaceae | Alexandrium | 2 |
| Alexandrium_ostenfeldii | Alveolata | Dinoflagellata | Dinophyceae | Gonyaulacales | Pyrocystaceae | Alexandrium | 50 |
| Alexandrium_pacificum | Alveolata | Dinoflagellata | Dinophyceae | Gonyaulacales | Pyrocystaceae | Alexandrium | 2 |
| Alexandrium_pseudogonyaulax | Alveolata | Dinoflagellata | Dinophyceae | Gonyaulacales | Pyrocystaceae | Alexandrium | 6 |
| Alexandrium_tamarense | Alveolata | Dinoflagellata | Dinophyceae | Gonyaulacales | Pyrocystaceae | Alexandrium | 59 |
| Alexandrium_tamiyavanichii | Alveolata | Dinoflagellata | Dinophyceae | Gonyaulacales | Pyrocystaceae | Alexandrium | 1 |
| Alexandrium_taylorii | Alveolata | Dinoflagellata | Dinophyceae | Gonyaulacales | Pyrocystaceae | Alexandrium | 1 |
| Amphidinium_carterae | Alveolata | Dinoflagellata | Dinophyceae | Gymnodiniphycidae | Amphidiniaceae | Amphidinium | 8 |
| Amphidinium_gibbosum | Alveolata | Dinoflagellata | Dinophyceae | Gymnodiniphycidae | Amphidiniaceae | Amphidinium | 1 |
| Amphidinium_klebsii | Alveolata | Dinoflagellata | Dinophyceae | Gymnodiniphycidae | Amphidiniaceae | Amphidinium | 1 |
| Amphidinium_operculatum | Alveolata | Dinoflagellata | Dinophyceae | Gymnodiniphycidae | Amphidiniaceae | Amphidinium | 2 |
| Amphidoma_languida | Alveolata | Dinoflagellata | Dinophyceae | Gonyaulacales | Amphidomataceae | Amphidoma | 3 |

| Azadinium_dexteroporum | Alveolata | Dinoflagellata | Dinophyceae | Peridiniales | Amphidomataceae | Azadinium | 2 |
|----------------------------|-----------|----------------|-------------|------------------|-----------------|---------------|----|
| Azadinium_poporum | Alveolata | Dinoflagellata | Dinophyceae | Peridiniales | Amphidomataceae | Azadinium | 3 |
| Azadinium_spinosum | Alveolata | Dinoflagellata | Dinophyceae | Peridiniales | Amphidomataceae | Azadinium | 5 |
| Centrodinium_punctatum | Alveolata | Dinoflagellata | Dinophyceae | Gonyaulacales | Ostreopsidaceae | Centrodinium | 2 |
| Coolia_canariensis | Alveolata | Dinoflagellata | Dinophyceae | Peridiniphycidae | Gonyaulacales | Coolia | 2 |
| Coolia_malayensis | Alveolata | Dinoflagellata | Dinophyceae | Peridiniphycidae | Gonyaulacales | Coolia | 1 |
| Dinophysis_acuminata | Alveolata | Dinoflagellata | Dinophyceae | Dinophysiales | Dinophysiaceae | Dinophysis | 5 |
| Dinophysis_acuta | Alveolata | Dinoflagellata | Dinophyceae | Dinophysiales | Dinophysiaceae | Dinophysis | 1 |
| Dinophysis_caudata | Alveolata | Dinoflagellata | Dinophyceae | Dinophysiales | Dinophysiaceae | Dinophysis | 2 |
| Dinophysis_fortii | Alveolata | Dinoflagellata | Dinophyceae | Dinophysiales | Dinophysiaceae | Dinophysis | 1 |
| Dinophysis_infundibulum | Alveolata | Dinoflagellata | Dinophyceae | Dinophysiales | Dinophysiaceae | Dinophysis | 1 |
| Dinophysis_miles | Alveolata | Dinoflagellata | Dinophyceae | Dinophysiales | Dinophysiaceae | Dinophysis | 1 |
| Dinophysis_norvegica | Alveolata | Dinoflagellata | Dinophyceae | Dinophysiales | Dinophysiaceae | Dinophysis | 3 |
| Dinophysis_tripos | Alveolata | Dinoflagellata | Dinophyceae | Dinophysiales | Dinophysiaceae | Dinophysis | 1 |
| Fukuyoa_paulensis | Alveolata | Dinoflagellata | Dinophyceae | Gonyaulacales | Pyrocystaceae | Fukuyoa | 1 |
| Fukuyoa_ruetzleri | Alveolata | Dinoflagellata | Dinophyceae | Gonyaulacales | Pyrocystaceae | Fukuyoa | 2 |
| Fukuyoa_yasumotoi | Alveolata | Dinoflagellata | Dinophyceae | Gonyaulacales | Pyrocystaceae | Fukuyoa | 10 |
| Gambierdiscus_australes | Alveolata | Dinoflagellata | Dinophyceae | Gonyaulacales | Pyrocystaceae | Gambierdiscus | 18 |
| Gambierdiscus_balechii | Alveolata | Dinoflagellata | Dinophyceae | Gonyaulacales | Pyrocystaceae | Gambierdiscus | 4 |
| Gambierdiscus_belizeanus | Alveolata | Dinoflagellata | Dinophyceae | Gonyaulacales | Pyrocystaceae | Gambierdiscus | 7 |
| Gambierdiscus_caribaeus | Alveolata | Dinoflagellata | Dinophyceae | Gonyaulacales | Pyrocystaceae | Gambierdiscus | 1 |
| Gambierdiscus_carolinianus | Alveolata | Dinoflagellata | Dinophyceae | Gonyaulacales | Pyrocystaceae | Gambierdiscus | 5 |
| Gambierdiscus_carpenteri | Alveolata | Dinoflagellata | Dinophyceae | Gonyaulacales | Pyrocystaceae | Gambierdiscus | 8 |
| Gambierdiscus_cheloniae | Alveolata | Dinoflagellata | Dinophyceae | Gonyaulacales | Pyrocystaceae | Gambierdiscus | 1 |
| Gambierdiscus_excentricus | Alveolata | Dinoflagellata | Dinophyceae | Gonyaulacales | Pyrocystaceae | Gambierdiscus | 1 |
| Gambierdiscus_holmesii | Alveolata | Dinoflagellata | Dinophyceae | Gonyaulacales | Pyrocystaceae | Gambierdiscus | 2 |
| | | | | | | | |

| Gambierdiscus_honu | Alveolata | Dinoflagellata | Dinophyceae | Gonyaulacales | Pyrocystaceae | Gambierdiscus | 1 |
|---|-----------|----------------|-------------|---------------|------------------|-----------------|----|
| Gambierdiscus_lapillus | Alveolata | Dinoflagellata | Dinophyceae | Gonyaulacales | Pyrocystaceae | Gambierdiscus | 1 |
| Gambierdiscus_lewisii | Alveolata | Dinoflagellata | Dinophyceae | Gonyaulacales | Pyrocystaceae | Gambierdiscus | 1 |
| Gambierdiscus_pacificus | Alveolata | Dinoflagellata | Dinophyceae | Gonyaulacales | Pyrocystaceae | Gambierdiscus | 10 |
| Gambierdiscus_polynesiensis | Alveolata | Dinoflagellata | Dinophyceae | Gonyaulacales | Pyrocystaceae | Gambierdiscus | 6 |
| Gambierdiscus_scabrosus | Alveolata | Dinoflagellata | Dinophyceae | Gonyaulacales | Pyrocystaceae | Gambierdiscus | 55 |
| Gambierdiscus_silvae | Alveolata | Dinoflagellata | Dinophyceae | Gonyaulacales | Pyrocystaceae | Gambierdiscus | 2 |
| Gambierdiscus_toxicus | Alveolata | Dinoflagellata | Dinophyceae | Gonyaulacales | Pyrocystaceae | Gambierdiscus | 14 |
| Gonyaulax_bohaiensis | Alveolata | Dinoflagellata | Dinophyceae | Gonyaulacales | Gonyaulacaceae | Gonyaulax | 1 |
| Gonyaulax_spinifera | Alveolata | Dinoflagellata | Dinophyceae | Gonyaulacales | Gonyaulacaceae | Gonyaulax | 6 |
| Gymnodinium_catenatum | Alveolata | Dinoflagellata | Dinophyceae | Gymnodiniales | Gymnodiniaceae | Gymnodinium | 11 |
| Heterocapsa_bohaiensis | Alveolata | Dinoflagellata | Dinophyceae | Peridiniales | Heterocapsaceae | Heterocapsa | 2 |
| Heterocapsa_borneoensis | Alveolata | Dinoflagellata | Dinophyceae | Peridiniales | Heterocapsaceae | Heterocapsa | 1 |
| Heterocapsa_circularisquama | Alveolata | Dinoflagellata | Dinophyceae | Peridiniales | Heterocapsaceae | Heterocapsa | 1 |
| Heterocapsa_horiguchii | Alveolata | Dinoflagellata | Dinophyceae | Peridiniales | Heterocapsaceae | Heterocapsa | 1 |
| Heterocapsa_niei | Alveolata | Dinoflagellata | Dinophyceae | Peridiniales | Heterocapsaceae | Heterocapsa | 3 |
| Heterocapsa_pygmaea | Alveolata | Dinoflagellata | Dinophyceae | Peridiniales | Heterocapsaceae | Heterocapsa | 10 |
| Karenia_bicuneiformis | Alveolata | Dinoflagellata | Dinophyceae | Gymnodiniales | Kareniaceae | Karenia | 1 |
| Karenia_brevis | Alveolata | Dinoflagellata | Dinophyceae | Gymnodiniales | Kareniaceae | Karenia | 14 |
| Karenia_mikimotoi | Alveolata | Dinoflagellata | Dinophyceae | Gymnodiniales | Kareniaceae | Karenia | 7 |
| Karenia_papilionacea | Alveolata | Dinoflagellata | Dinophyceae | Gymnodiniales | Kareniaceae | Karenia | 1 |
| Karenia_selliformis | Alveolata | Dinoflagellata | Dinophyceae | Gymnodiniales | Kareniaceae | Karenia | 1 |
| Karlodinium_veneficum | Alveolata | Dinoflagellata | Dinophyceae | Gymnodiniales | Kareniaceae | Karlodinium | 25 |
| Lingulodinium_polyedra | Alveolata | Dinoflagellata | Dinophyceae | Gonyaulacales | Lingulodiniaceae | Lingulodinium | 10 |
| ${\it Margale fidinium_fulve scens}$ | Alveolata | Dinoflagellata | Dinophyceae | Gymnodiniales | Gymnodiniaceae | Margalefidinium | 3 |
| ${\it Margale fidinium_polykrikoides}$ | Alveolata | Dinoflagellata | Dinophyceae | Gymnodiniales | Gymnodiniaceae | Margalefidinium | 27 |
| | | | | | | | |

| Ostreopsis_ovata | Alveolata | Dinoflagellata | Dinophyceae | Gonyaulacales | Pyrocystaceae | Ostreopsis | 3 |
|------------------------------------|---------------|----------------|------------------|-------------------|-----------------|-----------------|----|
| Ostreopsis_rhodesiae | Alveolata | Dinoflagellata | Dinophyceae | Gonyaulacales | Pyrocystaceae | Ostreopsis | 3 |
| Ostreopsis_siamensis | Alveolata | Dinoflagellata | Dinophyceae | Gonyaulacales | Pyrocystaceae | Ostreopsis | 2 |
| Pfiesteria_piscicida | Alveolata | Dinoflagellata | Dinophyceae | Thoracosphaerales | Pfiesteriaceae | Pfiesteria | 35 |
| Pfiesteria_shumwayae | Alveolata | Dinoflagellata | Dinophyceae | Thoracosphaerales | Pfiesteriaceae | Pfiesteria | 1 |
| Phalacroma_mitra | Alveolata | Dinoflagellata | Dinophyceae | Dinophysiales | Oxyphysiaceae | Phalacroma | 4 |
| Phalacroma_rotundatum | Alveolata | Dinoflagellata | Dinophyceae | Dinophysiales | Oxyphysaceae | Phalacroma | 2 |
| Polykrikos_hartmannii | Alveolata | Dinoflagellata | Dinophyceae | Gymnodiniales | Gymnodiniaceae | Polykrikos | 1 |
| Prorocentrum_cassubicum | Alveolata | Dinoflagellata | Dinophyceae | Prorocentrales | Prorocentraceae | Prorocentrum | 1 |
| Prorocentrum_concavum | Alveolata | Dinoflagellata | Dinophyceae | Prorocentrales | Prorocentraceae | Prorocentrum | 1 |
| Prorocentrum_cordatum | Alveolata | Dinoflagellata | Dinophyceae | Prorocentrales | Prorocentraceae | Prorocentrum | 12 |
| Prorocentrum_emarginatum | Alveolata | Dinoflagellata | Dinophyceae | Prorocentrales | Prorocentraceae | Prorocentrum | 2 |
| Prorocentrum_hoffmannianum | Alveolata | Dinoflagellata | Dinophyceae | Prorocentrales | Prorocentraceae | Prorocentrum | 2 |
| Prorocentrum_leve | Alveolata | Dinoflagellata | Dinophyceae | Prorocentrales | Prorocentraceae | Prorocentrum | 3 |
| Prorocentrum_lima | Alveolata | Dinoflagellata | Dinophyceae | Prorocentrales | Prorocentraceae | Prorocentrum | 12 |
| Prorocentrum_mexicanum | Alveolata | Dinoflagellata | Dinophyceae | Prorocentrales | Prorocentraceae | Prorocentrum | 3 |
| Prorocentrum_panamense | Alveolata | Dinoflagellata | Dinophyceae | Prorocentrales | Prorocentraceae | Prorocentrum | 1 |
| Prorocentrum_rhathymum | Alveolata | Dinoflagellata | Dinophyceae | Prorocentrales | Prorocentraceae | Prorocentrum | 6 |
| Prorocentrum_texanum | Alveolata | Dinoflagellata | Dinophyceae | Prorocentrales | Prorocentraceae | Prorocentrum | 1 |
| Protoceratium_reticulatum | Alveolata | Dinoflagellata | Dinophyceae | Gonyaulacales | Pyrocystaceae | Protoceratium | 6 |
| Protoperidinium_crassipes | Alveolata | Dinoflagellata | Dinophyceae | Peridiniphycidae | Peridiniales | Protoperidinium | 3 |
| Pyrodinium_bahamense | Alveolata | Dinoflagellata | Dinophyceae | Gonyaulacales | Pyrocystaceae | Pyrodinium | 6 |
| Pyrodinium_bahamense_varcompressum | Alveolata | Dinoflagellata | Dinophyceae | Gonyaulacales | Pyrocystaceae | Pyrodinium | 6 |
| Takayama_acrotrocha | Alveolata | Dinoflagellata | Dinophyceae | Gymnodiniphycidae | Kareniaceae | Takayama | 1 |
| Vicicitus globosus | | | | 51. | 51. 1 | | |
| 3 | Stramenopiles | Ochrophyta | Dictyochophyceae | Dictyochales | Dictyochaceae | Vicicitus | 1 |

| Chattonella_marina_varantiqua | Stramenopiles | Ochrophyta | Raphidophyceae | Chattonellales | Chattonellaceae | Chattonella | 4 |
|--------------------------------------|---------------|------------|------------------|-------------------|--------------------------------|-------------------|----|
| Chattonella_marina_varovata | Stramenopiles | Ochrophyta | Raphidophyceae | Chattonellales | Chattonellaceae | Chattonella | 3 |
| Chattonella_subsalsa | Stramenopiles | Ochrophyta | Raphidophyceae | Chattonellales | Chattonellaceae | Chattonella | 12 |
| Fibrocapsa_japonica | Stramenopiles | Ochrophyta | Raphidophyceae | Chattonellales | Chattonellaceae | Fibrocapsa | 4 |
| Heterosigma_akashiwo | Stramenopiles | Ochrophyta | Raphidophyceae | Chattonellales | Chattonellaceae | Heterosigma | 53 |
| Nitzschia_bizertensis | Stramenopiles | Ochrophyta | Diatomea | Bacillariophytina | Bacillariophyceae | Nitzschia | 1 |
| Pseudochattonella_farcimen | Stramenopiles | Ochrophyta | Dictyochophyceae | Florenciellales | Florenciellales incertae sedis | Pseudochattonella | 3 |
| $Pseudochattonella_verruculosa$ | Stramenopiles | Ochrophyta | Dictyochophyceae | Florenciellales | Florenciellales incertae sedis | Pseudochattonella | 6 |
| Pseudo-nitzschia_australis | Stramenopiles | Ochrophyta | Diatomea | Bacillariophytina | Bacillariophyceae | Pseudo-nitzschia | 3 |
| Pseudo-nitzschia_batesiana | Stramenopiles | Ochrophyta | Diatomea | Bacillariophytina | Bacillariophyceae | Pseudo-nitzschia | 1 |
| Pseudo-nitzschia_brasiliana | Stramenopiles | Ochrophyta | Diatomea | Bacillariophytina | Bacillariophyceae | Pseudo-nitzschia | 3 |
| Pseudo-nitzschia_caciantha | Stramenopiles | Ochrophyta | Diatomea | Bacillariophytina | Bacillariophyceae | Pseudo-nitzschia | 1 |
| Pseudo-nitzschia_calliantha | Stramenopiles | Ochrophyta | Diatomea | Bacillariophytina | Bacillariophyceae | Pseudo-nitzschia | 1 |
| Pseudo-nitzschia_cuspidata | Stramenopiles | Ochrophyta | Diatomea | Bacillariophytina | Bacillariophyceae | Pseudo-nitzschia | 2 |
| Pseudo-nitzschia_delicatissima | Stramenopiles | Ochrophyta | Diatomea | Bacillariophytina | Bacillariophyceae | Pseudo-nitzschia | 1 |
| Pseudo-nitzschia_fraudulenta | Stramenopiles | Ochrophyta | Diatomea | Bacillariophytina | Bacillariophyceae | Pseudo-nitzschia | 1 |
| Pseudo-nitzschia_fukuyoi | Stramenopiles | Ochrophyta | Diatomea | Bacillariophytina | Bacillariophyceae | Pseudo-nitzschia | 3 |
| Pseudo-nitzschia_galaxiae | Stramenopiles | Ochrophyta | Diatomea | Bacillariophytina | Bacillariophyceae | Pseudo-nitzschia | 1 |
| Pseudo-nitzschia_granii | Stramenopiles | Ochrophyta | Diatomea | Bacillariophytina | Bacillariophyceae | Pseudo-nitzschia | 1 |
| Pseudo-nitzschia_kodamae | Stramenopiles | Ochrophyta | Diatomea | Bacillariophytina | Bacillariophyceae | Pseudo-nitzschia | 1 |
| Pseudo-nitzschia_lundholmiae | Stramenopiles | Ochrophyta | Diatomea | Bacillariophytina | Bacillariophyceae | Pseudo-nitzschia | 2 |
| Pseudo-nitzschia_multiseries | Stramenopiles | Ochrophyta | Diatomea | Bacillariophytina | Bacillariophyceae | Pseudo-nitzschia | 6 |
| Pseudo-nitzschia_pseudodelicatissima | Stramenopiles | Ochrophyta | Diatomea | Bacillariophytina | Bacillariophyceae | Pseudo-nitzschia | 1 |
| Pseudo-nitzschia_pungens | Stramenopiles | Ochrophyta | Diatomea | Bacillariophytina | Bacillariophyceae | Pseudo-nitzschia | 4 |
| Pseudo-nitzschia_seriata | Stramenopiles | Ochrophyta | Diatomea | Bacillariophytina | Bacillariophyceae | Pseudo-nitzschia | 1 |
| Pseudo-nitzschia_simulans | Stramenopiles | Ochrophyta | Diatomea | Bacillariophytina | Bacillariophyceae | Pseudo-nitzschia | 1 |
| | | | | | | | |

| Pseudo-nitzschia_subcurvata | Stramenopiles | Ochrophyta | Diatomea | Bacillariophytina | Bacillariophyceae | Pseudo-nitzschia | 1 |
|---------------------------------|---------------|------------|-------------------|-------------------|----------------------|------------------|----|
| Pseudo-nitzschia_turgidula | Stramenopiles | Ochrophyta | Diatomea | Bacillariophytina | Bacillariophyceae | Pseudo-nitzschia | 2 |
| $Chrysochromulina_leadbeateri$ | Chromista | Haptophyta | Prymnesiophyceae | Prymnesiales | Chrysochromulinaceae | Chrysochromulina | 1 |
| Phaeocystis_globosa | Chromista | Haptophyta | Prymnesiophyceae | Phaeocystales | Phaeocystaceae | Phaeocystis | 22 |
| Phaeocystis_pouchetii | Chromista | Haptophyta | Prymnesiophyceae | Phaeocystales | Phaeocystaceae | Phaeocystis | 4 |
| Prymnesium_calathiferum | Chromista | Haptophyta | Coccolithophyceae | Prymnesiales | Prymnesiaceae | Prymnesium | 2 |
| Prymnesium_faveolatum | Chromista | Haptophyta | Coccolithophyceae | Prymnesiales | Prymnesiaceae | Prymnesium | 1 |
| Prymnesium_parvum | Chromista | Haptophyta | Coccolithophyceae | Prymnesiales | Prymnesiaceae | Prymnesium | 4 |
| Prymnesium_polylepis | Chromista | Haptophyta | Coccolithophyceae | Prymnesiales | Prymnesiaceae | Prymnesium | 17 |
| Prymnesium_zebrinum | Chromista | Haptophyta | Coccolithophyceae | Prymnesiales | Prymnesiaceae | Prymnesium | 1 |

3 The taxonomic composition and sequence number of harmful algal species in HTMaDB.

| Species_database | Kingdom | Phylum | Class | Order | Family | Genus | SeqNumber |
|------------------------------|---------------|----------------|---------------|--------------------|---------------------|------------------|-----------|
| Alexandrium_cohorticula | Alveolata | Dinoflagellata | Dinophyceae | Gonyaulacales | Pyrocystaceae | Alexandrium | 1 |
| Alexandrium_margalefii | Alveolata | Dinoflagellata | Dinophyceae | Gonyaulacales | Pyrocystaceae | Alexandrium | 1 |
| Alexandrium_insuetum | Alveolata | Dinoflagellata | Dinophyceae | Gonyaulacales | Pyrocystaceae | Alexandrium | 2 |
| Asterionellopsis_glacialis | Stramenopiles | Ochrophyta | Diatomea | Bacillariophytina | Bacillariophyceae | Asterionellopsis | 4 |
| Aureococcus_anophagefferens | Stramenopiles | Ochrophyta | Pelagophyceae | Pelagomonadales | Pelagomonadaceae | Aureococcus | 20 |
| Aureoumbra_lagunensis | Stramenopiles | Ochrophyta | Pelagophyceae | Sarcinochrysidales | Sarcinochrysidaceae | Aureoumbra | 3 |
| Blixaea_quinquecornis | Alveolata | Dinoflagellata | Dinophyceae | Peridiniphycidae | Peridiniales | Blixaea | 2 |
| Chaetoceros_affinis | Stramenopiles | Ochrophyta | Diatomea | Chaetocerotales | Chaetocerotaceae | Chaetoceros | 1 |
| Chaetoceros_atlanticus | Stramenopiles | Ochrophyta | Diatomea | Chaetocerotales | Chaetocerotaceae | Chaetoceros | 1 |
| Chaetoceros_constrictus | Stramenopiles | Ochrophyta | Diatomea | Chaetocerotales | Chaetocerotaceae | Chaetoceros | 2 |
| Chaetoceros_convolutus | Stramenopiles | Ochrophyta | Diatomea | Chaetocerotales | Chaetocerotaceae | Chaetoceros | 1 |
| Chaetoceros_curvisetus | Stramenopiles | Ochrophyta | Diatomea | Chaetocerotales | Chaetocerotaceae | Chaetoceros | 1 |
| Chaetoceros_danicus | Stramenopiles | Ochrophyta | Diatomea | Chaetocerotales | Chaetocerotaceae | Chaetoceros | 4 |
| Chaetoceros_debilis | Stramenopiles | Ochrophyta | Diatomea | Chaetocerotales | Chaetocerotaceae | Chaetoceros | 3 |
| Chaetoceros_decipiens | Stramenopiles | Ochrophyta | Diatomea | Chaetocerotales | Chaetocerotaceae | Chaetoceros | 1 |
| Chaetoceros_diadema | Stramenopiles | Ochrophyta | Diatomea | Chaetocerotales | Chaetocerotaceae | Chaetoceros | 1 |
| Chaetoceros_laevisporus | Stramenopiles | Ochrophyta | Diatomea | Chaetocerotales | Chaetocerotaceae | Chaetoceros | 3 |
| Chaetoceros_pseudocurvisetus | Stramenopiles | Ochrophyta | Diatomea | Chaetocerotales | Chaetocerotaceae | Chaetoceros | 2 |
| Chaetoceros_rostratus | Stramenopiles | Ochrophyta | Diatomea | Chaetocerotales | Chaetocerotaceae | Chaetoceros | 5 |
| Chaetoceros_socialis | Stramenopiles | Ochrophyta | Diatomea | Chaetocerotales | Chaetocerotaceae | Chaetoceros | 2 |
| Chaetoceros_simplex | Stramenopiles | Ochrophyta | Diatomea | Chaetocerotales | Chaetocerotaceae | Chaetoceros | 1 |
| Chaetoceros_tenuissimus | Stramenopiles | Ochrophyta | Diatomea | Chaetocerotales | Chaetocerotaceae | Chaetoceros | 1 |
| Chaetoceros_throndsenii | Stramenopiles | Ochrophyta | Diatomea | Chaetocerotales | Chaetocerotaceae | Chaetoceros | 4 |

| Chaetoceros_tortissimus | Stramenopiles | Ochrophyta | Diatomea | Chaetocerotales | Chaetocerotaceae | Chaetoceros | 2 |
|----------------------------|----------------|----------------|------------------|---------------------|-------------------|----------------|----|
| Chaetoceros_lorenzianus | Stramenopiles | Ochrophyta | Diatomea | Chaetocerotales | Chaetocerotaceae | Chaetoceros | 1 |
| Chaetoceros_wighamii | Stramenopiles | Ochrophyta | Diatomea | Chaetocerotales | Chaetocerotaceae | Chaetoceros | 1 |
| Coscinodiscus_radiatus | Stramenopiles | Ochrophyta | Diatomea | Coscinodiscales | Coscinodiscaceae | Coscinodiscus | 4 |
| Coolia_monotis | Alveolata | Dinoflagellata | Dinophyceae | Peridiniphycidae | Gonyaulacales | Coolia | 4 |
| Cylindrotheca_closterium | Stramenopiles | Ochrophyta | Diatomea | Bacillariophytina | Bacillariophyceae | Cylindrotheca | 46 |
| Cyclotella_meneghiniana | Stramenopiles | Ochrophyta | Diatomea | Bacillariophytina | Mediophyceae | Cyclotella | 29 |
| Dictyocha_fibula | Stramenopiles | Ochrophyta | Dictyochophyceae | Dictyochales | Dictyochaceae | Dictyocha | 1 |
| Dictyocha_speculum | Stramenopiles | Ochrophyta | Dictyochophyceae | Dictyochales | Dictyochaceae | Dictyocha | 1 |
| Dinophysis_hastata | Alveolata | Dinoflagellata | Dinophyceae | Dinophysiales | Dinophysaceae | Dinophysis | 1 |
| Diplopsalis_lenticula | Alveolata | Dinoflagellata | Dinophyceae | Peridiniphycidae | Peridiniales | Diplopsalopsis | 1 |
| Dunaliella_salina | Chloroplastida | Chlorophyta | Chlorophyceae | Chlamydomonadales | Dunaliellaceae | Dunaliella | 39 |
| Eucampia_zodiacus | Stramenopiles | Ochrophyta | Diatomea | Bacillariophytina | Mediophyceae | Eucampia | 1 |
| Eutreptiella_gymnastica | Discoba | Euglenozoa | Euglenophyceae | Eutreptiales | Eutreptiaceae | Eutreptiella | 5 |
| Euglena_viridis | Discoba | Euglenozoa | Euglenophyceae | Euglenales | Euglenaceae | Euglena | 1 |
| Gonyaulax_digitale | Alveolata | Dinoflagellata | Dinophyceae | Peridiniphycidae | Gonyaulacales | Gonyaulax | 1 |
| Gonyaulax_polygramma | Alveolata | Dinoflagellata | Dinophyceae | Peridiniphycidae | Gonyaulacales | Gonyaulax | 2 |
| Gonyaulax_verior | Alveolata | Dinoflagellata | Dinophyceae | Peridiniphycidae | Gonyaulacales | Gonyaulax | 1 |
| Gymnodinium_aureolum | Alveolata | Dinoflagellata | Dinophyceae | Gymnodiniphycidae | Gymnodiniaceae | Gymnodinium | 7 |
| Gymnodinium_impudicum | Alveolata | Dinoflagellata | Dinophyceae | Gymnodiniphycidae | Gymnodiniaceae | Gymnodinium | 5 |
| Guinardia_delicatula | Stramenopiles | Ochrophyta | Diatomea | Coscinodiscophytina | Rhizosolenids | Guinardia | 5 |
| Guinardia_flaccida | Stramenopiles | Ochrophyta | Diatomea | Coscinodiscophytina | Rhizosolenids | Guinardia | 5 |
| $Halamphora_coffeiformis$ | Stramenopiles | Ochrophyta | Diatomea | Bacillariophytina | Bacillariophyceae | Halamphora | 8 |
| Heterocapsa_rotundata | Alveolata | Dinoflagellata | Dinophyceae | Peridiniphycidae | Peridiniales | Heterocapsa | 2 |
| Heterocapsa_triquetra | Alveolata | Dinoflagellata | Dinophyceae | Peridiniphycidae | Peridiniales | Heterocapsa | 5 |
| Heterocapsa_pygmaea | Alveolata | Dinoflagellata | Dinophyceae | Peridiniphycidae | Peridiniales | Heterocapsa | 2 |
| | | | | | | | |

| Lepidodinium_chlorophorum Alveolata Dinoflagellata Dinophyceae Gymnodiniphycidae Gymnodiniaceae Lepidodinium Leptocylindrus_danicus Stramenopiles Ochrophyta Diatomea Coscinodiscophytina Rhizosolenids Leptocylindrus | 4 4 2 |
|--|-------------|
| | • |
| | 2 |
| Leptocylindrus_minimus Stramenopiles Ochrophyta Diatomea Coscinodiscophytina Rhizosolenids Leptocylindrus | _ |
| Levanderina_fissa Alveolata Dinoflagellata Dinophyceae Gymnodiniphycidae Gymnodiniales familia incertae sedis Levanderina | 1 |
| Lithodesmioides_polymorpha Stramenopiles Ochrophyta Diatomea Bacillariophytina Mediophyceae Lithodesmioides | 1 |
| Noctiluca_scintillans Alveolata Dinoflagellata Dinophyceae Noctilucales Noctilucaceae Noctiluca | 5 |
| Nitzschia_longissima Stramenopiles Ochrophyta Diatomea Bacillariophytina Bacillariophyceae Nitzschia | 3 |
| Ostreopsis_lenticularis Alveolata Dinoflagellata Dinophyceae Gonyaulacales Pyrocystaceae Ostreopsis | 1 |
| Paralia_sulcata Stramenopiles Ochrophyta Diatomea Coscinodiscophytina Melosirids Paralia | 2 |
| Phalacroma_rapa Alveolata Dinoflagellata Dinophyceae Dinophysiales Oxyphysiaceae Phalacroma | 3 |
| Plagioselmis_prolonga Chromista Cryptophyta Cryptophyceae Pyrenomonadales Geminigeraceae Plagioselmis | 2 |
| Plagioselmis_nannoplanctica Chromista Cryptophyta Cryptophyceae Pyrenomonadales Geminigeraceae Plagioselmis | 3 |
| Proboscia_alata Stramenopiles Ochrophyta Diatomea Coscinodiscophytina Rhizosolenids Proboscia | 2 |
| Prorocentrum_dentatum Alveolata Dinoflagellata Dinophyceae Prorocentrales Prorocentraceae Prorocentrum | 3 |
| Prorocentrum_donghaiense Alveolata Dinoflagellata Dinophyceae Prorocentrales Prorocentraceae Prorocentrum | 5 |
| Prorocentrum_gracile Alveolata Dinoflagellata Dinophyceae Prorocentrales Prorocentraceae Prorocentrum | 1 |
| Prorocentrum_micans Alveolata Dinoflagellata Dinophyceae Prorocentrales Prorocentraceae Prorocentrum | 10 |
| Prorocentrum_redfieldii Alveolata Dinoflagellata Dinophyceae Prorocentrales Prorocentraceae Prorocentrum | 3 |
| Prorocentrum_triestinum Alveolata Dinoflagellata Dinophyceae Prorocentrales Prorocentraceae Prorocentrum | 3 |
| Protodinium_simplex Alveolata Dinoflagellata Dinophyceae Gymnodiniphycidae Suessiaceae Protodinium | 3 |
| Pseudocochlodinium_profundisulcus Alveolata Dinoflagellata Dinophyceae Dinophyceae ordo incertae sedis Dinophyceae familia incertae sedis Pseudocochlodini | ım 1 |
| Pyrophacus_steinii Alveolata Dinoflagellata Dinophyceae Gonyaulacales Pyrocystaceae Pyrophacus | 1 |
| Scrippsiella_acuminata Alveolata Dinoflagellata Dinophyceae Peridiniphycidae Thoracosphaeraceae Scrippsiella | 8 |
| Scrippsiella_trochoidea Alveolata Dinoflagellata Dinophyceae Peridiniphycidae Thoracosphaeraceae Scrippsiella | 6 |
| Skeletonema_tropicum Stramenopiles Ochrophyta Diatomea Bacillariophytina Mediophyceae Skeletonema | 7 |

| Skeletonema_costatum | Stramenopiles | Ochrophyta | Diatomea | Bacillariophytina | Mediophyceae | Skeletonema | 29 |
|--------------------------------|---------------|----------------|---------------|---------------------|-------------------|----------------|----|
| Skeletonema_marinoi | Stramenopiles | Ochrophyta | Diatomea | Bacillariophytina | Mediophyceae | Skeletonema | 48 |
| Skeletonema_menzelii | Stramenopiles | Ochrophyta | Diatomea | Bacillariophytina | Mediophyceae | Skeletonema | 1 |
| Skeletonema_pseudocostatum | Stramenopiles | Ochrophyta | Diatomea | Bacillariophytina | Mediophyceae | Skeletonema | 17 |
| Takayama_pulchella | Alveolata | Dinoflagellata | Dinophyceae | Gymnodiniphycidae | Kareniaceae | Takayama | 1 |
| Takayama_xiamenensis | Alveolata | Dinoflagellata | Dinophyceae | Gymnodiniphycidae | Kareniaceae | Takayama | 1 |
| Teleaulax_acuta | Chromista | Cryptophyta | Cryptophyceae | Pyrenomonadales | Geminigeraceae | Teleaulax | 3 |
| Tenuicylindrus_belgicus | Stramenopiles | Ochrophyta | Diatomea | Coscinodiscophytina | Rhizosolenids | Tenuicylindrus | 3 |
| Thalassiosira_allenii | Stramenopiles | Ochrophyta | Diatomea | Bacillariophytina | Mediophyceae | Thalassiosira | 1 |
| Thalassiosira_curviseriata | Stramenopiles | Ochrophyta | Diatomea | Bacillariophytina | Mediophyceae | Thalassiosira | 2 |
| Thalassiosira_gravida | Stramenopiles | Ochrophyta | Diatomea | Bacillariophytina | Mediophyceae | Thalassiosira | 8 |
| Thalassiosira_lundiana | Stramenopiles | Ochrophyta | Diatomea | Bacillariophytina | Mediophyceae | Thalassiosira | 1 |
| Thalassiosira_mala | Stramenopiles | Ochrophyta | Diatomea | Bacillariophytina | Mediophyceae | Thalassiosira | 1 |
| Thalassiosira_minima | Stramenopiles | Ochrophyta | Diatomea | Bacillariophytina | Mediophyceae | Thalassiosira | 5 |
| Thalassiosira_minuscula | Stramenopiles | Ochrophyta | Diatomea | Bacillariophytina | Mediophyceae | Thalassiosira | 3 |
| $Thalassionema_nitzschioides$ | Stramenopiles | Ochrophyta | Diatomea | Bacillariophytina | Mediophyceae | Thalassionema | 1 |
| Thalassiosira_pseudonana | Stramenopiles | Ochrophyta | Diatomea | Bacillariophytina | Mediophyceae | Thalassiosira | 18 |
| Thalassiosira_punctigera | Stramenopiles | Ochrophyta | Diatomea | Bacillariophytina | Mediophyceae | Thalassiosira | 7 |
| Thalassiosira_rotula | Stramenopiles | Ochrophyta | Diatomea | Bacillariophytina | Mediophyceae | Thalassiosira | 19 |
| Thalassiosira_weissflogii | Stramenopiles | Ochrophyta | Diatomea | Bacillariophytina | Mediophyceae | Thalassiosira | 10 |
| Tripos_dens | Alveolata | Dinoflagellata | Dinophyceae | Gonyaulacales | Ceratiaceae | Tripos | 1 |
| Tripos_furca | Alveolata | Dinoflagellata | Dinophyceae | Gonyaulacales | Ceratiaceae | Tripos | 6 |
| Tripos_fusus | Alveolata | Dinoflagellata | Dinophyceae | Gonyaulacales | Ceratiaceae | Tripos | 2 |
| Tripos_lineatus | Alveolata | Dinoflagellata | Dinophyceae | Gonyaulacales | Ceratiaceae | Tripos | 1 |
| Tripos_muelleri | Alveolata | Dinoflagellata | Dinophyceae | Gonyaulacales | Ceratiaceae | Tripos | 1 |
| $Tryblionella_compressa$ | Stramenopiles | Ochrophyta | Diatomea | Bacillariophytina | Bacillariophyceae | Tryblionella | 2 |
| | | | | | | | |

Appendix 2

This section provides detailed sampling information in this study.

1 The detailed information of sampling locations in Hong Kong coastal seawater.

| Sample ID | Description | Longitude | Latitude |
|-----------|-----------------|------------|-----------|
| S1 | Eastern Waters | 114.218333 | 22.408611 |
| S2 | Eastern Waters | 114.205 | 22.450556 |
| S3 | Eastern Waters | 114.274167 | 22.46 |
| S4 | Eastern Waters | 114.275278 | 22.434444 |
| S5 | Eastern Waters | 114.335 | 22.476944 |
| S6 | Eastern Waters | 114.3071 | 22.523683 |
| S7 | Eastern Waters | 114.3618 | 22.57305 |
| S8 | Eastern Waters | 114.40945 | 22.48765 |
| S9 | Eastern Waters | 114.289444 | 22.343056 |
| S10 | Eastern Waters | 114.296389 | 22.286111 |
| S11 | Eastern Waters | 114.3325 | 22.276111 |
| S12 | Eastern Waters | 114.343333 | 22.249722 |
| S13 | Eastern Waters | 114.304167 | 22.289444 |
| S14 | Eastern Waters | 114.379722 | 22.265 |
| S15 | Eastern Waters | 114.394167 | 22.355 |
| S16 | Southern Waters | 114.322222 | 22.366667 |
| S17 | Southern Waters | 114.257222 | 22.234167 |
| S18 | Southern Waters | 114.256389 | 22.205 |
| S19 | Southern Waters | 114.286389 | 22.155278 |
| S20 | Southern Waters | 114.175556 | 22.1875 |
| S21 | Southern Waters | 114.189444 | 22.227222 |
| S22 | Southern Waters | 114.130833 | 22.228056 |
| S23 | Southern Waters | 114.073333 | 22.278333 |
| S24 | Southern Waters | 114.055 | 22.353333 |
| S25 | Southern Waters | 114.119722 | 22.331667 |
| S26 | Southern Waters | 114.117222 | 22.311667 |
| S27 | Southern Waters | 114.1525 | 22.299167 |
| S28 | Southern Waters | 114.009722 | 22.238889 |
| S29 | Southern Waters | 114.002778 | 22.188056 |
| S30 | Southern Waters | 113.939444 | 22.185556 |
| S31 | Southern Waters | 113.881389 | 22.166944 |

| S32 | Western Waters | 113.827983 | 22.224317 |
|-----|----------------|------------|-----------|
| S33 | Western Waters | 113.975 | 22.323056 |
| S34 | Western Waters | 113.875 | 22.289722 |
| S35 | Western Waters | 113.919722 | 22.431944 |
| S36 | Western Waters | 113.905278 | 22.361111 |

2. The detailed information of sampling locations of coastal plastisphere.

| Sample ID | Group | Description | Region | Longitude | Latitude |
|----------------|---------|----------------|-----------|-----------|----------|
| EnvHKSumMc1 | Ambient | Marine Culture | Hong Kong | 114.32278 | 22.36583 |
| EnvHKSumMR1 | Ambient | Marine Reserve | Hong Kong | 114.27630 | 22.25900 |
| EnvHKSumMR2 | Ambient | Marine Reserve | Hong Kong | 114.27630 | 22.25900 |
| EnvHKSumMR3 | Ambient | Marine Reserve | Hong Kong | 114.27630 | 22.25900 |
| EnvHKSumMR4 | Ambient | Marine Reserve | Hong Kong | 114.27630 | 22.25900 |
| EnvHKSumPort1 | Ambient | Port | Hong Kong | 114.27446 | 22.38152 |
| EnvHKSumWWTP2 | Ambient | WWTP | Hong Kong | 114.11727 | 22.31137 |
| EnvHKWinBeach1 | Ambient | Beach | Hong Kong | 114.18930 | 22.23140 |
| EnvHKWinBeach2 | Ambient | Beach | Hong Kong | 114.18930 | 22.23140 |
| EnvHKWinBeach3 | Ambient | Beach | Hong Kong | 114.18930 | 22.23140 |
| EnvHKWinEs1 | Ambient | Estuary | Hong Kong | 114.21694 | 22.40194 |
| EnvHKWinEs2 | Ambient | Estuary | Hong Kong | 114.21694 | 22.40194 |
| EnvHKWinEs4 | Ambient | Estuary | Hong Kong | 114.21694 | 22.40194 |
| EnvHKWinEs5 | Ambient | Estuary | Hong Kong | 114.21694 | 22.40194 |
| EnvHKWinMc1 | Ambient | Marine Culture | Hong Kong | 113.98209 | 22.46892 |
| EnvHKWinMc10 | Ambient | Marine Culture | Hong Kong | 114.32278 | 22.36583 |
| EnvHKWinMc11 | Ambient | Marine Culture | Hong Kong | 114.13045 | 22.22385 |
| EnvHKWinMc12 | Ambient | Marine Culture | Hong Kong | 114.13045 | 22.22385 |
| EnvHKWinMc13 | Ambient | Marine Culture | Hong Kong | 114.13045 | 22.22385 |
| EnvHKWinMc15 | Ambient | Marine Culture | Hong Kong | 114.13045 | 22.22385 |
| EnvHKWinMc2 | Ambient | Marine Culture | Hong Kong | 113.98209 | 22.46892 |
| EnvHKWinMc3 | Ambient | Marine Culture | Hong Kong | 113.98209 | 22.46892 |
| EnvHKWinMc4 | Ambient | Marine Culture | Hong Kong | 113.98209 | 22.46892 |
| EnvHKWinMc5 | Ambient | Marine Culture | Hong Kong | 113.98209 | 22.46892 |
| EnvHKWinMc6 | Ambient | Marine Culture | Hong Kong | 114.32278 | 22.36583 |
| EnvHKWinMc7 | Ambient | Marine Culture | Hong Kong | 114.32278 | 22.36583 |
| EnvHKWinMc8 | Ambient | Marine Culture | Hong Kong | 114.32278 | 22.36583 |
| EnvHKWinMc9 | Ambient | Marine Culture | Hong Kong | 114.32278 | 22.36583 |
| EnvHKWinMR1 | Ambient | Marine Reserve | Hong Kong | 114.27632 | 22.25902 |
| EnvHKWinPort1 | Ambient | Port | Hong Kong | 114.27446 | 22.38152 |
| EnvHKWinPort2 | Ambient | Port | Hong Kong | 114.27446 | 22.38152 |
| EnvHKWinPort3 | Ambient | Port | Hong Kong | 114.27446 | 22.38152 |
| EnvHKWinPort4 | Ambient | Port | Hong Kong | 114.27446 | 22.38152 |
| EnvHKWinPort5 | Ambient | Port | Hong Kong | 114.27446 | 22.38152 |
| | | | | | |

| E HIZW' WWED1 | A 1 ' 4 | WWTD | 11 17 | 114 11727 | 22 21127 |
|----------------|---------|----------------|-----------|-----------|----------|
| EnvHKWinWWTP1 | Ambient | WWTP | Hong Kong | 114.11727 | 22.31137 |
| EnvHKWinWWTP2 | Ambient | WWTP | Hong Kong | 114.11727 | 22.31137 |
| EnvHKWinWWTP3 | Ambient | WWTP | Hong Kong | 114.11727 | 22.31137 |
| EnvHKWinWWTP4 | Ambient | WWTP | Hong Kong | 114.11727 | 22.31137 |
| EnvQDSumBeach1 | Ambient | Beach | Hong Kong | 120.49433 | 36.09085 |
| EnvQDSumBeach2 | Ambient | Beach | Hong Kong | 120.49433 | 36.09085 |
| EnvQDSumBeach4 | Ambient | Beach | Hong Kong | 120.49433 | 36.09085 |
| EnvQDSumBeach5 | Ambient | Beach | Hong Kong | 120.49433 | 36.09085 |
| EnvQDSumEs1 | Ambient | Estuary | Hong Kong | 120.31518 | 36.25501 |
| EnvQDSumEs2 | Ambient | Estuary | Hong Kong | 120.31518 | 36.25501 |
| EnvQDSumEs3 | Ambient | Estuary | Hong Kong | 120.31518 | 36.25501 |
| EnvQDSumMc1 | Ambient | Marine Culture | Hong Kong | 120.23318 | 36.19174 |
| EnvQDSumMc2 | Ambient | Marine Culture | Hong Kong | 120.23318 | 36.19174 |
| EnvQDSumMc4 | Ambient | Marine Culture | Hong Kong | 120.23318 | 36.19174 |
| EnvQDSumMc5 | Ambient | Marine Culture | Hong Kong | 120.23318 | 36.19174 |
| EnvQDSumMR1 | Ambient | Marine Reserve | Hong Kong | 120.17625 | 35.88025 |
| EnvQDSumMR2 | Ambient | Marine Reserve | Hong Kong | 120.17625 | 35.88025 |
| EnvQDSumMR3 | Ambient | Marine Reserve | Hong Kong | 120.17625 | 35.88025 |
| EnvQDSumPort1 | Ambient | Port | Hong Kong | 120.30023 | 36.07129 |
| EnvQDSumPort2 | Ambient | Port | Hong Kong | 120.30023 | 36.07129 |
| EnvQDSumPort3 | Ambient | Port | Hong Kong | 120.30023 | 36.07129 |
| EnvQDSumPort4 | Ambient | Port | Hong Kong | 120.30023 | 36.07129 |
| EnvQDSumPort5 | Ambient | Port | Hong Kong | 120.30023 | 36.07129 |
| EnvQDSumWWTP1 | Ambient | WWTP | Hong Kong | 120.34792 | 36.15289 |
| EnvQDSumWWTP2 | Ambient | WWTP | Hong Kong | 120.34792 | 36.15289 |
| EnvQDSumWWTP3 | Ambient | WWTP | Hong Kong | 120.34792 | 36.15289 |
| EnvQDSumWWTP4 | Ambient | WWTP | Hong Kong | 120.34792 | 36.15289 |
| EnvQDSumWWTP5 | Ambient | WWTP | Hong Kong | 120.34792 | 36.15289 |
| EnvQDWinBeach1 | Ambient | Beach | Hong Kong | 120.49433 | 36.09085 |
| EnvQDWinBeach2 | Ambient | Beach | Hong Kong | 120.49433 | 36.09085 |
| EnvQDWinBeach3 | Ambient | Beach | Hong Kong | 120.49433 | 36.09085 |
| EnvQDWinBeach4 | Ambient | Beach | Hong Kong | 120.49433 | 36.09085 |
| EnvQDWinBeach5 | Ambient | Beach | Hong Kong | 120.49433 | 36.09085 |
| EnvQDWinEs1 | Ambient | Estuary | Hong Kong | 120.31518 | 36.25501 |
| EnvQDWinEs2 | Ambient | Estuary | Hong Kong | 120.31518 | 36.25501 |
| EnvQDWinEs3 | Ambient | Estuary | Hong Kong | 120.31518 | 36.25501 |
| EnvQDWinEs4 | Ambient | Estuary | Hong Kong | 120.31518 | 36.25501 |
| | | | | | |

| EnvQDWinEs5 | Ambient | Estuary | Hong Kong | 120.31518 | 36.25501 |
|----------------|--------------|----------------|-----------|-----------|----------|
| EnvQDWinMc1 | Ambient | Marine Culture | Hong Kong | 120.23318 | 36.19174 |
| EnvQDWinMc2 | Ambient | Marine Culture | Hong Kong | 120.23318 | 36.19174 |
| EnvQDWinMc3 | Ambient | Marine Culture | Hong Kong | 120.23318 | 36.19174 |
| EnvQDWinMc4 | Ambient | Marine Culture | Hong Kong | 120.23318 | 36.19174 |
| EnvQDWinMc5 | Ambient | Marine Culture | Hong Kong | 120.23318 | 36.19174 |
| EnvQDWinMR1 | Ambient | Marine Reserve | Hong Kong | 120.17625 | 35.88025 |
| EnvQDWinMR2 | Ambient | Marine Reserve | Hong Kong | 120.17625 | 35.88025 |
| EnvQDWinMR3 | Ambient | Marine Reserve | Hong Kong | 120.17625 | 35.88025 |
| EnvQDWinMR4 | Ambient | Marine Reserve | Hong Kong | 120.17625 | 35.88025 |
| EnvQDWinMR5 | Ambient | Marine Reserve | Hong Kong | 120.17625 | 35.88025 |
| EnvQDWinPort1 | Ambient | Port | Hong Kong | 120.30023 | 36.07129 |
| EnvQDWinPort2 | Ambient | Port | Hong Kong | 120.30023 | 36.07129 |
| EnvQDWinPort3 | Ambient | Port | Hong Kong | 120.30023 | 36.07129 |
| EnvQDWinPort4 | Ambient | Port | Hong Kong | 120.30023 | 36.07129 |
| EnvQDWinPort5 | Ambient | Port | Hong Kong | 120.30023 | 36.07129 |
| EnvQDWinWWTP1 | Ambient | WWTP | Hong Kong | 120.34792 | 36.15289 |
| EnvQDWinWWTP2 | Ambient | WWTP | Hong Kong | 120.34792 | 36.15289 |
| EnvQDWinWWTP3 | Ambient | WWTP | Hong Kong | 120.34792 | 36.15289 |
| EnvQDWinWWTP4 | Ambient | WWTP | Hong Kong | 120.34792 | 36.15289 |
| EnvQDWinWWTP5 | Ambient | WWTP | Hong Kong | 120.34792 | 36.15289 |
| PlaHKSumMc1 | Plastisphere | Marine Culture | Hong Kong | 114.32278 | 22.36583 |
| PlaHKSumMR1 | Plastisphere | Marine Reserve | Hong Kong | 114.27630 | 22.25900 |
| PlaHKSumMR2 | Plastisphere | Marine Reserve | Hong Kong | 114.27630 | 22.25900 |
| PlaHKSumMR3 | Plastisphere | Marine Reserve | Hong Kong | 114.27630 | 22.25900 |
| PlaHKSumMR4 | Plastisphere | Marine Reserve | Hong Kong | 114.27630 | 22.25900 |
| PlaHKSumPort1 | Plastisphere | Port | Hong Kong | 114.27446 | 22.38152 |
| PlaHKSumWWTP2 | Plastisphere | WWTP | Hong Kong | 114.11727 | 22.31137 |
| PlaHKWinBeach1 | Plastisphere | Beach | Hong Kong | 114.18930 | 22.23140 |
| PlaHKWinBeach2 | Plastisphere | Beach | Hong Kong | 114.18930 | 22.23140 |
| PlaHKWinBeach3 | Plastisphere | Beach | Hong Kong | 114.18930 | 22.23140 |
| PlaHKWinEs1 | Plastisphere | Estuary | Hong Kong | 114.21694 | 22.40194 |
| PlaHKWinEs2 | Plastisphere | Estuary | Hong Kong | 114.21694 | 22.40194 |
| PlaHKWinEs4 | Plastisphere | Estuary | Hong Kong | 114.21694 | 22.40194 |
| PlaHKWinEs5 | Plastisphere | Estuary | Hong Kong | 114.21694 | 22.40194 |
| PlaHKWinMc1 | Plastisphere | Marine Culture | Hong Kong | 113.98209 | 22.46892 |
| PlaHKWinMc10 | Plastisphere | Marine Culture | Hong Kong | 114.32278 | 22.36583 |

| PlaHKWinMc11 | Plastisphere | Marine Culture | Hong Kong | 114.13045 | 22.22385 |
|----------------|--------------|----------------|-----------|-----------|----------|
| PlaHKWinMc12 | Plastisphere | Marine Culture | Hong Kong | 114.13045 | 22.22385 |
| PlaHKWinMc13 | Plastisphere | Marine Culture | Hong Kong | 114.13045 | 22.22385 |
| PlaHKWinMc15 | Plastisphere | Marine Culture | Hong Kong | 114.13045 | 22.22385 |
| PlaHKWinMc2 | Plastisphere | Marine Culture | Hong Kong | 113.98209 | 22.46892 |
| PlaHKWinMc3 | Plastisphere | Marine Culture | Hong Kong | 113.98209 | 22.46892 |
| PlaHKWinMc4 | Plastisphere | Marine Culture | Hong Kong | 113.98209 | 22.46892 |
| PlaHKWinMc5 | Plastisphere | Marine Culture | Hong Kong | 113.98209 | 22.46892 |
| PlaHKWinMc6 | Plastisphere | Marine Culture | Hong Kong | 114.32278 | 22.36583 |
| PlaHKWinMc7 | Plastisphere | Marine Culture | Hong Kong | 114.32278 | 22.36583 |
| PlaHKWinMc8 | Plastisphere | Marine Culture | Hong Kong | 114.32278 | 22.36583 |
| PlaHKWinMc9 | Plastisphere | Marine Culture | Hong Kong | 114.32278 | 22.36583 |
| PlaHKWinMR1 | Plastisphere | Marine Reserve | Hong Kong | 114.27632 | 22.25902 |
| PlaHKWinPort1 | Plastisphere | Port | Hong Kong | 114.27446 | 22.38152 |
| PlaHKWinPort2 | Plastisphere | Port | Hong Kong | 114.27446 | 22.38152 |
| PlaHKWinPort3 | Plastisphere | Port | Hong Kong | 114.27446 | 22.38152 |
| PlaHKWinPort4 | Plastisphere | Port | Hong Kong | 114.27446 | 22.38152 |
| PlaHKWinPort5 | Plastisphere | Port | Hong Kong | 114.27446 | 22.38152 |
| PlaHKWinWWTP1 | Plastisphere | WWTP | Hong Kong | 114.11727 | 22.31137 |
| PlaHKWinWWTP2 | Plastisphere | WWTP | Hong Kong | 114.11727 | 22.31137 |
| PlaHKWinWWTP3 | Plastisphere | WWTP | Hong Kong | 114.11727 | 22.31137 |
| PlaHKWinWWTP4 | Plastisphere | WWTP | Hong Kong | 114.11727 | 22.31137 |
| PlaQDSumBeach1 | Plastisphere | Beach | Qingdao | 120.49433 | 36.09085 |
| PlaQDSumBeach2 | Plastisphere | Beach | Qingdao | 120.49433 | 36.09085 |
| PlaQDSumBeach4 | Plastisphere | Beach | Qingdao | 120.49433 | 36.09085 |
| PlaQDSumBeach5 | Plastisphere | Beach | Qingdao | 120.49433 | 36.09085 |
| PlaQDSumEs1 | Plastisphere | Estuary | Qingdao | 120.31518 | 36.25501 |
| PlaQDSumEs2 | Plastisphere | Estuary | Qingdao | 120.31518 | 36.25501 |
| PlaQDSumEs3 | Plastisphere | Estuary | Qingdao | 120.31518 | 36.25501 |
| PlaQDSumMc1 | Plastisphere | Marine Culture | Qingdao | 120.23318 | 36.19174 |
| PlaQDSumMc2 | Plastisphere | Marine Culture | Qingdao | 120.23318 | 36.19174 |
| PlaQDSumMc4 | Plastisphere | Marine Culture | Qingdao | 120.23318 | 36.19174 |
| PlaQDSumMc5 | Plastisphere | Marine Culture | Qingdao | 120.23318 | 36.19174 |
| PlaQDSumMR1 | Plastisphere | Marine Reserve | Qingdao | 120.17625 | 35.88025 |
| PlaQDSumMR2 | Plastisphere | Marine Reserve | Qingdao | 120.17625 | 35.88025 |
| PlaQDSumMR3 | Plastisphere | Marine Reserve | Qingdao | 120.17625 | 35.88025 |
| PlaQDSumPort1 | Plastisphere | Port | Qingdao | 120.30023 | 36.07129 |

| PlaQDSumPort2 | Plastisphere | Port | Qingdao | 120.30023 | 36.07129 |
|----------------|--------------|----------------|---------|-----------|----------|
| PlaQDSumPort3 | Plastisphere | Port | Qingdao | 120.30023 | 36.07129 |
| PlaQDSumPort4 | Plastisphere | Port | Qingdao | 120.30023 | 36.07129 |
| PlaQDSumPort5 | Plastisphere | Port | Qingdao | 120.30023 | 36.07129 |
| PlaQDSumWWTP1 | Plastisphere | WWTP | Qingdao | 120.34792 | 36.15289 |
| PlaQDSumWWTP2 | Plastisphere | WWTP | Qingdao | 120.34792 | 36.15289 |
| PlaQDSumWWTP3 | Plastisphere | WWTP | Qingdao | 120.34792 | 36.15289 |
| PlaQDSumWWTP4 | Plastisphere | WWTP | Qingdao | 120.34792 | 36.15289 |
| PlaQDSumWWTP5 | Plastisphere | WWTP | Qingdao | 120.34792 | 36.15289 |
| PlaQDWinBeach1 | Plastisphere | Beach | Qingdao | 120.49433 | 36.09085 |
| PlaQDWinBeach2 | Plastisphere | Beach | Qingdao | 120.49433 | 36.09085 |
| PlaQDWinBeach3 | Plastisphere | Beach | Qingdao | 120.49433 | 36.09085 |
| PlaQDWinBeach4 | Plastisphere | Beach | Qingdao | 120.49433 | 36.09085 |
| PlaQDWinBeach5 | Plastisphere | Beach | Qingdao | 120.49433 | 36.09085 |
| PlaQDWinEs1 | Plastisphere | Estuary | Qingdao | 120.31518 | 36.25501 |
| PlaQDWinEs2 | Plastisphere | Estuary | Qingdao | 120.31518 | 36.25501 |
| PlaQDWinEs3 | Plastisphere | Estuary | Qingdao | 120.31518 | 36.25501 |
| PlaQDWinEs4 | Plastisphere | Estuary | Qingdao | 120.31518 | 36.25501 |
| PlaQDWinEs5 | Plastisphere | Estuary | Qingdao | 120.31518 | 36.25501 |
| PlaQDWinMc1 | Plastisphere | Marine Culture | Qingdao | 120.23318 | 36.19174 |
| PlaQDWinMc2 | Plastisphere | Marine Culture | Qingdao | 120.23318 | 36.19174 |
| PlaQDWinMc3 | Plastisphere | Marine Culture | Qingdao | 120.23318 | 36.19174 |
| PlaQDWinMc4 | Plastisphere | Marine Culture | Qingdao | 120.23318 | 36.19174 |
| PlaQDWinMc5 | Plastisphere | Marine Culture | Qingdao | 120.23318 | 36.19174 |
| PlaQDWinMR1 | Plastisphere | Marine Reserve | Qingdao | 120.17625 | 35.88025 |
| PlaQDWinMR2 | Plastisphere | Marine Reserve | Qingdao | 120.17625 | 35.88025 |
| PlaQDWinMR3 | Plastisphere | Marine Reserve | Qingdao | 120.17625 | 35.88025 |
| PlaQDWinMR4 | Plastisphere | Marine Reserve | Qingdao | 120.17625 | 35.88025 |
| PlaQDWinMR5 | Plastisphere | Marine Reserve | Qingdao | 120.17625 | 35.88025 |
| PlaQDWinPort1 | Plastisphere | Port | Qingdao | 120.30023 | 36.07129 |
| PlaQDWinPort2 | Plastisphere | Port | Qingdao | 120.30023 | 36.07129 |
| PlaQDWinPort3 | Plastisphere | Port | Qingdao | 120.30023 | 36.07129 |
| PlaQDWinPort4 | Plastisphere | Port | Qingdao | 120.30023 | 36.07129 |
| PlaQDWinPort5 | Plastisphere | Port | Qingdao | 120.30023 | 36.07129 |
| PlaQDWinWWTP1 | Plastisphere | WWTP | Qingdao | 120.34792 | 36.15289 |
| PlaQDWinWWTP2 | Plastisphere | WWTP | Qingdao | 120.34792 | 36.15289 |
| PlaQDWinWWTP3 | Plastisphere | WWTP | Qingdao | 120.34792 | 36.15289 |

| PlaQDWinWWTP4 | Plastisphere | WWTP | Qingdao | 120.34792 | 36.15289 |
|---------------|--------------|------|---------|-----------|----------|
| PlaQDWinWWTP5 | Plastisphere | WWTP | Qingdao | 120.34792 | 36.15289 |

Appendix 3

This section provides a database of the experimental results of the thesis.

1 In situ environmental parameters and nutrients in Hong Kong coastal seawater

| Sample ID | In-situ para | Nutrients (μg·L ⁻¹) | | | | | | | |
|-----------|--------------|---------------------------------|------|-----------------------|-------------------|-------------------|------------------------------|--------------------------------|--------------------------------|
| | Temp (°C) | Salinity | pН | DO | NO ₃ - | NO ₂ - | NH ₄ ⁺ | PO ₄ ³ - | SiO ₃ ²⁻ |
| | | | | (mg L ⁻¹) | | | | | |
| Wet_S1 | 30.2 | 17.5 | 8.22 | 5.52 | 16.7 | 7.47 | 42.1 | 31.5 | 1487 |
| Wet_S2 | 33.1 | 24.9 | 8.50 | 2.39 | 7.35 | 1.52 | 55.8 | 48.2 | 605 |
| Wet_S3 | 32.9 | 27.2 | 8.35 | 8.75 | 6.96 | 1.33 | 31.6 | 45.4 | 263 |
| Wet_S4 | 31.3 | 26.2 | 8.38 | 5.55 | 7.89 | 1.94 | 41.7 | 10.8 | 478 |
| Wet_S5 | 30.7 | 31.9 | 8.26 | 5.57 | 7.33 | 1.60 | 33.2 | 8.26 | 189 |
| Wet_S6 | 30.8 | 32.1 | 8.18 | 8.21 | 7.05 | 1.23 | 26.0 | 5.16 | 206 |
| Wet_S7 | 30.9 | 30.2 | 8.23 | 5.83 | 7.15 | 1.46 | 48.5 | 18.3 | 157 |
| Wet_S8 | 30.4 | 30.5 | 8.25 | 5.55 | 7.33 | 1.39 | 40.2 | 8.18 | 114 |
| Wet_S9 | 31.6 | 31.9 | 8.19 | 9.50 | 6.59 | 1.06 | 35.9 | 39.1 | 228 |
| Wet_S10 | 30.4 | 32.6 | 8.24 | 9.60 | 6.75 | 1.17 | 55.5 | 17.3 | 121 |
| Wet_S11 | 30.4 | 32.5 | 8.29 | 9.70 | 6.76 | 1.38 | 42.6 | 11.2 | 126 |
| Wet_S12 | 30.2 | 31.5 | 8.24 | 10.2 | 6.58 | 0 | 47.4 | 7.05 | 100 |
| Wet_S13 | 30.3 | 32.6 | 8.25 | 10.0 | 6.66 | 1.18 | 57.1 | 10.4 | 89.2 |
| Wet_S14 | 30.4 | 31.6 | 8.25 | 10.1 | 6.79 | 1.85 | 47.2 | 11.8 | 81.3 |
| Wet_S15 | 30.7 | 32.0 | 8.27 | 9.80 | 6.80 | 1.52 | 31.1 | 9.03 | 95.5 |
| Wet_S16 | 31.5 | 32.4 | 8.06 | 8.10 | 6.67 | 1.17 | 39.9 | 16.0 | 189 |
| Wet_S17 | 28.1 | 31.3 | 8.08 | 5.56 | 14.3 | 10.6 | 62.0 | 11.8 | 129 |
| Wet_S18 | 27.9 | 31.1 | 8.10 | 5.83 | 15.1 | 11.4 | 52.3 | 19.9 | 127 |
| Wet_S19 | 28.9 | 30.5 | 8.17 | 6.48 | 11.0 | 6.33 | 41.2 | 17.8 | 100 |
| Wet_S20 | 29.0 | 29.6 | 8.15 | 6.39 | 17.9 | 14.3 | 42.9 | 19.5 | 114 |
| Wet_S21 | 28.8 | 29.7 | 8.15 | 6.65 | 17.3 | 12.5 | 38.6 | 19.0 | 110 |
| Wet_S22 | 28.4 | 28.5 | 8.06 | 5.07 | 43.1 | 39.0 | 66.1 | 17.9 | 250 |
| Wet_S23 | 29.0 | 27.3 | 8.14 | 6.00 | 47.4 | 43.1 | 75.8 | 34.9 | 311 |
| Wet_S24 | 29.3 | 24.5 | 7.96 | 4.78 | 86.4 | 85.2 | 79.2 | 64.6 | 642 |
| Wet_S25 | 28.1 | 28.2 | 7.95 | 3.74 | 46.7 | 42.9 | 130 | 49.6 | 447 |
| Wet_S26 | 28.1 | 27.3 | 7.74 | 3.27 | 66.7 | 64.6 | 2101 | 103.0 | 635 |
| Wet_S27 | 28.6 | 28.2 | 7.97 | 5.96 | 39.9 | 36.4 | 120 | 17.1 | 380 |
| Wet_S28 | 29.3 | 28.0 | 8.15 | 5.95 | 113 | 115 | 12.7 | 77.5 | 864 |

| Wet_S29 | 29.6 | 27.9 | 8.31 | 7.80 | 34.9 | 29.0 | 99.6 | 42.7 | 306 |
|---------|------|------|------|------|------|------|------|------|------|
| Wet_S30 | 30.1 | 28.2 | 8.31 | 7.17 | 23.9 | 20.1 | 66.9 | 47.8 | 207 |
| Wet_S31 | 30.1 | 27.7 | 8.14 | 6.67 | 18.0 | 13.0 | 31.1 | 24.9 | 143 |
| Wet_S32 | 30.8 | 21.3 | 7.94 | 5.48 | 27.1 | 22.0 | 77.7 | 0 | 170 |
| Wet_S33 | 30.2 | 22.1 | 7.90 | 5.74 | 168 | 175 | 2.43 | 86.7 | 1181 |
| Wet_S34 | 30.4 | 23.4 | 7.96 | 5.48 | 122 | 125 | 27.6 | 47.3 | 660 |
| Wet_S35 | 30.6 | 20.3 | 7.81 | 5.29 | 113 | 114 | 0 | 34.9 | 681 |
| Wet_S36 | 31.0 | 18.8 | 7.88 | 6.52 | 149 | 153 | 53.4 | 91.2 | 992 |
| Dry_S1 | 19.7 | 31.8 | 7.76 | 6.74 | 8.12 | 7.33 | 85.7 | 4.80 | 258 |
| Dry_S2 | 20.2 | 31.9 | 7.82 | 6.18 | 3.02 | 2.40 | 34.1 | 3.10 | 99 |
| Dry_S3 | 19.6 | 32.0 | 7.83 | 5.91 | 2.93 | 1.69 | 13.1 | 2.99 | 105 |
| Dry_S4 | 19.7 | 32.1 | 7.67 | 6.57 | 2.30 | 1.42 | 13.5 | 6.36 | 139 |
| Dry_S5 | 19.1 | 32.6 | 7.90 | 6.54 | 2.24 | 1.56 | 11.7 | 8.86 | 123 |
| Dry_S6 | 20.1 | 32.8 | 8.07 | 5.92 | 1.84 | 0.95 | 20.5 | 4.57 | 87.6 |
| Dry_S7 | 19.6 | 32.9 | 7.98 | 5.86 | 1.72 | 0.42 | 28.9 | 5.62 | 63.4 |
| Dry_S8 | 18.9 | 33.0 | 8.06 | 5.74 | 0 | 0.93 | 15.8 | 4.20 | 63.3 |
| Dry_S9 | 18.8 | 33.3 | 7.92 | 5.47 | 2.33 | 0.93 | 45.2 | 2.89 | 74.6 |
| Dry_S10 | 20.0 | 33.6 | 7.89 | 5.78 | 2.53 | 1.44 | 34.9 | 7.61 | 89.6 |
| Dry_S11 | 19.5 | 33.6 | 7.88 | 5.72 | 5.65 | 4.28 | 51.9 | 8.63 | 113 |
| Dry_S12 | 19.5 | 33.9 | 7.73 | 5.54 | 3.82 | 3.02 | 26.2 | 4.91 | 82.6 |
| Dry_S13 | 19.5 | 33.7 | 7.88 | 5.87 | 4.97 | 3.79 | 64.5 | 7.70 | 108 |
| Dry_S14 | 19.2 | 33.9 | 7.72 | 5.75 | 7.29 | 6.05 | 50.7 | 7.67 | 114 |
| Dry_S15 | 19.1 | 33.6 | 7.83 | 5.84 | 6.42 | 5.42 | 52.5 | 7.53 | 125 |
| Dry_S16 | 19.3 | 33.3 | 7.79 | 5.99 | 2.11 | 1.96 | 52.3 | 5.95 | 96.1 |
| Dry_S17 | 18.3 | 33.1 | 7.86 | 5.90 | 6.28 | 5.69 | 74.2 | 6.51 | 62.4 |
| Dry_S18 | 18.2 | 33.4 | 7.19 | 5.87 | 8.62 | 7.90 | 85.3 | 12.0 | 129 |
| Dry_S19 | 20.4 | 34.0 | 7.78 | 6.05 | 3.05 | 2.32 | 52.8 | 5.43 | 92.6 |
| Dry_S20 | 18.2 | 33.3 | 7.68 | 6.12 | 4.84 | 4.21 | 43.1 | 7.84 | 113 |
| Dry_S21 | 18.5 | 33.0 | 7.85 | 6.74 | 2.66 | 2.01 | 46.7 | 5.90 | 75.0 |
| Dry_S22 | 19.2 | 32.7 | 7.82 | 6.18 | 2.79 | 2.53 | 125 | 10.2 | 69.7 |
| Dry_S23 | 18.4 | 33.1 | 7.72 | 5.91 | 4.99 | 4.26 | 144 | 7.87 | 79.2 |
| Dry_S24 | 18.7 | 32.7 | 8.06 | 6.57 | 4.76 | 4.42 | 141 | 16.4 | 75.5 |
| Dry_S25 | 18.8 | 32.5 | 7.91 | 6.54 | 5.61 | 4.98 | 226 | 13.1 | 74.1 |
| Dry_S26 | 18.6 | 33.0 | 7.87 | 5.92 | 7.22 | 6.12 | 127 | 6.36 | 67.6 |
| Dry_S27 | 18.9 | 31.4 | 7.80 | 5.86 | 9.53 | 8.85 | 1212 | 12.7 | 165 |
| Dry_S28 | 18.8 | 32.6 | 7.87 | 5.74 | 6.55 | 5.59 | 135 | 3.81 | 111 |
| Dry_S29 | 18.5 | 32.2 | 7.78 | 5.47 | 4.22 | 3.50 | 113 | 0.12 | 82.4 |
| | | | | | | | | | |

| Dry_S30 | 19.1 | 33.0 | 7.88 | 5.78 | 3.78 | 3.38 | 72.0 | 2.45 | 75.0 |
|---------|------|------|------|------|------|------|------|------|------|
| Dry_S31 | 19.1 | 33.2 | 7.86 | 5.72 | 4.67 | 3.92 | 72.7 | 1.91 | 81.8 |
| Dry_S32 | 18.9 | 32.9 | 7.83 | 5.54 | 5.17 | 4.72 | 72.7 | 3.18 | 81.1 |
| Dry_S33 | 19.6 | 30.8 | 7.95 | 5.87 | 9.63 | 8.58 | 152 | 10.2 | 52.1 |
| Dry_S34 | 19.1 | 32.9 | 7.91 | 5.75 | 4.27 | 3.50 | 45.2 | 1.28 | 78.2 |
| Dry_S35 | 19.7 | 30.6 | 7.85 | 5.84 | 7.69 | 6.80 | 108 | 9.61 | 61.9 |
| Dry_S36 | 19.4 | 30.4 | 8.04 | 5.99 | 10.8 | 9.63 | 135 | 8.43 | 76.0 |

2 Concentrations of LATs in dissolved seawater and SPM in sampling location.

| Sample | | Wet season (ng L ⁻¹) | | | | | | | Dry season (ng L ⁻¹) | | | | | | | |
|-----------------|-------|----------------------------------|----------|--|-------|---|------|--|----------------------------------|---------|----------|-------|-------|------|------|---------------------|
| Sample location | | Dissolv | ed phase | ; | | SP | M | | | Dissolv | ed phase | | | S | PM | |
| location | PTX-2 | OA | GYM | DTX-1 | PTX-2 | OA | GYM | DTX-1 | PTX-2 | OA | GYM | DTX-1 | PTX-2 | OA | GYM | DTX-1 |
| S1 | 0.85 | 0.68 | 0.18 | <lod< td=""><td>0.16</td><td><lod< td=""><td>0.01</td><td><lod< td=""><td>0.31</td><td>1.23</td><td>1.43</td><td>0.03</td><td>0.18</td><td>0.00</td><td>0.03</td><td><lod< td=""></lod<></td></lod<></td></lod<></td></lod<> | 0.16 | <lod< td=""><td>0.01</td><td><lod< td=""><td>0.31</td><td>1.23</td><td>1.43</td><td>0.03</td><td>0.18</td><td>0.00</td><td>0.03</td><td><lod< td=""></lod<></td></lod<></td></lod<> | 0.01 | <lod< td=""><td>0.31</td><td>1.23</td><td>1.43</td><td>0.03</td><td>0.18</td><td>0.00</td><td>0.03</td><td><lod< td=""></lod<></td></lod<> | 0.31 | 1.23 | 1.43 | 0.03 | 0.18 | 0.00 | 0.03 | <lod< td=""></lod<> |
| S2 | 0.31 | 1.12 | 0.31 | <lod< td=""><td>0.06</td><td><lod< td=""><td>0.01</td><td><lod< td=""><td>0.69</td><td>1.94</td><td>1.76</td><td>0.04</td><td>0.22</td><td>0.00</td><td>0.05</td><td><lod< td=""></lod<></td></lod<></td></lod<></td></lod<> | 0.06 | <lod< td=""><td>0.01</td><td><lod< td=""><td>0.69</td><td>1.94</td><td>1.76</td><td>0.04</td><td>0.22</td><td>0.00</td><td>0.05</td><td><lod< td=""></lod<></td></lod<></td></lod<> | 0.01 | <lod< td=""><td>0.69</td><td>1.94</td><td>1.76</td><td>0.04</td><td>0.22</td><td>0.00</td><td>0.05</td><td><lod< td=""></lod<></td></lod<> | 0.69 | 1.94 | 1.76 | 0.04 | 0.22 | 0.00 | 0.05 | <lod< td=""></lod<> |
| S3 | 1.38 | 0.78 | 0.18 | <lod< td=""><td>0.07</td><td><lod< td=""><td>0.01</td><td><lod< td=""><td>1.01</td><td>1.10</td><td>0.98</td><td>0.05</td><td>0.36</td><td>0.00</td><td>0.07</td><td><lod< td=""></lod<></td></lod<></td></lod<></td></lod<> | 0.07 | <lod< td=""><td>0.01</td><td><lod< td=""><td>1.01</td><td>1.10</td><td>0.98</td><td>0.05</td><td>0.36</td><td>0.00</td><td>0.07</td><td><lod< td=""></lod<></td></lod<></td></lod<> | 0.01 | <lod< td=""><td>1.01</td><td>1.10</td><td>0.98</td><td>0.05</td><td>0.36</td><td>0.00</td><td>0.07</td><td><lod< td=""></lod<></td></lod<> | 1.01 | 1.10 | 0.98 | 0.05 | 0.36 | 0.00 | 0.07 | <lod< td=""></lod<> |
| S4 | 1.39 | 1.02 | 0.29 | <lod< td=""><td>0.04</td><td><lod< td=""><td>0.02</td><td><lod< td=""><td>1.08</td><td>1.89</td><td>1.68</td><td>0.06</td><td>0.29</td><td>0.00</td><td>0.02</td><td><lod< td=""></lod<></td></lod<></td></lod<></td></lod<> | 0.04 | <lod< td=""><td>0.02</td><td><lod< td=""><td>1.08</td><td>1.89</td><td>1.68</td><td>0.06</td><td>0.29</td><td>0.00</td><td>0.02</td><td><lod< td=""></lod<></td></lod<></td></lod<> | 0.02 | <lod< td=""><td>1.08</td><td>1.89</td><td>1.68</td><td>0.06</td><td>0.29</td><td>0.00</td><td>0.02</td><td><lod< td=""></lod<></td></lod<> | 1.08 | 1.89 | 1.68 | 0.06 | 0.29 | 0.00 | 0.02 | <lod< td=""></lod<> |
| S5 | 1.97 | 1.03 | 0.16 | <lod< td=""><td>0.01</td><td><lod< td=""><td>0.00</td><td><lod< td=""><td>1.06</td><td>1.64</td><td>1.39</td><td>0.11</td><td>0.53</td><td>0.00</td><td>0.02</td><td><lod< td=""></lod<></td></lod<></td></lod<></td></lod<> | 0.01 | <lod< td=""><td>0.00</td><td><lod< td=""><td>1.06</td><td>1.64</td><td>1.39</td><td>0.11</td><td>0.53</td><td>0.00</td><td>0.02</td><td><lod< td=""></lod<></td></lod<></td></lod<> | 0.00 | <lod< td=""><td>1.06</td><td>1.64</td><td>1.39</td><td>0.11</td><td>0.53</td><td>0.00</td><td>0.02</td><td><lod< td=""></lod<></td></lod<> | 1.06 | 1.64 | 1.39 | 0.11 | 0.53 | 0.00 | 0.02 | <lod< td=""></lod<> |
| S6 | 1.56 | 1.54 | 0.23 | <lod< td=""><td>0.05</td><td><lod< td=""><td>0.03</td><td><lod< td=""><td>5.38</td><td>1.60</td><td>0.45</td><td>0.64</td><td>1.89</td><td>0.00</td><td>0.01</td><td><lod< td=""></lod<></td></lod<></td></lod<></td></lod<> | 0.05 | <lod< td=""><td>0.03</td><td><lod< td=""><td>5.38</td><td>1.60</td><td>0.45</td><td>0.64</td><td>1.89</td><td>0.00</td><td>0.01</td><td><lod< td=""></lod<></td></lod<></td></lod<> | 0.03 | <lod< td=""><td>5.38</td><td>1.60</td><td>0.45</td><td>0.64</td><td>1.89</td><td>0.00</td><td>0.01</td><td><lod< td=""></lod<></td></lod<> | 5.38 | 1.60 | 0.45 | 0.64 | 1.89 | 0.00 | 0.01 | <lod< td=""></lod<> |
| S7 | 0.10 | 0.31 | 0.01 | <lod< td=""><td>0.04</td><td><lod< td=""><td>0.02</td><td><lod< td=""><td>0.90</td><td>0.63</td><td>0.30</td><td>0.10</td><td>1.21</td><td>0.10</td><td>0.01</td><td>< LOD</td></lod<></td></lod<></td></lod<> | 0.04 | <lod< td=""><td>0.02</td><td><lod< td=""><td>0.90</td><td>0.63</td><td>0.30</td><td>0.10</td><td>1.21</td><td>0.10</td><td>0.01</td><td>< LOD</td></lod<></td></lod<> | 0.02 | <lod< td=""><td>0.90</td><td>0.63</td><td>0.30</td><td>0.10</td><td>1.21</td><td>0.10</td><td>0.01</td><td>< LOD</td></lod<> | 0.90 | 0.63 | 0.30 | 0.10 | 1.21 | 0.10 | 0.01 | < LOD |
| S8 | 0.87 | 1.51 | 0.06 | <lod< td=""><td>0.04</td><td><lod< td=""><td>0.01</td><td><lod< td=""><td>0.93</td><td>0.37</td><td>0.26</td><td>0.09</td><td>1.78</td><td>0.11</td><td>0.01</td><td>< LOD</td></lod<></td></lod<></td></lod<> | 0.04 | <lod< td=""><td>0.01</td><td><lod< td=""><td>0.93</td><td>0.37</td><td>0.26</td><td>0.09</td><td>1.78</td><td>0.11</td><td>0.01</td><td>< LOD</td></lod<></td></lod<> | 0.01 | <lod< td=""><td>0.93</td><td>0.37</td><td>0.26</td><td>0.09</td><td>1.78</td><td>0.11</td><td>0.01</td><td>< LOD</td></lod<> | 0.93 | 0.37 | 0.26 | 0.09 | 1.78 | 0.11 | 0.01 | < LOD |
| S9 | 1.13 | 0.48 | 0.34 | <lod< td=""><td>0.11</td><td><lod< td=""><td>0.00</td><td><lod< td=""><td>0.58</td><td>0.57</td><td>0.05</td><td>0.06</td><td>0.63</td><td>0.03</td><td>0.00</td><td>< LOD</td></lod<></td></lod<></td></lod<> | 0.11 | <lod< td=""><td>0.00</td><td><lod< td=""><td>0.58</td><td>0.57</td><td>0.05</td><td>0.06</td><td>0.63</td><td>0.03</td><td>0.00</td><td>< LOD</td></lod<></td></lod<> | 0.00 | <lod< td=""><td>0.58</td><td>0.57</td><td>0.05</td><td>0.06</td><td>0.63</td><td>0.03</td><td>0.00</td><td>< LOD</td></lod<> | 0.58 | 0.57 | 0.05 | 0.06 | 0.63 | 0.03 | 0.00 | < LOD |
| S10 | 1.83 | 0.98 | 0.46 | <lod< td=""><td>0.00</td><td><lod< td=""><td>0.00</td><td><lod< td=""><td>0.15</td><td>0.18</td><td>0.00</td><td>0.03</td><td>1.69</td><td>0.05</td><td>0.02</td><td>0.11</td></lod<></td></lod<></td></lod<> | 0.00 | <lod< td=""><td>0.00</td><td><lod< td=""><td>0.15</td><td>0.18</td><td>0.00</td><td>0.03</td><td>1.69</td><td>0.05</td><td>0.02</td><td>0.11</td></lod<></td></lod<> | 0.00 | <lod< td=""><td>0.15</td><td>0.18</td><td>0.00</td><td>0.03</td><td>1.69</td><td>0.05</td><td>0.02</td><td>0.11</td></lod<> | 0.15 | 0.18 | 0.00 | 0.03 | 1.69 | 0.05 | 0.02 | 0.11 |
| S11 | 1.44 | 0.44 | 0.74 | <lod< td=""><td>0.28</td><td><lod< td=""><td>0.01</td><td><lod< td=""><td>0.41</td><td>0.52</td><td>0.00</td><td>0.04</td><td>0.97</td><td>0.02</td><td>0.00</td><td>0.05</td></lod<></td></lod<></td></lod<> | 0.28 | <lod< td=""><td>0.01</td><td><lod< td=""><td>0.41</td><td>0.52</td><td>0.00</td><td>0.04</td><td>0.97</td><td>0.02</td><td>0.00</td><td>0.05</td></lod<></td></lod<> | 0.01 | <lod< td=""><td>0.41</td><td>0.52</td><td>0.00</td><td>0.04</td><td>0.97</td><td>0.02</td><td>0.00</td><td>0.05</td></lod<> | 0.41 | 0.52 | 0.00 | 0.04 | 0.97 | 0.02 | 0.00 | 0.05 |
| S12 | 1.51 | 0.11 | 0.51 | <lod< td=""><td>0.20</td><td><lod< td=""><td>0.01</td><td><lod< td=""><td>0.24</td><td>0.34</td><td>0.06</td><td>0.02</td><td>0.08</td><td>0.01</td><td>0.00</td><td>< LOD</td></lod<></td></lod<></td></lod<> | 0.20 | <lod< td=""><td>0.01</td><td><lod< td=""><td>0.24</td><td>0.34</td><td>0.06</td><td>0.02</td><td>0.08</td><td>0.01</td><td>0.00</td><td>< LOD</td></lod<></td></lod<> | 0.01 | <lod< td=""><td>0.24</td><td>0.34</td><td>0.06</td><td>0.02</td><td>0.08</td><td>0.01</td><td>0.00</td><td>< LOD</td></lod<> | 0.24 | 0.34 | 0.06 | 0.02 | 0.08 | 0.01 | 0.00 | < LOD |
| S13 | 0.97 | 0.46 | 0.78 | <lod< td=""><td>0.20</td><td><lod< td=""><td>0.01</td><td><lod< td=""><td>0.65</td><td>0.31</td><td>0.18</td><td>0.04</td><td>0.79</td><td>0.01</td><td>0.00</td><td>< LOD</td></lod<></td></lod<></td></lod<> | 0.20 | <lod< td=""><td>0.01</td><td><lod< td=""><td>0.65</td><td>0.31</td><td>0.18</td><td>0.04</td><td>0.79</td><td>0.01</td><td>0.00</td><td>< LOD</td></lod<></td></lod<> | 0.01 | <lod< td=""><td>0.65</td><td>0.31</td><td>0.18</td><td>0.04</td><td>0.79</td><td>0.01</td><td>0.00</td><td>< LOD</td></lod<> | 0.65 | 0.31 | 0.18 | 0.04 | 0.79 | 0.01 | 0.00 | < LOD |
| S14 | 1.29 | 0.31 | 0.32 | <lod< td=""><td>0.24</td><td><lod< td=""><td>0.03</td><td><lod< td=""><td>0.09</td><td>0.18</td><td>0.09</td><td>0.01</td><td>1.25</td><td>0.02</td><td>0.00</td><td>< LOD</td></lod<></td></lod<></td></lod<> | 0.24 | <lod< td=""><td>0.03</td><td><lod< td=""><td>0.09</td><td>0.18</td><td>0.09</td><td>0.01</td><td>1.25</td><td>0.02</td><td>0.00</td><td>< LOD</td></lod<></td></lod<> | 0.03 | <lod< td=""><td>0.09</td><td>0.18</td><td>0.09</td><td>0.01</td><td>1.25</td><td>0.02</td><td>0.00</td><td>< LOD</td></lod<> | 0.09 | 0.18 | 0.09 | 0.01 | 1.25 | 0.02 | 0.00 | < LOD |
| S15 | 1.00 | 0.39 | 0.55 | <lod< td=""><td>0.25</td><td><lod< td=""><td>0.01</td><td><lod< td=""><td>0.58</td><td>0.32</td><td>0.00</td><td>0.09</td><td>1.73</td><td>0.04</td><td>0.00</td><td><lod< td=""></lod<></td></lod<></td></lod<></td></lod<> | 0.25 | <lod< td=""><td>0.01</td><td><lod< td=""><td>0.58</td><td>0.32</td><td>0.00</td><td>0.09</td><td>1.73</td><td>0.04</td><td>0.00</td><td><lod< td=""></lod<></td></lod<></td></lod<> | 0.01 | <lod< td=""><td>0.58</td><td>0.32</td><td>0.00</td><td>0.09</td><td>1.73</td><td>0.04</td><td>0.00</td><td><lod< td=""></lod<></td></lod<> | 0.58 | 0.32 | 0.00 | 0.09 | 1.73 | 0.04 | 0.00 | <lod< td=""></lod<> |
| S16 | 1.03 | 0.62 | 0.43 | <lod< td=""><td>0.18</td><td><lod< td=""><td>0.01</td><td><lod< td=""><td>0.93</td><td>0.94</td><td>0.11</td><td>0.12</td><td>0.69</td><td>0.03</td><td>0.00</td><td><lod< td=""></lod<></td></lod<></td></lod<></td></lod<> | 0.18 | <lod< td=""><td>0.01</td><td><lod< td=""><td>0.93</td><td>0.94</td><td>0.11</td><td>0.12</td><td>0.69</td><td>0.03</td><td>0.00</td><td><lod< td=""></lod<></td></lod<></td></lod<> | 0.01 | <lod< td=""><td>0.93</td><td>0.94</td><td>0.11</td><td>0.12</td><td>0.69</td><td>0.03</td><td>0.00</td><td><lod< td=""></lod<></td></lod<> | 0.93 | 0.94 | 0.11 | 0.12 | 0.69 | 0.03 | 0.00 | <lod< td=""></lod<> |
| S17 | 0.50 | 0.17 | 0.34 | <lod< td=""><td>0.23</td><td><lod< td=""><td>0.01</td><td><lod< td=""><td>1.46</td><td>0.73</td><td>0.13</td><td>0.14</td><td>0.58</td><td>0.03</td><td>0.00</td><td><lod< td=""></lod<></td></lod<></td></lod<></td></lod<> | 0.23 | <lod< td=""><td>0.01</td><td><lod< td=""><td>1.46</td><td>0.73</td><td>0.13</td><td>0.14</td><td>0.58</td><td>0.03</td><td>0.00</td><td><lod< td=""></lod<></td></lod<></td></lod<> | 0.01 | <lod< td=""><td>1.46</td><td>0.73</td><td>0.13</td><td>0.14</td><td>0.58</td><td>0.03</td><td>0.00</td><td><lod< td=""></lod<></td></lod<> | 1.46 | 0.73 | 0.13 | 0.14 | 0.58 | 0.03 | 0.00 | <lod< td=""></lod<> |

| S | S18 | 0.63 | 0.09 | 0.14 | <lod< td=""><td>0.02</td><td><lod< td=""><td>0.00</td><td><lod< td=""><td>0.51</td><td>0.45</td><td>0.10</td><td>0.09</td><td>0.32</td><td>0.00</td><td>0.00</td><td><lod< td=""></lod<></td></lod<></td></lod<></td></lod<> | 0.02 | <lod< td=""><td>0.00</td><td><lod< td=""><td>0.51</td><td>0.45</td><td>0.10</td><td>0.09</td><td>0.32</td><td>0.00</td><td>0.00</td><td><lod< td=""></lod<></td></lod<></td></lod<> | 0.00 | <lod< td=""><td>0.51</td><td>0.45</td><td>0.10</td><td>0.09</td><td>0.32</td><td>0.00</td><td>0.00</td><td><lod< td=""></lod<></td></lod<> | 0.51 | 0.45 | 0.10 | 0.09 | 0.32 | 0.00 | 0.00 | <lod< td=""></lod<> |
|---|-----|------|------|------|--|------|---|------|--|------|------|------|------|------|------|------|---------------------|
| S | S19 | 0.68 | 0.19 | 0.16 | <lod< td=""><td>0.14</td><td><lod< td=""><td>0.00</td><td><lod< td=""><td>0.02</td><td>0.03</td><td>0.01</td><td>0.00</td><td>0.09</td><td>0.00</td><td>0.00</td><td><lod< td=""></lod<></td></lod<></td></lod<></td></lod<> | 0.14 | <lod< td=""><td>0.00</td><td><lod< td=""><td>0.02</td><td>0.03</td><td>0.01</td><td>0.00</td><td>0.09</td><td>0.00</td><td>0.00</td><td><lod< td=""></lod<></td></lod<></td></lod<> | 0.00 | <lod< td=""><td>0.02</td><td>0.03</td><td>0.01</td><td>0.00</td><td>0.09</td><td>0.00</td><td>0.00</td><td><lod< td=""></lod<></td></lod<> | 0.02 | 0.03 | 0.01 | 0.00 | 0.09 | 0.00 | 0.00 | <lod< td=""></lod<> |
| S | S20 | 0.48 | 0.07 | 0.08 | <lod< td=""><td>0.03</td><td><lod< td=""><td>0.00</td><td><lod< td=""><td>0.64</td><td>0.38</td><td>0.00</td><td>0.06</td><td>1.69</td><td>0.04</td><td>0.00</td><td><lod< td=""></lod<></td></lod<></td></lod<></td></lod<> | 0.03 | <lod< td=""><td>0.00</td><td><lod< td=""><td>0.64</td><td>0.38</td><td>0.00</td><td>0.06</td><td>1.69</td><td>0.04</td><td>0.00</td><td><lod< td=""></lod<></td></lod<></td></lod<> | 0.00 | <lod< td=""><td>0.64</td><td>0.38</td><td>0.00</td><td>0.06</td><td>1.69</td><td>0.04</td><td>0.00</td><td><lod< td=""></lod<></td></lod<> | 0.64 | 0.38 | 0.00 | 0.06 | 1.69 | 0.04 | 0.00 | <lod< td=""></lod<> |
| S | S21 | 0.64 | 0.08 | 0.13 | <lod< td=""><td>0.03</td><td><lod< td=""><td>0.00</td><td><lod< td=""><td>0.93</td><td>0.86</td><td>0.00</td><td>0.00</td><td>0.88</td><td>0.03</td><td>0.00</td><td><lod< td=""></lod<></td></lod<></td></lod<></td></lod<> | 0.03 | <lod< td=""><td>0.00</td><td><lod< td=""><td>0.93</td><td>0.86</td><td>0.00</td><td>0.00</td><td>0.88</td><td>0.03</td><td>0.00</td><td><lod< td=""></lod<></td></lod<></td></lod<> | 0.00 | <lod< td=""><td>0.93</td><td>0.86</td><td>0.00</td><td>0.00</td><td>0.88</td><td>0.03</td><td>0.00</td><td><lod< td=""></lod<></td></lod<> | 0.93 | 0.86 | 0.00 | 0.00 | 0.88 | 0.03 | 0.00 | <lod< td=""></lod<> |
| S | S22 | 0.59 | 0.21 | 0.14 | <lod< td=""><td>0.05</td><td><lod< td=""><td>0.01</td><td><lod< td=""><td>0.88</td><td>1.07</td><td>0.00</td><td>0.00</td><td>2.37</td><td>0.04</td><td>0.00</td><td><lod< td=""></lod<></td></lod<></td></lod<></td></lod<> | 0.05 | <lod< td=""><td>0.01</td><td><lod< td=""><td>0.88</td><td>1.07</td><td>0.00</td><td>0.00</td><td>2.37</td><td>0.04</td><td>0.00</td><td><lod< td=""></lod<></td></lod<></td></lod<> | 0.01 | <lod< td=""><td>0.88</td><td>1.07</td><td>0.00</td><td>0.00</td><td>2.37</td><td>0.04</td><td>0.00</td><td><lod< td=""></lod<></td></lod<> | 0.88 | 1.07 | 0.00 | 0.00 | 2.37 | 0.04 | 0.00 | <lod< td=""></lod<> |
| S | S23 | 0.78 | 0.18 | 0.09 | <lod< td=""><td>0.12</td><td><lod< td=""><td>0.00</td><td><lod< td=""><td>0.18</td><td>0.56</td><td>0.00</td><td>0.00</td><td>0.80</td><td>0.03</td><td>0.00</td><td><lod< td=""></lod<></td></lod<></td></lod<></td></lod<> | 0.12 | <lod< td=""><td>0.00</td><td><lod< td=""><td>0.18</td><td>0.56</td><td>0.00</td><td>0.00</td><td>0.80</td><td>0.03</td><td>0.00</td><td><lod< td=""></lod<></td></lod<></td></lod<> | 0.00 | <lod< td=""><td>0.18</td><td>0.56</td><td>0.00</td><td>0.00</td><td>0.80</td><td>0.03</td><td>0.00</td><td><lod< td=""></lod<></td></lod<> | 0.18 | 0.56 | 0.00 | 0.00 | 0.80 | 0.03 | 0.00 | <lod< td=""></lod<> |
| S | S24 | 0.21 | 0.37 | 0.01 | <lod< td=""><td>0.07</td><td><lod< td=""><td>0.00</td><td><lod< td=""><td>0.70</td><td>0.67</td><td>0.01</td><td>0.00</td><td>1.56</td><td>0.03</td><td>0.00</td><td><lod< td=""></lod<></td></lod<></td></lod<></td></lod<> | 0.07 | <lod< td=""><td>0.00</td><td><lod< td=""><td>0.70</td><td>0.67</td><td>0.01</td><td>0.00</td><td>1.56</td><td>0.03</td><td>0.00</td><td><lod< td=""></lod<></td></lod<></td></lod<> | 0.00 | <lod< td=""><td>0.70</td><td>0.67</td><td>0.01</td><td>0.00</td><td>1.56</td><td>0.03</td><td>0.00</td><td><lod< td=""></lod<></td></lod<> | 0.70 | 0.67 | 0.01 | 0.00 | 1.56 | 0.03 | 0.00 | <lod< td=""></lod<> |
| S | S25 | 0.19 | 0.15 | 0.05 | <lod< td=""><td>0.07</td><td><lod< td=""><td>0.02</td><td><lod< td=""><td>0.18</td><td>0.37</td><td>0.11</td><td>0.00</td><td>0.89</td><td>0.00</td><td>0.00</td><td><lod< td=""></lod<></td></lod<></td></lod<></td></lod<> | 0.07 | <lod< td=""><td>0.02</td><td><lod< td=""><td>0.18</td><td>0.37</td><td>0.11</td><td>0.00</td><td>0.89</td><td>0.00</td><td>0.00</td><td><lod< td=""></lod<></td></lod<></td></lod<> | 0.02 | <lod< td=""><td>0.18</td><td>0.37</td><td>0.11</td><td>0.00</td><td>0.89</td><td>0.00</td><td>0.00</td><td><lod< td=""></lod<></td></lod<> | 0.18 | 0.37 | 0.11 | 0.00 | 0.89 | 0.00 | 0.00 | <lod< td=""></lod<> |
| S | S26 | 0.55 | 0.26 | 0.05 | <lod< td=""><td>0.05</td><td><lod< td=""><td>0.00</td><td><lod< td=""><td>0.21</td><td>0.34</td><td>0.17</td><td>0.00</td><td>0.84</td><td>0.00</td><td>0.00</td><td><lod< td=""></lod<></td></lod<></td></lod<></td></lod<> | 0.05 | <lod< td=""><td>0.00</td><td><lod< td=""><td>0.21</td><td>0.34</td><td>0.17</td><td>0.00</td><td>0.84</td><td>0.00</td><td>0.00</td><td><lod< td=""></lod<></td></lod<></td></lod<> | 0.00 | <lod< td=""><td>0.21</td><td>0.34</td><td>0.17</td><td>0.00</td><td>0.84</td><td>0.00</td><td>0.00</td><td><lod< td=""></lod<></td></lod<> | 0.21 | 0.34 | 0.17 | 0.00 | 0.84 | 0.00 | 0.00 | <lod< td=""></lod<> |
| S | S27 | 0.38 | 0.38 | 0.05 | <lod< td=""><td>0.02</td><td><lod< td=""><td>0.01</td><td><lod< td=""><td>1.47</td><td>0.58</td><td>0.17</td><td>0.00</td><td>1.14</td><td>0.00</td><td>0.00</td><td><lod< td=""></lod<></td></lod<></td></lod<></td></lod<> | 0.02 | <lod< td=""><td>0.01</td><td><lod< td=""><td>1.47</td><td>0.58</td><td>0.17</td><td>0.00</td><td>1.14</td><td>0.00</td><td>0.00</td><td><lod< td=""></lod<></td></lod<></td></lod<> | 0.01 | <lod< td=""><td>1.47</td><td>0.58</td><td>0.17</td><td>0.00</td><td>1.14</td><td>0.00</td><td>0.00</td><td><lod< td=""></lod<></td></lod<> | 1.47 | 0.58 | 0.17 | 0.00 | 1.14 | 0.00 | 0.00 | <lod< td=""></lod<> |
| S | S28 | 0.25 | 0.33 | 0.02 | <lod< td=""><td>0.10</td><td><lod< td=""><td>0.02</td><td><lod< td=""><td>0.19</td><td>0.40</td><td>0.22</td><td>0.02</td><td>0.37</td><td>0.00</td><td>0.00</td><td><lod< td=""></lod<></td></lod<></td></lod<></td></lod<> | 0.10 | <lod< td=""><td>0.02</td><td><lod< td=""><td>0.19</td><td>0.40</td><td>0.22</td><td>0.02</td><td>0.37</td><td>0.00</td><td>0.00</td><td><lod< td=""></lod<></td></lod<></td></lod<> | 0.02 | <lod< td=""><td>0.19</td><td>0.40</td><td>0.22</td><td>0.02</td><td>0.37</td><td>0.00</td><td>0.00</td><td><lod< td=""></lod<></td></lod<> | 0.19 | 0.40 | 0.22 | 0.02 | 0.37 | 0.00 | 0.00 | <lod< td=""></lod<> |
| S | S29 | 0.2 | 0.33 | 0.01 | <lod< td=""><td>0.09</td><td><lod< td=""><td>0.00</td><td><lod< td=""><td>0.30</td><td>0.33</td><td>0.15</td><td>0.03</td><td>0.82</td><td>0.02</td><td>0.00</td><td><lod< td=""></lod<></td></lod<></td></lod<></td></lod<> | 0.09 | <lod< td=""><td>0.00</td><td><lod< td=""><td>0.30</td><td>0.33</td><td>0.15</td><td>0.03</td><td>0.82</td><td>0.02</td><td>0.00</td><td><lod< td=""></lod<></td></lod<></td></lod<> | 0.00 | <lod< td=""><td>0.30</td><td>0.33</td><td>0.15</td><td>0.03</td><td>0.82</td><td>0.02</td><td>0.00</td><td><lod< td=""></lod<></td></lod<> | 0.30 | 0.33 | 0.15 | 0.03 | 0.82 | 0.02 | 0.00 | <lod< td=""></lod<> |
| S | S30 | 0.11 | 0.15 | 0.01 | <lod< td=""><td>0.09</td><td><lod< td=""><td>0.00</td><td><lod< td=""><td>0.23</td><td>0.31</td><td>0.10</td><td>0.02</td><td>0.57</td><td>0.00</td><td>0.00</td><td><lod< td=""></lod<></td></lod<></td></lod<></td></lod<> | 0.09 | <lod< td=""><td>0.00</td><td><lod< td=""><td>0.23</td><td>0.31</td><td>0.10</td><td>0.02</td><td>0.57</td><td>0.00</td><td>0.00</td><td><lod< td=""></lod<></td></lod<></td></lod<> | 0.00 | <lod< td=""><td>0.23</td><td>0.31</td><td>0.10</td><td>0.02</td><td>0.57</td><td>0.00</td><td>0.00</td><td><lod< td=""></lod<></td></lod<> | 0.23 | 0.31 | 0.10 | 0.02 | 0.57 | 0.00 | 0.00 | <lod< td=""></lod<> |
| S | S31 | 0.55 | 0.26 | 0.05 | <lod< td=""><td>0.14</td><td><lod< td=""><td>0.02</td><td><lod< td=""><td>0.57</td><td>0.47</td><td>0.15</td><td>0.04</td><td>1.15</td><td>0.03</td><td>0.00</td><td><lod< td=""></lod<></td></lod<></td></lod<></td></lod<> | 0.14 | <lod< td=""><td>0.02</td><td><lod< td=""><td>0.57</td><td>0.47</td><td>0.15</td><td>0.04</td><td>1.15</td><td>0.03</td><td>0.00</td><td><lod< td=""></lod<></td></lod<></td></lod<> | 0.02 | <lod< td=""><td>0.57</td><td>0.47</td><td>0.15</td><td>0.04</td><td>1.15</td><td>0.03</td><td>0.00</td><td><lod< td=""></lod<></td></lod<> | 0.57 | 0.47 | 0.15 | 0.04 | 1.15 | 0.03 | 0.00 | <lod< td=""></lod<> |
| S | S32 | 0.12 | 0.15 | 0.01 | <lod< td=""><td>0.05</td><td><lod< td=""><td>0.00</td><td><lod< td=""><td>0.35</td><td>0.42</td><td>0.12</td><td>0.02</td><td>4.90</td><td>0.00</td><td>0.00</td><td><lod< td=""></lod<></td></lod<></td></lod<></td></lod<> | 0.05 | <lod< td=""><td>0.00</td><td><lod< td=""><td>0.35</td><td>0.42</td><td>0.12</td><td>0.02</td><td>4.90</td><td>0.00</td><td>0.00</td><td><lod< td=""></lod<></td></lod<></td></lod<> | 0.00 | <lod< td=""><td>0.35</td><td>0.42</td><td>0.12</td><td>0.02</td><td>4.90</td><td>0.00</td><td>0.00</td><td><lod< td=""></lod<></td></lod<> | 0.35 | 0.42 | 0.12 | 0.02 | 4.90 | 0.00 | 0.00 | <lod< td=""></lod<> |
| S | 533 | 0.18 | 0.08 | 0.04 | <lod< td=""><td>0.04</td><td><lod< td=""><td>0.00</td><td><lod< td=""><td>0.29</td><td>0.71</td><td>0.03</td><td>0.06</td><td>1.14</td><td>0.02</td><td>0.00</td><td><lod< td=""></lod<></td></lod<></td></lod<></td></lod<> | 0.04 | <lod< td=""><td>0.00</td><td><lod< td=""><td>0.29</td><td>0.71</td><td>0.03</td><td>0.06</td><td>1.14</td><td>0.02</td><td>0.00</td><td><lod< td=""></lod<></td></lod<></td></lod<> | 0.00 | <lod< td=""><td>0.29</td><td>0.71</td><td>0.03</td><td>0.06</td><td>1.14</td><td>0.02</td><td>0.00</td><td><lod< td=""></lod<></td></lod<> | 0.29 | 0.71 | 0.03 | 0.06 | 1.14 | 0.02 | 0.00 | <lod< td=""></lod<> |
| S | 534 | 0.35 | 0.12 | 0.02 | <lod< td=""><td>0.05</td><td><lod< td=""><td>0.00</td><td><lod< td=""><td>0.35</td><td>0.32</td><td>0.11</td><td>0.01</td><td>1.08</td><td>0.00</td><td>0.00</td><td><lod< td=""></lod<></td></lod<></td></lod<></td></lod<> | 0.05 | <lod< td=""><td>0.00</td><td><lod< td=""><td>0.35</td><td>0.32</td><td>0.11</td><td>0.01</td><td>1.08</td><td>0.00</td><td>0.00</td><td><lod< td=""></lod<></td></lod<></td></lod<> | 0.00 | <lod< td=""><td>0.35</td><td>0.32</td><td>0.11</td><td>0.01</td><td>1.08</td><td>0.00</td><td>0.00</td><td><lod< td=""></lod<></td></lod<> | 0.35 | 0.32 | 0.11 | 0.01 | 1.08 | 0.00 | 0.00 | <lod< td=""></lod<> |
| S | 535 | 0.2 | 0.16 | 0.02 | <lod< td=""><td>0.04</td><td><lod< td=""><td>0.00</td><td><lod< td=""><td>0.29</td><td>0.79</td><td>0.01</td><td>0.06</td><td>0.74</td><td>0.00</td><td>0.00</td><td><lod< td=""></lod<></td></lod<></td></lod<></td></lod<> | 0.04 | <lod< td=""><td>0.00</td><td><lod< td=""><td>0.29</td><td>0.79</td><td>0.01</td><td>0.06</td><td>0.74</td><td>0.00</td><td>0.00</td><td><lod< td=""></lod<></td></lod<></td></lod<> | 0.00 | <lod< td=""><td>0.29</td><td>0.79</td><td>0.01</td><td>0.06</td><td>0.74</td><td>0.00</td><td>0.00</td><td><lod< td=""></lod<></td></lod<> | 0.29 | 0.79 | 0.01 | 0.06 | 0.74 | 0.00 | 0.00 | <lod< td=""></lod<> |
| S | S36 | 0.19 | 0.12 | 0.01 | <lod< td=""><td>0.03</td><td><lod< td=""><td>0.00</td><td><lod< td=""><td>0.58</td><td>0.75</td><td>0.00</td><td>0.10</td><td>1.36</td><td>0.00</td><td>0.00</td><td><lod< td=""></lod<></td></lod<></td></lod<></td></lod<> | 0.03 | <lod< td=""><td>0.00</td><td><lod< td=""><td>0.58</td><td>0.75</td><td>0.00</td><td>0.10</td><td>1.36</td><td>0.00</td><td>0.00</td><td><lod< td=""></lod<></td></lod<></td></lod<> | 0.00 | <lod< td=""><td>0.58</td><td>0.75</td><td>0.00</td><td>0.10</td><td>1.36</td><td>0.00</td><td>0.00</td><td><lod< td=""></lod<></td></lod<> | 0.58 | 0.75 | 0.00 | 0.10 | 1.36 | 0.00 | 0.00 | <lod< td=""></lod<> |
| | | | | | | | | | | | | | | 1 | | | |

3 In situ environmental parameters and nutrients of plastisphere surrounding environment.

| Sample ID | Temp | pН | Salinity | DOC | NO ₃ - | PO ₄ ³ - |
|----------------|--------|------|----------|-------|-------------------|--------------------------------|
| EnvHKSumMc1 | 27.8 | 8.3 | 32 | 3.274 | 2 | 0.44 |
| EnvHKSumMR1 | 30.1 | 8.2 | 32 | 7.297 | 2.4 | 0.21 |
| EnvHKSumMR2 | 30.2 | 8.11 | 33 | 5.681 | 2.9 | 0.21 |
| EnvHKSumMR3 | 30.2 | 8.16 | 29 | 19.6 | 2 | 0.24 |
| EnvHKSumMR4 | 30.3 | 8.16 | 29 | 13.48 | 1.1 | 0.28 |
| EnvHKSumPort1 | 31.4 | 8.01 | 23 | 12.97 | 1.7 | 0.15 |
| EnvHKSumWWTP2 | 30.8 | 7.44 | 27 | 3.829 | 2.6 | 0.22 |
| EnvHKWinBeach1 | 18.967 | 7.81 | 35.63 | 5.138 | 3.3 | 0.19 |
| EnvHKWinBeach2 | 18.966 | 7.8 | 35.66 | 5.128 | 2.3 | 0.37 |
| EnvHKWinBeach3 | 18.985 | 7.81 | 35.68 | 5.294 | 1.5 | 0.3 |
| EnvHKWinEs1 | 20.071 | 7.95 | 34.95 | 7.056 | 1 | 0.36 |
| EnvHKWinEs2 | 20.047 | 7.91 | 34.98 | 7.015 | 1.5 | 0.28 |
| EnvHKWinEs4 | 20.02 | 7.89 | 34.89 | 3.996 | 0.5 | 0.05 |
| EnvHKWinEs5 | 19.92 | 7.9 | 35.27 | 4.555 | 1.2 | 0.04 |
| EnvHKWinMc1 | 22.134 | 7.59 | 29.54 | 7.467 | 1.6 | 1.71 |
| EnvHKWinMc10 | 16.164 | 7.9 | 34.54 | 2.81 | 1.8 | 0.18 |
| EnvHKWinMc11 | 17.81 | 7.99 | 35.44 | 2.647 | 1.7 | 0.53 |
| EnvHKWinMc12 | 17.93 | 7.97 | 35.48 | 2.653 | 1.8 | 0.57 |
| EnvHKWinMc13 | 17.998 | 8.01 | 35.47 | 2.656 | 1 | 0.35 |
| EnvHKWinMc15 | 17.982 | 8 | 35.43 | 2.727 | 1 | 0.33 |
| EnvHKWinMc2 | 22.278 | 7.6 | 29.43 | 9.211 | 1.3 | 1.59 |
| EnvHKWinMc3 | 22.345 | 7.56 | 29.44 | 7.804 | 1.3 | 1.54 |
| EnvHKWinMc4 | 21.869 | 7.41 | 29.41 | 8.466 | 1.8 | 1.21 |
| EnvHKWinMc5 | 22.026 | 7.53 | 29.69 | 17.86 | 1.4 | 1.15 |
| EnvHKWinMc6 | 16.185 | 7.95 | 34.33 | 4.122 | 1.3 | 0.27 |
| EnvHKWinMc7 | 16.158 | 7.91 | 34.54 | 3.753 | 1.8 | 0.15 |
| EnvHKWinMc8 | 16.189 | 7.86 | 34.54 | 3.813 | 1.6 | 0.27 |
| EnvHKWinMc9 | 16.099 | 7.88 | 34.54 | 2.73 | 1.4 | 0.08 |
| EnvHKWinMR1 | 15.652 | 8.01 | 34.21 | 2.809 | 1.7 | 0.65 |
| EnvHKWinPort1 | 17.456 | 7.89 | 26.19 | 2.792 | 1.1 | 1.65 |
| EnvHKWinPort2 | 18.402 | 7.99 | 34.07 | 2.751 | 1 | 0.46 |
| EnvHKWinPort3 | 18.689 | 8.09 | 33.35 | 3.046 | 1.2 | 0.41 |
| EnvHKWinPort4 | 18.569 | 8.02 | 34.42 | 2.396 | 1 | 0.69 |

| EnvHKWinPort5 | 18.16 | 7.99 | 35.34 | 4.131 | 1.2 | 0.31 |
|----------------|--------|------|-------|--------|-------|-------|
| EnvHKWinWWTP1 | 17.881 | 7.84 | 35.03 | 3.843 | 0.9 | 0.47 |
| EnvHKWinWWTP2 | 17.894 | 7.84 | 34.58 | 3.611 | 0.8 | 0.58 |
| EnvHKWinWWTP3 | 18.35 | 7.83 | 34.08 | 4.325 | 1.6 | 0.36 |
| EnvHKWinWWTP4 | 18.113 | 7.8 | 33.95 | 4.191 | 1.5 | 0.52 |
| EnvQDSumBeach1 | 27.5 | 8.3 | 24.5 | 1.3821 | 0.171 | 0.189 |
| EnvQDSumBeach2 | 27.7 | 8.2 | 24 | 1.066 | 0.035 | 0.318 |
| EnvQDSumBeach4 | 27.8 | 8.2 | 24.4 | 2.1802 | 0.039 | 0.19 |
| EnvQDSumBeach5 | 28 | 8.3 | 23.9 | 1.1783 | 0.085 | 0.239 |
| EnvQDSumEs1 | 30.8 | 8.6 | 19.6 | 3.0857 | 0.498 | 0.134 |
| EnvQDSumEs2 | 30.8 | 8.7 | 18.2 | 3.1053 | 0.619 | 0.142 |
| EnvQDSumEs3 | 30.3 | 8.9 | 20.6 | 1.564 | 0.34 | 0.222 |
| EnvQDSumMc1 | 30.7 | 8.2 | 23.2 | 1.8028 | 0.099 | 0.211 |
| EnvQDSumMc2 | 31.3 | 8.2 | 23.1 | 1.7373 | 0.104 | 0.308 |
| EnvQDSumMc4 | 31.1 | 8.2 | 23.6 | 1.5619 | 0.081 | 0.293 |
| EnvQDSumMc5 | 31.6 | 8.2 | 23.4 | 1.685 | 0.086 | 0.2 |
| EnvQDSumMR1 | 28.7 | 8.3 | 24.6 | 2.2215 | 0.009 | 0.245 |
| EnvQDSumMR2 | 28.2 | 8.3 | 24.4 | 1.3169 | 0.009 | 0.267 |
| EnvQDSumMR3 | 28 | 8.3 | 24.7 | 0.6398 | 0.009 | 0.332 |
| EnvQDSumPort1 | 27.8 | 8.1 | 23.5 | 0.6984 | 0.02 | 0.209 |
| EnvQDSumPort2 | 29.5 | 8.1 | 24.1 | 1.5084 | 0.075 | 0.218 |
| EnvQDSumPort3 | 28.2 | 8.2 | 24.3 | 1.6302 | 0.201 | 0.151 |
| EnvQDSumPort4 | 28.9 | 8.2 | 24.2 | 0.5761 | 0.069 | 0.272 |
| EnvQDSumPort5 | 27.8 | 8.3 | 24.3 | 0.08 | 0.005 | 0.32 |
| EnvQDSumWWTP1 | 30.8 | 7.9 | 24.3 | 2.1985 | 0.272 | 0.166 |
| EnvQDSumWWTP2 | 28.5 | 8.2 | 24.3 | 0.9544 | 0.183 | 0.223 |
| EnvQDSumWWTP3 | 28.4 | 8.2 | 24.6 | 1.0662 | 0.143 | 0.262 |
| EnvQDSumWWTP4 | 28.1 | 8.1 | 24.8 | 1.0851 | 0.076 | 0.21 |
| EnvQDSumWWTP5 | 28 | 8.2 | 24.8 | 1.0615 | 0.051 | 0.273 |
| EnvQDWinBeach1 | 6.2 | 8.5 | 27.8 | 6.11 | 0.008 | 0.108 |
| EnvQDWinBeach2 | 7 | 8.49 | 27.7 | 6.155 | 0.008 | 0.128 |
| EnvQDWinBeach3 | 6.6 | 8.52 | 27.7 | 5.297 | 0.228 | 0.12 |
| EnvQDWinBeach4 | 7 | 8.51 | 27.7 | 5.67 | 0.016 | 0.082 |
| EnvQDWinBeach5 | 6.1 | 8.5 | 27.8 | 5.807 | 0.011 | 0.003 |
| EnvQDWinEs1 | 7.2 | 8.08 | 11.2 | 19.015 | 2.701 | 0.098 |
| EnvQDWinEs2 | 7.5 | 8.1 | 8.44 | 19.468 | 5.555 | 0.166 |
| EnvQDWinEs3 | 7.2 | 8.37 | 13.7 | 20.168 | 5.08 | 0.116 |
| | | | | | | |

| EnvQDWinEs4 | 7.4 | 8.42 | 8.72 | 19.153 | 8.502 | 0.14 |
|---------------|-----|------|------|--------|-------|-------|
| EnvQDWinEs5 | 7.6 | 8.38 | 8.66 | 18.007 | 3.181 | 0.099 |
| EnvQDWinMc1 | 4.5 | 8.49 | 27.3 | 6.836 | 0.045 | 0.152 |
| EnvQDWinMc2 | 4.6 | 8.5 | 28.1 | 6.532 | 0.025 | 0.12 |
| EnvQDWinMc3 | 5.1 | 8.52 | 27.8 | 6.154 | 0.042 | 0.117 |
| EnvQDWinMc4 | 5 | 8.5 | 27.5 | 4.525 | 0.027 | 0.098 |
| EnvQDWinMc5 | 6.8 | 8.52 | 26.9 | 6.406 | 0.075 | 0.106 |
| EnvQDWinMR1 | 6.2 | 8.5 | 27 | 4.551 | 0.017 | 0.001 |
| EnvQDWinMR2 | 5.5 | 8.52 | 26.8 | 5.511 | 0.007 | 0.002 |
| EnvQDWinMR3 | 6.3 | 8.49 | 27.6 | 5.877 | 0.001 | 0.001 |
| EnvQDWinMR4 | 5.6 | 8.53 | 27.2 | 5.215 | 0.008 | 0.003 |
| EnvQDWinMR5 | 6.1 | 8.49 | 26.9 | 5.876 | 0.01 | 0.001 |
| EnvQDWinPort1 | 6.1 | 8.46 | 26.9 | 7.147 | 0.008 | 0.001 |
| EnvQDWinPort2 | 5.9 | 8.45 | 27.3 | 7.658 | 0.018 | 0.001 |
| EnvQDWinPort3 | 5.7 | 8.44 | 27.6 | 11.553 | 0.036 | 0.001 |
| EnvQDWinPort4 | 6.3 | 8.49 | 27.7 | 3.473 | 0.066 | 0.001 |
| EnvQDWinPort5 | 5.7 | 8.46 | 27.2 | 3.855 | 0.121 | 0.001 |
| EnvQDWinWWTP1 | 5.7 | 8.42 | 27 | 8.437 | 0.251 | 0.079 |
| EnvQDWinWWTP2 | 6.3 | 8.44 | 26.7 | 7.571 | 0.285 | 0.093 |
| EnvQDWinWWTP3 | 6 | 8.44 | 26.7 | 8.144 | 0.357 | 0.074 |
| EnvQDWinWWTP4 | 6.7 | 8.44 | 26.7 | 7.348 | 0.327 | 0.081 |
| EnvQDWinWWTP5 | 6.2 | 8.42 | 26.7 | 7.878 | 0.245 | 0.117 |

4 Polymer identification of plastics.

| Site | Polymer | HQI | Site | Polymer | HQI |
|---------|-------------------------------|-------|--------|-------------------------------|-------|
| Beach | Resin | 81.22 | WWTP | Polystyrene | 72.73 |
| | Ethylene propylene copolymer | 81.24 | | Polypropylenes | 72.89 |
| | Polypropylenes | 82.93 | | Polypropylenes | 78.85 |
| | Polypropylenes | 86.63 | | Propylene-ethylene copolymer | 85.7 |
| | Polypropylenes | 91.8 | | Polypropylene-co- ethylene | 94.18 |
| Reserve | Propylene-ethylene copolymer | 74.43 | Port-1 | Polystyrene | 79.85 |
| | Polyethylene terephthalate | 76.65 | | Ethylene-Propylene copolymer | 82.67 |
| | Polyethylenes | 82.9 | | Polyethylenes | 85.68 |
| | Polyamide-6,6 | 91.6 | | Polypropylenes | 86.96 |
| | Low-moecular | 91.82 | | Polyethylenes | 88.67 |

| Port-2 | polyethylene Polyethylenes | 76.74 | Estuary-1 | Polyethylene, chlorinated | 73.38 |
|-------------------|---|----------------|------------------|---|----------------|
| | Polyethylene-co- Propylene copolymer | 81.87 | | Polyethylenes | 86.42 |
| | Polypropylenes Polyethylene Terephthalate | 88.33 91.33 | | Ethylene copolymer Polyethylenes | 87.78 90.04 |
| | Polyethylenes | 92.28 | | Polypropylene isotactic | 90.85 |
| Estuary-2 | Polyethylenes | 81.21 | | Low-molecular polyethylene | 91.42 |
| | Polypropylenes | 85.46 | Marine culture-1 | Polypropylene | 79.74 |
| | Polyethylene Terephthalate | 90.05 | | Polyethylenes | 81.71 |
| | Polyethylenes | 90.47 | | Polyethylene Terephthalate | 82.12 |
| | Polyethylenes | 90.72 | | Propylene-ethylene copolymer | 83.88 |
| Marine culture-5 | Polystyrene | 72.75 | | Polypropylenes | 84.54 |
| | Polypropylenes | 79.7 | | Polyethylene-co- Propylene copolymer | 89.68 |
| | Polyester | 81.05 | | Ethylene-propylene copolymer | 93.53 |
| | Polypropylenes | 82 | Marine culture-6 | Polyethylene Terephthalate | 70.28 |
| | Butadiene rubber | 84.82 | | Polypropylenes | 80 |
| | Polyethylene-co- | 92.72 | | Polystyrenes | 81.09 |
| | Propylene copolymer | | | | |
| | Ethylene-propylene | 93.78 | | Polypropylenes | 89.8 |
| | copolymer | | | | |
| Marine culture-10 | Polystyrene | 74.07 | | Polypropylenes | 93.12 |
| | Polyester urethane | 74.77 | | | |
| | Polyethylenes | 82.21 | | | |
| | Polypropylenes | 85.62 | | | |
| | Polypropylenes | 90.83 | | | |
| | Polyethylenes | 90.9 | | | |

Appendix 4

This section provides bioinformatic and statistical results of the thesis.

1 The differential ASVs between the plastisphere (enriched) and the ambient (depleted) phytoplankton community.

| ID | log ₂ CPM | Level |
|----------|----------------------|----------|----------|----------------------|----------|----------|----------------------|----------|----------|----------------------|----------|
| ASV_51 | 14.27694 | Enriched | ASV_873 | 9.04205 | Enriched | ASV_1195 | 7.8738 | Depleted | ASV_686 | 9.07972 | Depleted |
| ASV_29 | 14.8988 | Enriched | ASV_2505 | 7.04359 | Enriched | ASV_277 | 10.64113 | Depleted | ASV_1315 | 7.44628 | Depleted |
| ASV_370 | 11.0917 | Enriched | ASV_792 | 8.70156 | Enriched | ASV_745 | 8.30355 | Depleted | ASV_279 | 10.26384 | Depleted |
| ASV_239 | 12.184 | Enriched | ASV_1885 | 7.89681 | Enriched | ASV_304 | 10.29413 | Depleted | ASV_998 | 8.64273 | Depleted |
| ASV_135 | 12.90295 | Enriched | ASV_3181 | 6.74108 | Enriched | ASV_160 | 11.56931 | Depleted | ASV_932 | 7.85574 | Depleted |
| ASV_44 | 14.19868 | Enriched | ASV_2933 | 6.77713 | Enriched | ASV_1080 | 7.7326 | Depleted | ASV_493 | 9.3872 | Depleted |
| ASV_190 | 12.09706 | Enriched | ASV_195 | 11.08615 | Enriched | ASV_663 | 8.69255 | Depleted | ASV_682 | 8.0979 | Depleted |
| ASV_38 | 14.87331 | Enriched | ASV_2281 | 7.11727 | Enriched | ASV_1068 | 7.48669 | Depleted | ASV_1211 | 7.79563 | Depleted |
| ASV_249 | 12.07192 | Enriched | ASV_985 | 8.51717 | Enriched | ASV_15 | 15.23821 | Depleted | ASV_1555 | 7.23997 | Depleted |
| ASV_1358 | 8.28062 | Enriched | ASV_1326 | 8.17306 | Enriched | ASV_653 | 8.94893 | Depleted | ASV_4271 | 9.31407 | Depleted |
| ASV_457 | 10.44725 | Enriched | ASV_1152 | 7.91642 | Enriched | ASV_1239 | 7.31894 | Depleted | ASV_126 | 11.1056 | Depleted |
| ASV_828 | 9.12886 | Enriched | ASV_644 | 12.09975 | Enriched | ASV_639 | 9.81806 | Depleted | ASV_1047 | 7.83228 | Depleted |
| ASV_827 | 9.39247 | Enriched | ASV_561 | 9.64329 | Enriched | ASV_1304 | 7.33697 | Depleted | ASV_1117 | 7.7183 | Depleted |
| ASV_162 | 12.36436 | Enriched | ASV_902 | 8.86777 | Enriched | ASV_247 | 10.6357 | Depleted | ASV_318 | 10.18774 | Depleted |
| ASV_423 | 11.49086 | Enriched | ASV_1161 | 8.11915 | Enriched | ASV_555 | 9.03394 | Depleted | ASV_1222 | 7.40094 | Depleted |
| ASV_707 | 9.44698 | Enriched | ASV_2244 | 7.06992 | Enriched | ASV_810 | 8.99102 | Depleted | ASV_1034 | 7.69751 | Depleted |

| ASV_208 | 12.20323 | Enriched | ASV_188 | 12.1253 | Enriched | ASV_168 | 11.43672 | Depleted | ASV_1038 | 7.57413 | Depleted |
|----------|----------|----------|----------|----------|----------|----------|----------|----------|----------|----------|----------|
| ASV_55 | 14.93803 | Enriched | ASV_492 | 9.78311 | Enriched | ASV_634 | 8.73382 | Depleted | ASV_972 | 8.37969 | Depleted |
| ASV_118 | 13.03003 | Enriched | ASV_246 | 11.80886 | Enriched | ASV_861 | 8.10797 | Depleted | ASV_288 | 10.61649 | Depleted |
| ASV_70 | 13.84342 | Enriched | ASV_1249 | 8.07863 | Enriched | ASV_788 | 8.07142 | Depleted | ASV_1054 | 7.94009 | Depleted |
| ASV_137 | 12.55404 | Enriched | ASV_1032 | 9.29455 | Enriched | ASV_552 | 8.75201 | Depleted | ASV_167 | 12.09774 | Depleted |
| ASV_240 | 11.36267 | Enriched | ASV_705 | 9.02664 | Enriched | ASV_60 | 12.75308 | Depleted | ASV_1434 | 7.51513 | Depleted |
| ASV_363 | 10.78133 | Enriched | ASV_1614 | 7.26577 | Enriched | ASV_352 | 10.32539 | Depleted | ASV_940 | 7.9978 | Depleted |
| ASV_36 | 15.05311 | Enriched | ASV_933 | 8.63138 | Enriched | ASV_765 | 8.8643 | Depleted | ASV_627 | 8.80169 | Depleted |
| ASV_1324 | 8.22985 | Enriched | ASV_1656 | 7.59571 | Enriched | ASV_502 | 9.39861 | Depleted | ASV_521 | 9.55057 | Depleted |
| ASV_1467 | 8.26929 | Enriched | ASV_1512 | 7.49688 | Enriched | ASV_336 | 10.81066 | Depleted | ASV_916 | 7.97616 | Depleted |
| ASV_171 | 12.5131 | Enriched | ASV_325 | 10.38047 | Enriched | ASV_64 | 12.96203 | Depleted | ASV_384 | 9.94556 | Depleted |
| ASV_18 | 15.20766 | Enriched | ASV_2056 | 7.2178 | Enriched | ASV_1140 | 7.6471 | Depleted | ASV_1042 | 7.9035 | Depleted |
| ASV_224 | 11.94717 | Enriched | ASV_1591 | 9.32573 | Enriched | ASV_367 | 10.37899 | Depleted | ASV_1115 | 8.0868 | Depleted |
| ASV_1287 | 7.99647 | Enriched | ASV_1357 | 8.20289 | Enriched | ASV_749 | 8.2248 | Depleted | ASV_148 | 11.30772 | Depleted |
| ASV_150 | 12.21365 | Enriched | ASV_1010 | 8.64459 | Enriched | ASV_287 | 10.48627 | Depleted | ASV_856 | 8.18199 | Depleted |
| ASV_499 | 10.18812 | Enriched | ASV_4278 | 6.85903 | Enriched | ASV_438 | 10.61921 | Depleted | ASV_1063 | 7.67085 | Depleted |
| ASV_405 | 10.17404 | Enriched | ASV_1146 | 8.40051 | Enriched | ASV_726 | 8.57128 | Depleted | ASV_812 | 8.32538 | Depleted |
| ASV_117 | 12.93695 | Enriched | ASV_1708 | 7.53906 | Enriched | ASV_440 | 9.12727 | Depleted | ASV_800 | 8.53142 | Depleted |
| ASV_1888 | 7.28229 | Enriched | ASV_2144 | 7.07713 | Enriched | ASV_942 | 7.89887 | Depleted | ASV_63 | 13.04906 | Depleted |
| ASV_1380 | 8.38098 | Enriched | ASV_2020 | 7.37874 | Enriched | ASV_223 | 10.52093 | Depleted | ASV_329 | 10.30281 | Depleted |
| ASV_229 | 11.11439 | Enriched | ASV_1009 | 8.19085 | Enriched | ASV_631 | 8.97038 | Depleted | ASV_1345 | 7.2506 | Depleted |
| ASV_1433 | 7.73095 | Enriched | ASV_1806 | 7.39647 | Enriched | ASV_214 | 11.40476 | Depleted | ASV_599 | 9.24836 | Depleted |

| ASV_121 | 13.10264 | Enriched | ASV_28 | 13.84634 | Enriched | ASV_197 | 11.10437 | Depleted | ASV_814 | 8.12084 | Depleted |
|----------|----------|----------|----------|----------|----------|----------|----------|----------|----------|----------|----------|
| ASV_761 | 9.28349 | Enriched | ASV_2967 | 7.0915 | Enriched | ASV_443 | 9.45095 | Depleted | ASV_1029 | 7.9115 | Depleted |
| ASV_922 | 8.37668 | Enriched | ASV_1791 | 7.43835 | Enriched | ASV_192 | 10.75176 | Depleted | ASV_673 | 8.37615 | Depleted |
| ASV_1506 | 8.05922 | Enriched | ASV_259 | 10.94241 | Enriched | ASV_69 | 12.29824 | Depleted | ASV_1027 | 7.77894 | Depleted |
| ASV_677 | 9.45154 | Enriched | ASV_2774 | 6.71779 | Enriched | ASV_925 | 7.50168 | Depleted | ASV_935 | 7.92239 | Depleted |
| ASV_337 | 10.97035 | Enriched | ASV_179 | 13.61298 | Enriched | ASV_672 | 9.17364 | Depleted | ASV_1427 | 7.28252 | Depleted |
| ASV_799 | 9.01338 | Enriched | ASV_886 | 8.7673 | Enriched | ASV_754 | 8.29847 | Depleted | ASV_99 | 11.51951 | Depleted |
| ASV_477 | 10.05915 | Enriched | ASV_1124 | 8.5722 | Enriched | ASV_306 | 10.52432 | Depleted | ASV_772 | 8.80826 | Depleted |
| ASV_1605 | 7.60851 | Enriched | ASV_1170 | 8.22554 | Enriched | ASV_210 | 10.36778 | Depleted | ASV_1127 | 7.52184 | Depleted |
| ASV_1341 | 8.15287 | Enriched | ASV_703 | 9.3964 | Enriched | ASV_750 | 8.45302 | Depleted | ASV_1217 | 7.91768 | Depleted |
| ASV_376 | 10.24316 | Enriched | ASV_1739 | 10.1386 | Enriched | ASV_756 | 8.37702 | Depleted | ASV_1374 | 7.42831 | Depleted |
| ASV_204 | 11.89785 | Enriched | ASV_572 | 9.76255 | Enriched | ASV_658 | 8.72319 | Depleted | ASV_512 | 9.81602 | Depleted |
| ASV_955 | 9.02952 | Enriched | ASV_583 | 7.45696 | Enriched | ASV_365 | 10.15589 | Depleted | ASV_1060 | 7.76186 | Depleted |
| ASV_1962 | 7.30342 | Enriched | ASV_2969 | 6.76876 | Enriched | ASV_1355 | 7.51093 | Depleted | ASV_693 | 8.61299 | Depleted |
| ASV_1302 | 8.91889 | Enriched | ASV_2046 | 7.53915 | Enriched | ASV_713 | 8.13642 | Depleted | ASV_453 | 9.65067 | Depleted |
| ASV_189 | 11.96604 | Enriched | ASV_157 | 11.94915 | Depleted | ASV_737 | 7.85495 | Depleted | ASV_581 | 9.35887 | Depleted |
| ASV_93 | 12.77542 | Enriched | ASV_6 | 16.18883 | Depleted | ASV_193 | 11.65696 | Depleted | ASV_995 | 8.43104 | Depleted |
| ASV_387 | 10.82676 | Enriched | ASV_107 | 12.29854 | Depleted | ASV_1391 | 7.07378 | Depleted | ASV_1182 | 7.43642 | Depleted |
| ASV_850 | 8.89894 | Enriched | ASV_22 | 14.12823 | Depleted | ASV_149 | 11.83658 | Depleted | ASV_1252 | 7.7411 | Depleted |
| ASV_1457 | 7.78917 | Enriched | ASV_221 | 10.75624 | Depleted | ASV_533 | 9.17069 | Depleted | ASV_1228 | 7.72285 | Depleted |
| ASV_1090 | 7.91257 | Enriched | ASV_98 | 12.41815 | Depleted | ASV_481 | 8.92712 | Depleted | ASV_130 | 12.18017 | Depleted |
| ASV_1446 | 7.80209 | Enriched | ASV_116 | 12.17288 | Depleted | ASV_297 | 10.81258 | Depleted | ASV_4 | 16.76444 | Depleted |

| ASV_1774 | 7.27277 | Enriched | ASV_410 | 9.72433 | Depleted | ASV_565 | 8.60153 | Depleted | ASV_882 | 8.23296 | Depleted |
|----------|----------|----------|----------|----------|----------|----------|----------|----------|----------|----------|----------|
| ASV_219 | 11.78763 | Enriched | ASV_432 | 9.70195 | Depleted | ASV_217 | 10.65817 | Depleted | ASV_1162 | 8.04574 | Depleted |
| ASV_592 | 9.83439 | Enriched | ASV_341 | 9.8971 | Depleted | ASV_742 | 8.02348 | Depleted | ASV_636 | 8.57681 | Depleted |
| ASV_734 | 9.30324 | Enriched | ASV_434 | 10.00243 | Depleted | ASV_1259 | 7.50861 | Depleted | ASV_1462 | 7.1069 | Depleted |
| ASV_1296 | 7.89483 | Enriched | ASV_87 | 12.62018 | Depleted | ASV_1580 | 7.38008 | Depleted | ASV_905 | 7.983 | Depleted |
| ASV_747 | 9.35401 | Enriched | ASV_271 | 10.63706 | Depleted | ASV_250 | 10.78597 | Depleted | ASV_460 | 9.36927 | Depleted |
| ASV_324 | 11.04965 | Enriched | ASV_1022 | 9.38931 | Depleted | ASV_1095 | 7.67141 | Depleted | ASV_675 | 8.40292 | Depleted |
| ASV_49 | 14.0219 | Enriched | ASV_265 | 15.26419 | Depleted | ASV_511 | 9.24796 | Depleted | ASV_1103 | 7.86694 | Depleted |
| ASV_1366 | 7.6314 | Enriched | ASV_586 | 11.34399 | Depleted | ASV_248 | 10.82689 | Depleted | ASV_1153 | 7.99099 | Depleted |
| ASV_202 | 12.00932 | Enriched | ASV_522 | 9.45322 | Depleted | ASV_253 | 11.20338 | Depleted | ASV_1305 | 7.54442 | Depleted |
| ASV_422 | 10.03416 | Enriched | ASV_258 | 10.82056 | Depleted | ASV_735 | 8.32936 | Depleted | ASV_500 | 9.69932 | Depleted |
| ASV_1996 | 7.39775 | Enriched | ASV_147 | 13.94487 | Depleted | ASV_100 | 12.40005 | Depleted | ASV_21 | 15.04211 | Depleted |
| ASV_392 | 10.116 | Enriched | ASV_447 | 9.67732 | Depleted | ASV_411 | 10.2753 | Depleted | ASV_1093 | 8.1428 | NotSig |
| ASV_963 | 8.85196 | Enriched | ASV_20 | 14.17246 | Depleted | ASV_233 | 11.36921 | Depleted | ASV_305 | 8.62823 | NotSig |
| ASV_428 | 10.63665 | Enriched | ASV_23 | 14.51835 | Depleted | ASV_452 | 9.52771 | Depleted | ASV_591 | 9.43978 | NotSig |
| ASV_1294 | 8.00758 | Enriched | ASV_429 | 9.61736 | Depleted | ASV_997 | 7.96913 | Depleted | ASV_559 | 9.34428 | NotSig |
| ASV_787 | 9.17344 | Enriched | ASV_2 | 17.01591 | Depleted | ASV_266 | 10.75849 | Depleted | ASV_371 | 10.42992 | NotSig |
| ASV_299 | 11.20664 | Enriched | ASV_268 | 10.70847 | Depleted | ASV_957 | 7.47824 | Depleted | ASV_946 | 8.52128 | NotSig |
| ASV_83 | 13.25921 | Enriched | ASV_1181 | 7.37858 | Depleted | ASV_1186 | 7.27626 | Depleted | ASV_731 | 8.8513 | NotSig |
| ASV_951 | 8.50808 | Enriched | ASV_73 | 13.06342 | Depleted | ASV_1238 | 7.12332 | Depleted | ASV_1521 | 7.96726 | NotSig |
| ASV_1441 | 7.8537 | Enriched | ASV_1052 | 8.03395 | Depleted | ASV_151 | 12.02405 | Depleted | ASV_541 | 9.28014 | NotSig |
| ASV_2189 | 7.2164 | Enriched | ASV_338 | 9.64991 | Depleted | ASV_1024 | 7.86271 | Depleted | ASV_1201 | 7.56929 | NotSig |

| ASV_496 | 9.81768 | Enriched | ASV_85 | 12.67127 | Depleted | ASV_1236 | 7.45118 | Depleted | ASV_895 | 8.16899 | NotSig |
|----------|----------|----------|----------|----------|----------|----------|----------|----------|----------|----------|--------|
| ASV_2018 | 7.04847 | Enriched | ASV_456 | 9.70514 | Depleted | ASV_1048 | 8.46731 | Depleted | ASV_965 | 8.41456 | NotSig |
| ASV_1317 | 7.59799 | Enriched | ASV_725 | 8.15868 | Depleted | ASV_1370 | 7.14949 | Depleted | ASV_296 | 10.75604 | NotSig |
| ASV_1267 | 8.12369 | Enriched | ASV_775 | 8.74411 | Depleted | ASV_59 | 12.40503 | Depleted | ASV_1447 | 7.41949 | NotSig |
| ASV_1050 | 8.25887 | Enriched | ASV_865 | 7.70372 | Depleted | ASV_468 | 9.29335 | Depleted | ASV_538 | 9.87408 | NotSig |
| ASV_1670 | 7.81662 | Enriched | ASV_466 | 9.49389 | Depleted | ASV_732 | 8.64161 | Depleted | ASV_1311 | 8.06813 | NotSig |
| ASV_1338 | 8.24955 | Enriched | ASV_529 | 9.44002 | Depleted | ASV_588 | 9.2224 | Depleted | ASV_1531 | 7.30823 | NotSig |
| ASV_1501 | 8.21046 | Enriched | ASV_216 | 10.98475 | Depleted | ASV_907 | 8.05547 | Depleted | ASV_1527 | 7.55005 | NotSig |
| ASV_444 | 10.10465 | Enriched | ASV_628 | 8.75981 | Depleted | ASV_662 | 9.19416 | Depleted | ASV_630 | 9.71509 | NotSig |
| ASV_2199 | 7.14436 | Enriched | ASV_19 | 14.22622 | Depleted | ASV_506 | 9.81204 | Depleted | ASV_885 | 8.49645 | NotSig |
| ASV_790 | 9.25156 | Enriched | ASV_1213 | 7.50203 | Depleted | ASV_274 | 10.91048 | Depleted | ASV_571 | 8.82669 | NotSig |
| ASV_1248 | 8.03812 | Enriched | ASV_1137 | 7.89911 | Depleted | ASV_825 | 8.65034 | Depleted | ASV_323 | 10.7291 | NotSig |
| ASV_710 | 9.47399 | Enriched | ASV_475 | 10.12785 | Depleted | ASV_417 | 10.13119 | Depleted | ASV_574 | 9.55941 | NotSig |
| ASV_758 | 9.12953 | Enriched | ASV_446 | 9.92227 | Depleted | ASV_519 | 9.23832 | Depleted | ASV_733 | 8.3549 | NotSig |
| ASV_2130 | 7.21328 | Enriched | ASV_212 | 11.57344 | Depleted | ASV_1025 | 7.75219 | Depleted | ASV_1330 | 7.18894 | NotSig |
| ASV_1882 | 7.37305 | Enriched | ASV_364 | 10.12793 | Depleted | ASV_205 | 11.11638 | Depleted | ASV_1766 | 7.32748 | NotSig |
| ASV_868 | 8.66584 | Enriched | ASV_7 | 16.24191 | Depleted | ASV_926 | 8.17798 | Depleted | ASV_1126 | 7.89449 | NotSig |
| ASV_720 | 9.42483 | Enriched | ASV_312 | 10.15841 | Depleted | ASV_822 | 8.77058 | Depleted | ASV_1573 | 7.38997 | NotSig |
| ASV_1482 | 7.48312 | Enriched | ASV_515 | 15.25519 | Depleted | ASV_154 | 11.92886 | Depleted | ASV_1843 | 7.24919 | NotSig |
| ASV_977 | 8.62317 | Enriched | ASV_43 | 13.36514 | Depleted | ASV_971 | 7.99359 | Depleted | ASV_1189 | 8.25312 | NotSig |
| ASV_818 | 8.3949 | Enriched | ASV_391 | 10.07661 | Depleted | ASV_401 | 10.25279 | Depleted | ASV_349 | 8.76482 | NotSig |
| ASV_1114 | 8.00708 | Enriched | ASV_201 | 10.75577 | Depleted | ASV_920 | 8.03871 | Depleted | ASV_56 | 10.98571 | NotSig |

| ASV_1454 | 7.9037 | Enriched | ASV_585 | 8.92592 | Depleted | ASV_441 | 9.60772 | Depleted | ASV_158 | 11.60468 | NotSig |
|----------|----------|----------|---------|----------|----------|----------|----------|----------|----------|----------|--------|
| ASV_2831 | 6.80028 | Enriched | ASV_412 | 12.24663 | Depleted | ASV_689 | 9.11577 | Depleted | ASV_602 | 9.2051 | NotSig |
| ASV_261 | 10.45364 | Enriched | ASV_697 | 8.54856 | Depleted | ASV_1588 | 7.34621 | Depleted | ASV_1172 | 7.94028 | NotSig |
| ASV_2420 | 8.03003 | Enriched | ASV_448 | 9.53378 | Depleted | ASV_1764 | 7.15589 | Depleted | ASV_235 | 8.73525 | NotSig |
| ASV_199 | 11.16654 | Enriched | ASV_373 | 9.98695 | Depleted | ASV_1119 | 7.36238 | Depleted | ASV_276 | 8.55702 | NotSig |
| ASV_1171 | 8.19338 | Enriched | ASV_615 | 8.5363 | Depleted | ASV_779 | 8.7877 | Depleted | ASV_1312 | 7.96585 | NotSig |
| ASV_1129 | 8.35357 | Enriched | ASV_608 | 9.20426 | Depleted | ASV_131 | 11.99109 | Depleted | ASV_1877 | 7.27459 | NotSig |
| ASV_927 | 8.66013 | Enriched | ASV_664 | 8.4753 | Depleted | ASV_80 | 12.96885 | Depleted | ASV_836 | 8.00303 | NotSig |
| ASV_290 | 10.93584 | Enriched | ASV_437 | 9.54993 | Depleted | ASV_485 | 9.87469 | Depleted | ASV_721 | 8.21536 | NotSig |
| ASV_504 | 8.12233 | Enriched | ASV_763 | 8.50829 | Depleted | ASV_326 | 10.24109 | Depleted | ASV_1183 | 7.91912 | NotSig |
| ASV_1207 | 8.00318 | Enriched | ASV_284 | 10.4765 | Depleted | ASV_540 | 9.59456 | Depleted | | | |
| ASV_1478 | 7.84489 | Enriched | ASV_234 | 11.16498 | Depleted | ASV_381 | 10.4144 | Depleted | | | |
| ASV_778 | 8.92695 | Enriched | ASV_528 | 8.75337 | Depleted | ASV_209 | 11.20881 | Depleted | | | |

2 Identified harmful and toxic ASVs in plastisphere and ambient seawater.

| ID | Species | Type | ID | Species | Type | ID | Species | Type |
|----------|-----------------------------|---------|----------|-------------------|---------|----------|-------------------------------|---------|
| ASV_19 | Akashiwo sanguinea | Toxic | ASV_3021 | Dunaliella salina | Harmful | ASV_214 | Leptocylindrus danicus | Harmful |
| ASV_3137 | Akashiwo sanguinea | Toxic | ASV_3079 | Dunaliella salina | Harmful | ASV_2803 | Leptocylindrus danicus | Harmful |
| ASV_1222 | Alexandrium andersonii | Toxic | ASV_3131 | Dunaliella salina | Harmful | ASV_126 | Levanderina fissa | Harmful |
| ASV_3341 | Alexandrium andersonii | Toxic | ASV_3115 | Dunaliella salina | Harmful | ASV_1349 | Lingulodinium polyedra | Toxic |
| ASV_2215 | Alexandrium cohorticula | Harmful | ASV_3164 | Dunaliella salina | Harmful | ASV_2274 | Lingulodinium polyedra | Toxic |
| ASV_754 | Alexandrium hiranoi | Toxic | ASV_3187 | Dunaliella salina | Harmful | ASV_3681 | Margalefidinium fulvescens | Toxic |
| ASV_233 | Alexandrium ostenfeldii | Toxic | ASV_3301 | Dunaliella salina | Harmful | ASV_800 | Margalefidinium polykrikoides | Toxic |
| ASV_686 | Alexandrium ostenfeldii | Toxic | ASV_3348 | Dunaliella salina | Harmful | ASV_1189 | Nitzschia longissima | Harmful |
| ASV_167 | Alexandrium tamarense | Toxic | ASV_3358 | Dunaliella salina | Harmful | ASV_1739 | Nitzschia longissima | Harmful |
| ASV_485 | Alexandrium tamarense | Toxic | ASV_3363 | Dunaliella salina | Harmful | ASV_69 | Noctiluca scintillans | Harmful |
| ASV_2400 | Amphidinium klebsii | Toxic | ASV_3405 | Dunaliella salina | Harmful | ASV_100 | Noctiluca scintillans | Harmful |
| ASV_814 | Asterionellopsis glacialis | Harmful | ASV_3584 | Dunaliella salina | Harmful | ASV_223 | Noctiluca scintillans | Harmful |
| ASV_1589 | Aureococcus anophagefferens | Harmful | ASV_3631 | Dunaliella salina | Harmful | ASV_3551 | Noctiluca scintillans | Harmful |
| ASV_1809 | Aureococcus anophagefferens | Harmful | ASV_3840 | Dunaliella salina | Harmful | ASV_107 | Plagioselmis prolonga | Harmful |
| ASV_3590 | Aureococcus anophagefferens | Harmful | ASV_3884 | Dunaliella salina | Harmful | ASV_441 | Plagioselmis prolonga | Harmful |
| ASV_935 | Azadinium poporum | Toxic | ASV_3891 | Dunaliella salina | Harmful | ASV_1115 | Plagioselmis prolonga | Harmful |
| ASV_336 | Chaetoceros curvisetus | Harmful | ASV_3925 | Dunaliella salina | Harmful | ASV_1335 | Plagioselmis prolonga | Harmful |
| ASV_1788 | Chaetoceros debilis | Harmful | ASV_3943 | Dunaliella salina | Harmful | ASV_304 | Polykrikos hartmannii | Toxic |
| ASV_212 | Chaetoceros diadema | Harmful | ASV_3956 | Dunaliella salina | Harmful | ASV_411 | Prorocentrum cordatum | Toxic |
| ASV_3005 | Chaetoceros diadema | Harmful | ASV_4094 | Dunaliella salina | Harmful | ASV_452 | Prorocentrum cordatum | Toxic |

| ASV_2221 | Chaetoceros lorenzianus | Harmful | ASV_4181 | Dunaliella salina | Harmful | ASV_512 | Prorocentrum cordatum | Toxic |
|----------|------------------------------|---------|----------|---------------------|---------|----------|-----------------------|-------|
| ASV_4750 | Chaetoceros pseudocurvisetus | Harmful | ASV_4193 | Dunaliella salina | Harmful | ASV_572 | Prorocentrum cordatum | Toxic |
| ASV_3399 | Chaetoceros rostratus | Harmful | ASV_4194 | Dunaliella salina | Harmful | ASV_733 | Prorocentrum cordatum | Toxic |
| ASV_2824 | Chaetoceros simplex | Harmful | ASV_4205 | Dunaliella salina | Harmful | ASV_731 | Prorocentrum cordatum | Toxic |
| ASV_2330 | Chaetoceros socialis | Harmful | ASV_4211 | Dunaliella salina | Harmful | ASV_749 | Prorocentrum cordatum | Toxic |
| ASV_2883 | Chaetoceros socialis | Harmful | ASV_4277 | Dunaliella salina | Harmful | ASV_1038 | Prorocentrum cordatum | Toxic |
| ASV_2 | Chaetoceros tenuissimus | Harmful | ASV_4311 | Dunaliella salina | Harmful | ASV_1315 | Prorocentrum cordatum | Toxic |
| ASV_2533 | Chaetoceros throndsenii | Harmful | ASV_4338 | Dunaliella salina | Harmful | ASV_1464 | Prorocentrum cordatum | Toxic |
| ASV_3299 | Chrysochromulina leadbeateri | Toxic | ASV_4383 | Dunaliella salina | Harmful | ASV_1812 | Prorocentrum cordatum | Toxic |
| ASV_4113 | Chrysochromulina leadbeateri | Toxic | ASV_4406 | Dunaliella salina | Harmful | ASV_1877 | Prorocentrum cordatum | Toxic |
| ASV_4878 | Chrysochromulina leadbeateri | Toxic | ASV_4437 | Dunaliella salina | Harmful | ASV_2213 | Prorocentrum cordatum | Toxic |
| ASV_296 | Coscinodiscus radiatus | Harmful | ASV_4459 | Dunaliella salina | Harmful | ASV_2506 | Prorocentrum cordatum | Toxic |
| ASV_1236 | Coscinodiscus radiatus | Harmful | ASV_4704 | Dunaliella salina | Harmful | ASV_2784 | Prorocentrum cordatum | Toxic |
| ASV_2862 | Coscinodiscus radiatus | Harmful | ASV_4743 | Dunaliella salina | Harmful | ASV_2869 | Prorocentrum cordatum | Toxic |
| ASV_6 | Cyclotella meneghiniana | Harmful | ASV_4772 | Dunaliella salina | Harmful | ASV_3347 | Prorocentrum cordatum | Toxic |
| ASV_99 | Cyclotella meneghiniana | Harmful | ASV_4947 | Dunaliella salina | Harmful | ASV_3481 | Prorocentrum cordatum | Toxic |
| ASV_148 | Cyclotella meneghiniana | Harmful | ASV_4993 | Dunaliella salina | Harmful | ASV_3488 | Prorocentrum cordatum | Toxic |
| ASV_216 | Cyclotella meneghiniana | Harmful | ASV_5114 | Dunaliella salina | Harmful | ASV_3605 | Prorocentrum cordatum | Toxic |
| ASV_438 | Cyclotella meneghiniana | Harmful | ASV_160 | Eucampia zodiacus | Harmful | ASV_3689 | Prorocentrum cordatum | Toxic |
| ASV_810 | Cyclotella meneghiniana | Harmful | ASV_453 | Eucampia zodiacus | Harmful | ASV_3816 | Prorocentrum cordatum | Toxic |
| ASV_4422 | Cyclotella meneghiniana | Harmful | ASV_1162 | Eucampia zodiacus | Harmful | ASV_4252 | Prorocentrum cordatum | Toxic |
| ASV_118 | Cylindrotheca closterium | Harmful | ASV_365 | Fibrocapsa japonica | Toxic | ASV_4310 | Prorocentrum cordatum | Toxic |
| ASV_179 | Cylindrotheca closterium | Harmful | ASV_1576 | Fibrocapsa japonica | Toxic | ASV_4350 | Prorocentrum cordatum | Toxic |
| ASV_189 | Cylindrotheca closterium | Harmful | ASV_2055 | Fibrocapsa japonica | Toxic | ASV_4451 | Prorocentrum cordatum | Toxic |
| | | | | | | | | |

| ASV_246 | Cylindrotheca closterium | Harmful | ASV_2536 | Fibrocapsa japonica | Toxic | ASV_4910 | Prorocentrum cordatum | Toxic |
|----------|--------------------------|---------|----------|-------------------------|---------|----------|---------------------------|---------|
| ASV_423 | Cylindrotheca closterium | Harmful | ASV_3897 | Fibrocapsa japonica | Toxic | ASV_636 | Prorocentrum rhathymum | Toxic |
| ASV_559 | Cylindrotheca closterium | Harmful | ASV_4922 | Gonyaulax polygramma | Harmful | ASV_1859 | Prorocentrum rhathymum | Toxic |
| ASV_631 | Cylindrotheca closterium | Harmful | ASV_2319 | Gonyaulax spinifera | Toxic | ASV_2415 | Prorocentrum rhathymum | Toxic |
| ASV_799 | Cylindrotheca closterium | Harmful | ASV_3954 | Gonyaulax spinifera | Toxic | ASV_265 | Prorocentrum triestinum | Harmful |
| ASV_778 | Cylindrotheca closterium | Harmful | ASV_3287 | Gonyaulax verior | Harmful | ASV_1837 | Prorocentrum triestinum | Harmful |
| ASV_873 | Cylindrotheca closterium | Harmful | ASV_586 | Guinardia flaccida | Harmful | ASV_4514 | Protoceratium reticulatum | Toxic |
| ASV_1341 | Cylindrotheca closterium | Harmful | ASV_1182 | Gymnodinium catenatum | Toxic | ASV_20 | Protodinium simplex | Harmful |
| ASV_1467 | Cylindrotheca closterium | Harmful | ASV_1511 | Gymnodinium catenatum | Toxic | ASV_284 | Protodinium simplex | Harmful |
| ASV_1882 | Cylindrotheca closterium | Harmful | ASV_3218 | Gymnodinium catenatum | Toxic | ASV_448 | Protodinium simplex | Harmful |
| ASV_1946 | Cylindrotheca closterium | Harmful | ASV_171 | Halamphora coffeiformis | Harmful | ASV_2431 | Protodinium simplex | Harmful |
| ASV_2966 | Cylindrotheca closterium | Harmful | ASV_208 | Halamphora coffeiformis | Harmful | ASV_3445 | Protoperidinium crassipes | Toxic |
| ASV_2985 | Cylindrotheca closterium | Harmful | ASV_239 | Halamphora coffeiformis | Harmful | ASV_693 | Prymnesium parvum | Toxic |
| ASV_4397 | Cylindrotheca closterium | Harmful | ASV_370 | Halamphora coffeiformis | Harmful | ASV_1399 | Prymnesium parvum | Toxic |
| ASV_4496 | Cylindrotheca closterium | Harmful | ASV_538 | Halamphora coffeiformis | Harmful | ASV_1425 | Prymnesium parvum | Toxic |
| ASV_4847 | Cylindrotheca closterium | Harmful | ASV_707 | Halamphora coffeiformis | Harmful | ASV_1944 | Prymnesium parvum | Toxic |
| ASV_234 | Dictyocha fibula | Harmful | ASV_720 | Halamphora coffeiformis | Harmful | ASV_2234 | Prymnesium parvum | Toxic |
| ASV_3130 | Dinophysis acuminata | Toxic | ASV_747 | Halamphora coffeiformis | Harmful | ASV_3232 | Prymnesium parvum | Toxic |
| ASV_4797 | Dinophysis acuminata | Toxic | ASV_955 | Halamphora coffeiformis | Harmful | ASV_3670 | Prymnesium parvum | Toxic |
| ASV_1580 | Dinophysis miles | Toxic | ASV_965 | Halamphora coffeiformis | Harmful | ASV_4166 | Prymnesium parvum | Toxic |
| ASV_4135 | Dinophysis miles | Toxic | ASV_1114 | Halamphora coffeiformis | Harmful | ASV_4424 | Prymnesium parvum | Toxic |
| ASV_85 | Dunaliella salina | Harmful | ASV_1124 | Halamphora coffeiformis | Harmful | ASV_4967 | Prymnesium parvum | Toxic |
| ASV_193 | Dunaliella salina | Harmful | ASV_1172 | Halamphora coffeiformis | Harmful | ASV_4972 | Prymnesium parvum | Toxic |
| ASV_204 | Dunaliella salina | Harmful | ASV_1207 | Halamphora coffeiformis | Harmful | ASV_5053 | Prymnesium parvum | Toxic |
| | | | | | | | | |

| ASV_229 | Dunaliella salina | Harmful | ASV_1248 | Halamphora coffeiformis | Harmful | ASV_653 | Prymnesium polylepis | Toxic |
|----------|-------------------|---------|----------|-------------------------|---------|----------|-----------------------------------|---------|
| ASV_253 | Dunaliella salina | Harmful | ASV_1287 | Halamphora coffeiformis | Harmful | ASV_1137 | Prymnesium polylepis | Toxic |
| ASV_274 | Dunaliella salina | Harmful | ASV_1312 | Halamphora coffeiformis | Harmful | ASV_2682 | Prymnesium polylepis | Toxic |
| ASV_364 | Dunaliella salina | Harmful | ASV_1324 | Halamphora coffeiformis | Harmful | ASV_4585 | Prymnesium polylepis | Toxic |
| ASV_417 | Dunaliella salina | Harmful | ASV_1358 | Halamphora coffeiformis | Harmful | ASV_1119 | Pseudo-nitzschia australis | Toxic |
| ASV_437 | Dunaliella salina | Harmful | ASV_1501 | Halamphora coffeiformis | Harmful | ASV_2571 | Pseudo-nitzschia australis | Toxic |
| ASV_506 | Dunaliella salina | Harmful | ASV_1854 | Halamphora coffeiformis | Harmful | ASV_571 | Pseudo-nitzschia cuspidata | Toxic |
| ASV_585 | Dunaliella salina | Harmful | ASV_1962 | Halamphora coffeiformis | Harmful | ASV_23 | Pseudo-nitzschia delicatissima | Toxic |
| ASV_599 | Dunaliella salina | Harmful | ASV_2056 | Halamphora coffeiformis | Harmful | ASV_4271 | Pseudo-nitzschia delicatissima | Toxic |
| ASV_634 | Dunaliella salina | Harmful | ASV_2141 | Halamphora coffeiformis | Harmful | ASV_4448 | Pseudo-nitzschia delicatissima | Toxic |
| ASV_663 | Dunaliella salina | Harmful | ASV_2268 | Halamphora coffeiformis | Harmful | ASV_4662 | Pseudo-nitzschia delicatissima | Toxic |
| ASV_662 | Dunaliella salina | Harmful | ASV_2295 | Halamphora coffeiformis | Harmful | ASV_4901 | Pseudo-nitzschia delicatissima | Toxic |
| ASV_689 | Dunaliella salina | Harmful | ASV_2526 | Halamphora coffeiformis | Harmful | ASV_266 | Pseudochattonella verruculosa | Toxic |
| ASV_822 | Dunaliella salina | Harmful | ASV_2542 | Halamphora coffeiformis | Harmful | ASV_2484 | Pseudocochlodinium profundisulcus | Harmful |
| ASV_827 | Dunaliella salina | Harmful | ASV_2580 | Halamphora coffeiformis | Harmful | ASV_3202 | Pseudocochlodinium profundisulcus | Harmful |
| ASV_825 | Dunaliella salina | Harmful | ASV_2799 | Halamphora coffeiformis | Harmful | ASV_5111 | Pseudocochlodinium profundisulcus | Harmful |
| ASV_972 | Dunaliella salina | Harmful | ASV_3046 | Halamphora coffeiformis | Harmful | ASV_1195 | Pyrophacus steinii | Harmful |
| ASV_998 | Dunaliella salina | Harmful | ASV_3074 | Halamphora coffeiformis | Harmful | ASV_64 | Skeletonema marinoi | Harmful |
| ASV_1024 | Dunaliella salina | Harmful | ASV_3281 | Halamphora coffeiformis | Harmful | ASV_259 | Skeletonema marinoi | Harmful |
| ASV_1051 | Dunaliella salina | Harmful | ASV_3513 | Halamphora coffeiformis | Harmful | ASV_73 | Teleaulax acuta | Harmful |
| ASV_1291 | Dunaliella salina | Harmful | ASV_3613 | Halamphora coffeiformis | Harmful | ASV_197 | Teleaulax acuta | Harmful |
| | | | | | | | | |

| ASV_1323 | Dunaliella salina | Harmful | ASV_3712 | Halamphora coffeiformis | Harmful | ASV_287 | Teleaulax acuta | Harmful |
|----------|-------------------|---------|----------|-------------------------|---------|----------|----------------------------|---------|
| ASV_1347 | Dunaliella salina | Harmful | ASV_3779 | Halamphora coffeiformis | Harmful | ASV_697 | Teleaulax acuta | Harmful |
| ASV_1352 | Dunaliella salina | Harmful | ASV_3869 | Halamphora coffeiformis | Harmful | ASV_985 | Teleaulax acuta | Harmful |
| ASV_1474 | Dunaliella salina | Harmful | ASV_3872 | Halamphora coffeiformis | Harmful | ASV_1656 | Teleaulax acuta | Harmful |
| ASV_1449 | Dunaliella salina | Harmful | ASV_3940 | Halamphora coffeiformis | Harmful | ASV_1754 | Teleaulax acuta | Harmful |
| ASV_1539 | Dunaliella salina | Harmful | ASV_3969 | Halamphora coffeiformis | Harmful | ASV_1751 | Teleaulax acuta | Harmful |
| ASV_1614 | Dunaliella salina | Harmful | ASV_4210 | Halamphora coffeiformis | Harmful | ASV_1876 | Teleaulax acuta | Harmful |
| ASV_1632 | Dunaliella salina | Harmful | ASV_4964 | Halamphora coffeiformis | Harmful | ASV_1926 | Teleaulax acuta | Harmful |
| ASV_1645 | Dunaliella salina | Harmful | ASV_147 | Heterocapsa triquetra | Harmful | ASV_2074 | Teleaulax acuta | Harmful |
| ASV_1724 | Dunaliella salina | Harmful | ASV_456 | Heterosigma akashiwo | Toxic | ASV_2086 | Teleaulax acuta | Harmful |
| ASV_1735 | Dunaliella salina | Harmful | ASV_1153 | Heterosigma akashiwo | Toxic | ASV_2257 | Teleaulax acuta | Harmful |
| ASV_1906 | Dunaliella salina | Harmful | ASV_1441 | Heterosigma akashiwo | Toxic | ASV_2299 | Teleaulax acuta | Harmful |
| ASV_1890 | Dunaliella salina | Harmful | ASV_1531 | Heterosigma akashiwo | Toxic | ASV_2324 | Teleaulax acuta | Harmful |
| ASV_1952 | Dunaliella salina | Harmful | ASV_1643 | Heterosigma akashiwo | Toxic | ASV_2472 | Teleaulax acuta | Harmful |
| ASV_1983 | Dunaliella salina | Harmful | ASV_1774 | Heterosigma akashiwo | Toxic | ASV_3337 | Teleaulax acuta | Harmful |
| ASV_2018 | Dunaliella salina | Harmful | ASV_1956 | Heterosigma akashiwo | Toxic | ASV_3534 | Teleaulax acuta | Harmful |
| ASV_2081 | Dunaliella salina | Harmful | ASV_2126 | Heterosigma akashiwo | Toxic | ASV_4279 | Teleaulax acuta | Harmful |
| ASV_2197 | Dunaliella salina | Harmful | ASV_2632 | Heterosigma akashiwo | Toxic | ASV_4441 | Teleaulax acuta | Harmful |
| ASV_2235 | Dunaliella salina | Harmful | ASV_3861 | Heterosigma akashiwo | Toxic | ASV_4515 | Teleaulax acuta | Harmful |
| ASV_2244 | Dunaliella salina | Harmful | ASV_4028 | Heterosigma akashiwo | Toxic | ASV_4793 | Teleaulax acuta | Harmful |
| ASV_2284 | Dunaliella salina | Harmful | ASV_4072 | Heterosigma akashiwo | Toxic | ASV_5052 | Teleaulax acuta | Harmful |
| ASV_2283 | Dunaliella salina | Harmful | ASV_4558 | Heterosigma akashiwo | Toxic | ASV_1101 | Tenuicylindrus belgicus | Harmful |
| ASV_2361 | Dunaliella salina | Harmful | ASV_4647 | Heterosigma akashiwo | Toxic | ASV_349 | Thalassiosira allenii | Harmful |
| ASV_2406 | Dunaliella salina | Harmful | ASV_4783 | Heterosigma akashiwo | Toxic | ASV_1183 | Thalassiosira curviseriata | Harmful |
| | | | | | | | | |

| ASV_2481 | Dunaliella salina | Harmful | ASV_5082 | Heterosigma akashiwo | Toxic | ASV_3440 | Thalassiosira curviseriata | Harmful |
|----------|-------------------|---------|----------|---------------------------|---------|----------|----------------------------|---------|
| ASV_2587 | Dunaliella salina | Harmful | ASV_932 | Karenia mikimotoi | Toxic | ASV_3827 | Thalassiosira curviseriata | Harmful |
| ASV_2651 | Dunaliella salina | Harmful | ASV_1117 | Karenia mikimotoi | Toxic | ASV_21 | Thalassiosira mala | Harmful |
| ASV_2620 | Dunaliella salina | Harmful | ASV_861 | Karlodinium veneficum | Toxic | ASV_158 | Thalassiosira mala | Harmful |
| ASV_2689 | Dunaliella salina | Harmful | ASV_1025 | Karlodinium veneficum | Toxic | ASV_765 | Thalassiosira pseudonana | Harmful |
| ASV_2690 | Dunaliella salina | Harmful | ASV_2572 | Karlodinium veneficum | Toxic | ASV_279 | Thalassiosira weissflogii | Harmful |
| ASV_2714 | Dunaliella salina | Harmful | ASV_4856 | Karlodinium veneficum | Toxic | ASV_1550 | Thalassiosira weissflogii | Harmful |
| ASV_2765 | Dunaliella salina | Harmful | ASV_1510 | Lauderia annulata | Harmful | ASV_2823 | Thalassiosira weissflogii | Harmful |
| ASV_2774 | Dunaliella salina | Harmful | ASV_205 | Lepidodinium chlorophorum | Harmful | ASV_4168 | Thalassiosira weissflogii | Harmful |
| ASV_2860 | Dunaliella salina | Harmful | ASV_312 | Lepidodinium chlorophorum | Harmful | ASV_466 | Tripos fusus | Harmful |
| ASV_2933 | Dunaliella salina | Harmful | ASV_502 | Lepidodinium chlorophorum | Harmful | | | |
| ASV_3016 | Dunaliella salina | Harmful | ASV_4183 | Lepidodinium chlorophorum | Harmful | | | |

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Publications from the Current PhD Study

- Li Zhang[#], Anqi Xiong[#], Changchao Li, Xintong Liu, Xiaohua Zhang, Shihao Gong, Meng Yan, Xian Qin, Yang Liu, Zhangxi Hu, James Kar-Hei Fang, Huanfeng Duan, Hongbin Liu, Leo L. Chan, Ling N. Jin (2024). Ecological pattern of microalgal communities and associated risks in coastal ecosystems. *ISME Communications*. (Major revision)
- 2. **Li Zhang**[#], Ziqian Xu[#], Changchao Li, Xi Pan, Yijing Wang, Xiaohua Zhang, Xintong Liu, Anqi Xiong, Yan Wang, Xinru Zhang, Yongkang Zhao, Shuping Ji, Jian Liu, Meng Yan, James K.H. Fang, Ling N. Jin (2025). Microalgae and associated ecological risks in coastal plastispheres. *Water Research X* (under review)
- 3. Xintong Liu; Bo Liang; Siyu Yao; Anqi Xiong; Xiaohua Zhang; Yajing Sun; Li Zhang; Changchao Li; Yuefei Ruan; Meng Yan; Yuen Wa Ho; James Kar-Hei; Wang Fang, Bei; Frederic Leusch; Daniel Schlenk; Wenhua Liu; Kenneth Mei Yee Leung; Ling Jin, (2024). Interplay of Natural Toxins and Anthropogenic Contaminants in Species-Specific Seawater Cytotoxicity: Insights from Cetacean Skin Fibroblasts. *Environmental Science & Technology* (under review)